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19 December 2024

Taylorville Resource Park Ltd
c/o Barker & Associates
KerstinG@barker.co.nz

Enquiries to: Michael Durand

Dear Ms Ghisel

RC-2024-0122 and RC-2021-0097-V10 – DISCHARGES OF CONTAMINANTS AT TAYLORVILLE RESOURCE PARK

Further information request under section 92(1) of the Resource Management Act 1991

Thank you for your application to discharge collected leachate and stormwater to land (RC-2024-0122) from the Taylorville Resource Park Ltd landfill and the application (RC2021.0097-V10) to change consent condition 11 of consent RC2021.0097 to authorise the discharges applied in application RC2024.0122. Also thank you for the response provided on s92(1) request related to application RC2024.0122 in a letter from Engeo addressed to Bruce Tinnelly of TRP dated 1 November 2024 and the accompanying AEE dated 19 October 2024.

We have engaged Carl Steffens of PDP to review AEE provided. Please see attached the review provided by PDP. As a result of these assessments further information is requested under section 92(1) of the RMA as listed below.

This information is requested so that WCRC can complete its assessment of the environmental effects of the proposed discharge and to determine notification or non-notification of the application. The additional information requested below will also enable us to assess potential adverse effects on the Grey District Council's water supply and potential ecological effects on surface water.

Information requested

1. Hydraulic conductivity estimates from onsite testing (pump testing and/or slug testing). If slug testing is used, then please provide testing data from at least 3 bore locations on the upper terrace surface.
2. Updated mounding and contaminant modelling simulations using the hydraulic conductivity estimates derived from the requested testing. Based on the testing results, please take into account that the results from testing (particularly slug testing) may provide local values for hydraulic conductivity but not necessarily be representative of the average hydraulic conductivity over the area of the landfill. Therefore, please provide multiple modelling simulations informed by the testing results using a realistic range of potential hydraulic conductivity values for the strata beneath the upper terrace surface (i.e. a sensitivity assessment based on varying the hydraulic conductivity that is informed by the testing results).
3. Updated contaminant modelling simulations using input parameters more consistent with high parameter concentrations that have been measured in leachate and wastewater.

4. An ecological survey of the surface water features and aquatic habitat surrounding the site, including Noname Creek, the drain along Taylorville Road, including that part of the drain that begins at spring seep TCD1 and the streams on the southern side of Taylorville Road (excluding the Grey River) and an impact assessment on surface water quality and ecological values resulting from the proposed discharge.
5. A water quality comparison between water samples collected from the GDC water supply, the Grey River, groundwater upgradient of the landfill north of Taylorville Road and surface water samples collected at TRD2 and TRD4. The samples should include analysis for anions and cations and nitrogen species, in addition to other parameters sampled to date by TRPL. The purpose of the sampling is to inform whether the GDC water supply source is dominantly Grey River water and that there is minimal input to the supply bores from upgradient groundwater and surface drainage.
6. Yield versus drawdown data from the GDC supply bores to confirm the gravel deposits immediately adjacent to the Grey River are of high permeability with a high hydraulic connection to the Grey River.
7. XRD sample analysis from the Kaiata mudstone underlying the landfill to confirm clay mineralogy in the mudstone and confirm ENGEO comments that swelling within any fractures that may be present in the mudstone will block the transmission of groundwater via a fracture network.

Next steps

The steps set out in section 92A(1) apply to this request. TRPL must take one of the options in section 92(1) by 3 February 2025, which is 15 working days from the date of this request.

Decisions under sections 95A and 95B of the RMA are postponed until receipt of this information, unless section 95C applies.

If you have any queries regarding this matter, please contact Michael Durand on 021 035 2593 or e-mail on michael@planzconsultants.co.nz.

Yours faithfully,



Michael Durand
Consultant Planning Officer
West Coast Regional Council



memorandum



TO	Michael Durand	FROM	Carl Steffens
	Planz Consultants	DATE	18 December 2024
RE	Technical Review – TRPL short-term consent application to discharge contaminants to ground.		

Pattle Delamore Partners Limited (PDP) has been engaged by West Coast Regional Council (WCRC) to provide a technical review of a resource consent application by Taylorville Resource Park Limited. The application has been prepared by Barker and Associates Limited (B&A) with technical assessment provided by ENGEO Limited (ENGEO). The application seeks a temporary discharge (for up to six months) from the existing onsite retention pond to land via a soakage area. The maximum daily discharge is up to 800 m³, which represents a continuous discharge of approximately 9.3 L/s). A change to condition 11 of existing resource consent RC-2021-0097-03 is also sought to allow for potential contamination of groundwater or surface water outside of the site.

PDP has recently prepared a report which commented on the potential impacts of TRPL landfill on the nearby Grey District Council (GDC) water supply intake (supply to Greymouth). That report is titled Impacts of Taylorville Resource Park Limited Landfill on Grey District Council Coal Creek Water Supply Intake and is dated 11 October 2024. Based on limited available information, the report concluded that the risk of contamination to the water supply was likely to be low. However, the report also highlighted that there are considerable uncertainties regarding the risk posed by leachate migration and given the proximity of the GDC supply to the landfill, a very high level of assessment should be provided by the consent holder. To date, that information has not been provided.

While there is still considerable additional information required to assess the ongoing effects of the landfill on the environment, we note that an ENGEO memorandum dated 6 November 2024 reports that TRPL has recently installed eight additional monitoring bores (MB13-MB19s/d). ENGEO note that they have updated their groundwater flow contours based on groundwater elevation measurements in all existing and new bores at the site. They report that this updated information shows that groundwater flow is to the west-southwest and their memo presents a preliminary figure showing the updated contours that have been constructed by ENGEO. This is useful additional information, however we have not been able to review the data used to construct the contours and we also consider the presence of groundwater seeps along the Taylorville Road cutting indicate there is also a southerly component to groundwater flow that is not evident from the ENGEO contours. It is possible there could be groundwater divides at the site and more complex groundwater flow than indicated by the ENGEO contours.

Further to the existing risks to the environment (and human health based on the presence of the landfill in proximity to the water supply), the current proposal represents an additional discharge at the site that was not considered during the drafting of the previous October 2024 PDP report.

The ENGEO technical assessment relating to the current proposal is provided in their report titled Preliminary Hydrogeological Assessment – 287 Taylorville Road, Coal Creek, Greymouth, dated 19

September 2024 and which is appended to the main Planning report prepared by B&A. We understand that water from the retention pond will be piped to a proposed soakage area which will be located approximately 150 m east from the retention pond. ENGEO has carried out modelling to simulate the migration of contaminants in groundwater following the discharge. It is assumed that the discharge will occur to an approximately 10 m x 10 m soakage area constructed to a depth of at least 2 m below ground level (bgl).

ENGEO has presented the results of groundwater quality sampling in Appendix 5 of their report. This information has been collected at and in the vicinity of the site from January 2023 to present. The water quality monitoring includes samples collected at the following locations:

- ✧ Samples of Leachate or wastewater from landfill Cells A, B and C, sediment/settlement/retention ponds and sumps. These sample names include Leachate (Cell A), Leachate (Cell B/C), Sediment Pond, SWSDP, Settlement Stage 2, DC/C3 and Deep Sump.
- ✧ Upgradient groundwater spring/seep locations with sample names SPA, SPC and SPE.
- ✧ Downgradient spring/seep locations with sample names TCD1/Discharge, SP1, SP2 and SP3.
- ✧ Upgradient bore locations with sample names MB1, MB2s MB2d and MB3.
- ✧ Downgradient bore locations with sample names MB4s, MB4d, MB5 and TSB1.
- ✧ Upgradient drain locations with sample names TRD3 and TRD4.
- ✧ Downgradient drain locations with sample names SPF Pit/TRD1, TRD2 and TRD5
- ✧ Samples from Noname Creek to the west of the site on the opposite side of the road from the western site perimeter drain with sample names SPF US, SPF MS, SPF DS.

Based on our review of the water quality data we note leachate is being generated at the site and concentrations of many contaminants in leachate and wastewater are significantly elevated. We note that concentrations of dissolved manganese, arsenic and chromium have been recorded in leachate or wastewater well above the Water Services (Drinking Water Standards for New Zealand) Regulations 2022 (DWSNZ), human health based Maximum Acceptable Value (MAV) concentrations. Arsenic has been recorded at nearly 100 x the MAV concentration in one leachate sample. In addition, there are a significant number of leachate or wastewater samples where parameters exceed or significantly exceed the relevant Taumata Arowai Aesthetic Values for Drinking Water Notice 2022 (AVDWN) Guideline Value (GV) concentrations and/or Australia and New Zealand Guidelines for Fresh and Marine Water Quality 2018 (ANZG) Default Guideline Values (DGV) for ecological toxicants.

Samples to determine the impacts of the landfill on groundwater quality are very limited. We note that upgradient bore MB2d is drilled into mudstone and has significantly different water quality results than the other three upgradient bores. This indicates that the source rock may be impacting on the water quality results. As a result, downgradient bore MB4d (also drilled into mudstone) may not be a good bore for assessing the effects on groundwater migrating from the landfill through the more permeable overlying gravel strata (although much higher sulphate concentrations have been recorded in MB4d compared to MB2d which could be caused by leachate). Of the remaining three downgradient bores (MB4s, MB5 and TSB1), only TSB1 has intercepted groundwater and therefore this may be the only representative downgradient bore. As a result, there is significant uncertainty whether the potential effects on groundwater are well constrained. We note TSB1 has recorded dissolved aluminium at up to 19 times the concentrations recorded in the upgradient groundwater bores, however most other relevant parameters are not significantly different between the upgradient and downgradient gravel bores.

Of further significance to water quality we note the following:

- ✧ Samples collected from seeps SP1, SP2 and SP3 (below the landfill in the Taylorville Road cutting) typically show higher parameter concentrations than those measured from upgradient springs SPA, SPC and SPE. In particular, total hardness, electrical conductivity, dissolved calcium and magnesium, sulphate and Chemical Oxygen Demand are many times higher in the downgradient seep locations.
- ✧ Samples from downgradient seep TCD1/Discharge (the seep feeding the western perimeter drain) typically show poorer water quality than the Taylorville Road downgradient seeps, however it should be noted that there are only two individual samples that have been collected from the Taylorville Road seeps compared to nine samples from TCD1/Discharge.
- ✧ Dissolved manganese and dissolved arsenic have been measured at concentrations in TCD1/Discharge exceeding DWSNZ MAV concentrations, while dissolved arsenic, chromium, copper, zinc and manganese, in addition to electrical conductivity, total ammoniacal nitrogen and nitrate have been measured above or well above ecological DGV concentrations in this bore.
- ✧ Concentrations of dissolved aluminium and nitrate have been measured above ecological criteria in the upgradient springs, however significantly higher concentrations for these two parameters have been recorded in either TCD1/Discharge or the downgradient Taylorville Road Springs.
- ✧ Downgradient drain samples SPF Pit/TRD1, TRD2 and TRD5 have recorded higher concentrations for most parameters compared to samples collected from upgradient drains TRD3 and TRD4. All samples (upgradient and downgradient) have been collected from the roadside drain along Taylorville Road. We note dissolved arsenic in SPF Pit/TRD1 has been recorded on one occasion exceeding the DWSNZ MAV, while a significant number of ecological DGV exceedances have occurred for all three downstream drain locations.

Based on the above information, the landfill is generating leachate and it appears that this has impacted on the water quality in downgradient groundwater, springs/seeps and drains. We note that some sample locations have only a very limited number of samples.

Mounding Assessment

The infiltration basin will create a raised mounding of the underlying water table. Based on the proposed discharge of up to 800 m³/day, ENGEO has estimated the potential mounding at the water table. We note the calculations are based on an assumed horizontal hydraulic conductivity (K) of 84 m/d, which is assumed based on text book values for various strata descriptions. We consider the adopted value may not sufficiently take into account the potential for fine grained silt and clay in the matrix and note that the mounding model will potentially predict much higher levels of mounding if K is actually significantly lower. We consider onsite testing would be required to confirm the hydraulic conductivity and typically this would involve pumping testing or slug testing. Infiltration testing would also be advisable to confirm the infiltration capacity of the soils. At present we consider the mounding assessment is highly uncertain. Significant mounding has the potential to impact on the flow direction of contaminants contained in the discharge.

Effects on Groundwater Quality

The level of impacts to groundwater quality onsite are currently not well constrained as there are only limited downgradient sampling results and, as noted above, potentially only one bore within the site with relevant data. Groundwater underlying the site likely discharges into surrounding springs and drains and there may not be hydraulic connection with groundwater on surrounding properties, although that would ultimately need to be confirmed by offsite drilling and monitoring. To date no offsite drilling or monitoring has been carried out.

Our concerns regarding the actual hydraulic conductivity of the strata underlying the site (mentioned above in our comment on the groundwater mounding assessment) apply equally to the RISC5 contaminant transport modelling carried out by ENGEO. We consider that lower values of hydraulic conductivity may result in higher predicted contaminant concentrations due to less dilution and dispersion in the aquifer and consider that to more comprehensively assess the potential for contaminant migration further simulations using lower values of hydraulic conductivity should be carried out.

We also note that the model source concentrations are stated to be the highest concentrations detected in samples collected from locations in the water treatment system in July 2024. However, we note much higher concentrations have been recorded in the past than the input concentrations that have been used. While the most recent sample results may potentially be relevant for an initial discharge, if discharges are still occurring for up to 6 months, these later discharges could contain much greater concentrations than those modelled, as the impacts of rainfall dilution may be less than the current situation.

Based on the above comments, we consider additional mounding and model simulations would be required, by undertaking a sensitivity analysis of the model input parameters, to adequately address the potential effects on groundwater quality and the resulting impacts on the natural surface water discharge areas. We consider the risk to offsite groundwater (from direct groundwater flow) is likely low due to an expected lack of hydraulic connection, however the level of information provided to date to confirm that is currently low.

We also consider that while there are a significant number of samples that have been collected at the site to date, some sample locations do not have a lot of data. In addition, VOC and SVOCs are not included in the parameters that are sampled and therefore we can't be sure that all potential contaminants are sufficiently characterised.

Effects on Groundwater Quantity

ENGEO has assessed the effects on groundwater quantity based on the results of their mounding assessment. As described above, we consider the mounding assessment is highly uncertain and recommend that further simulations based on lower hydraulic conductivity values should be carried out. Ideally those simulations should be based on testing data from the site to constrain the hydraulic conductivity.

Effects on Surface Water Quality

Effects on surface water quality have been assessed by ENGEO in Noname Creek and the Grey River. The potential lack of connection between groundwater at the site and surface water in Noname Creek is noted by ENGEO and we consider this is likely relevant (although if mounding from the discharge is very high there could be potential for discharge to Noname Creek). In addition, dilution in the Grey River will be very large with minimal impacts. However, ENGEO has not assessed the potential effects in the drain that begins at TCD1, the extension of that drain along Taylorville Road and the surface water bodies on the alluvial gravels on the south side of Taylorville Road that receive flow inputs from the roadside drain.

Discharges from the landfill have impacted on these drains in the past and we consider the proposed discharge will likely result in further exceedances (possibly by significant amounts) of ecological limits in the drains which have not been considered. Further ecological information has been requested in the past, relevant to a previous application by TRPL, but has not been provided. As a result, we are unsure of the ecological status of the roadside drain or the surface water bodies on the southern side of the road. Therefore, based on the information provided, we can't confirm that the effects on surface water quality or ecology within these surface water bodies will be less than minor.

Effects on GDC Water Supply

As per our October 2024 report, we expect the risk to the GDC water supply from the operation of the TRPL landfill is likely low. However, we also note the supply services around 10,000 people and a landfill operation would typically not be appropriate within such close proximity to a public water supply. The proposed discharge to ground results in an additional risk to the GDC supply compared to that considered by PDP previously, as the applicant now seeks to discharge a large quantity of contaminated water directly to ground over a short time period. In addition, the proposed discharge area will be closer to an upgradient location, relative to the water supply, than most of the landfill cells and treatment system.

We consider the risk to the water supply from direct groundwater flow is still likely low, as there is unlikely to be a connection between the gravels underlying the landfill and those south of Taylorville Road (where the GDC bores are located). However, that is based solely on an assumed conceptual understanding of the groundwater and surface water system on the southern side of Taylorville Road. It would be best to confirm that is the case via a drilling and monitoring programme which has not been carried out to date. In addition, this assumes there is no potential for fracture flow in the mudstone that underlies the gravels at the site. ENGEO has previously indicated that clay minerals within the mudstone will swell and seal any fractures that may be present, however they have not provided any information to confirm the clay content within the mudstone underlying the site. This information could be provided via X-Ray Diffraction (XRD) analysis.

While the risk from direct groundwater flow is likely low between the upper and lower terrace, there is the potential for transport of contaminants from the site to the southern terrace via the roadside drain on Taylorville Road and the culverts that transport the water under the road.

We expect any water that makes its way under the road at the location of TRD5 is unlikely to impact on the water supply as it will be further separated from the water supply bores by other surface water features on the lower terrace.

The main risk is likely from flow under the road via the culvert at TRD2. Figure 1 of the March 2024 EHS Support report shows that the drain at TRD2 could receive seepage and flow from the drain impacted by leachate. There is a potential transport mechanism via the associated stream on the southern side of Taylorville Road where losses of flow to the underlying gravels could allow contaminants to reach the strata from which the GDC supply bores abstract water. In reality we expect that the GDC bores will derive most of their water from the Grey River and the extent of input from the gravels on the landfill side of the aquifer will be very limited. This would likely significantly dilute any contaminants that may reach the southern side of Taylorville Road. However, we also note we have not been provided with any information to confirm the performance of the GDC bores (yield versus drawdown in individual supply bores and drawdown interference effects induced in other supply bores) to confirm that is the case. In addition, water quality sampling comparing the GDC supply bore water, Grey River water, groundwater and drain water would provide useful information to determine the potential for upgradient groundwater to reach the GDC supply bores.

Furthermore, it can't be ruled out that the concentrations of impacted water that reach the roadside drain as a result of the proposed discharge won't be higher than those that have reached the drain in the past, due to the discharge of a significantly increased mass of contaminants to ground over a short time period. We recommend further modelling simulations should be carried out to address this uncertainty. Even after that, detailed monitoring of the effects from any discharge is likely to be required.

Conclusion

Based on our review of the ENGEO assessments, we consider their assessments could potentially underestimate the effects of the proposed discharge. We consider the hydraulic conductivity used in their

modelling could be too high and lower values of hydraulic conductivity would result in more mounding and may also result in higher contaminant concentrations in groundwater and surface water. In addition, we consider the input contaminant concentrations in their RISC5 model may be too low which will also result in lower output concentrations. As a result, further modelling could be provided.

We also consider that ENGEO should provide further information to confirm the potential surface water quality effects on the perimeter drains on the western boundary and along Taylorville Road and the potential ecological impacts on the drains surrounding the site and the downstream surface water features on the southern side of Taylorville Road.

Additional information should also be provided to confirm the potential for contamination of the GDC water supply via surface water impacted from the landfill which may cross Taylorville Road via the culvert at TRD2.

It is important to note that irrespective of the results of any further modelling, the assessment will still be highly theoretical and subject to uncertainty. The most certain way to assess the potential for impacts on the GDC water supply are via a detailed drilling and monitoring programme on the southern side of Taylorville Road. We have recommended such monitoring on a number of occasions but to date nothing has been carried out.

Without a site-specific monitoring programme on the GDC land and if WCRC decide to grant the consent application, it will be important that a robust monitoring programme is incorporated into consent conditions. We note the proposed monitoring programme set out in Section 7 of the ENGEO report recommends sampling from a number of groundwater monitoring bores downgradient of the discharge. We consider further bores should be added directly south of MB13 in case groundwater mounding is significant and affects groundwater flow direction. In addition, MB17 should also be monitored along with spring seeps and drain sample locations along Taylorville Road. It would be good for PDP to review any consent conditions prior to granting of a consent.

Limitations

This memorandum has been prepared by Pattle Delamore Partners Limited (PDP) on the basis of information provided by West Coast Regional Council and others (not directly contracted by PDP for the work), including Planz, Engeo and Barkers and Associates. PDP has not independently verified the provided information and has relied upon it being accurate and sufficient for use by PDP in preparing the memorandum. PDP accepts no responsibility for errors or omissions in, or the currency or sufficiency of, the provided information.

This memorandum has been prepared by PDP on the specific instructions of West Coast Regional Council for the limited purposes described in the memorandum. PDP accepts no liability if the memorandum is used for a different purpose or if it is used or relied on by any other person. Any such use or reliance will be solely at their own risk.

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Prepared by



Carl Steffens

Technical Director - Water Resources

Site Memorandum

Project Name	Taylorville Resource Park		
Memo Topic	RC-2024-0122 and RC-2021-0097-V10 - RFI Response		
Memo Date	30 January 2025	Date of Site Visit	N/A
ENGEO Memo No.	08	ENGEO Project No.	26200.000.001
To	Taylorville Resource Park Limited	Copy	Barker and Associates
Author	Lyn Nugent	Reviewer	Dave Robotham

ENGEO has reviewed the further information request relating to Taylorville Resource Park Limited's (TRPL) application to discharge collected leachate and stormwater to land (RC-2024-0012) and the application (RC2021.0097-V10) to change consent condition 11 of consent RC2021.0097. The further information request was issued by the West Coast Regional Council (WCRC, dated 19-Dec-24) under section 92(1) of the Resource Management Act 1991.

WCRC's questions (in grey) and ENGEO responses (in blue) are provided below.

1. Hydraulic conductivity estimates from onsite testing (pump testing and/or slug testing). If slug testing is used, then please provide testing data from at least 3 bore locations on the upper terrace surface

To assess for the hydraulic conductivity in the Pleistocene gravels, aquifer testing was undertaken at three on-site groundwater monitoring bores (MB13, MB14, and MB19d). Based on an assessment of the slug testing data using the method for unconfined aquifers (Bouwer & Rice, 1976) (refer Appendix 1), the hydraulic conductivity of the Pleistocene gravels ranges from 1×10^{-6} m/s to 1×10^{-5} m/s.

2. Updated mounding and contaminant modelling simulations using the hydraulic conductivity estimates derived from the requested testing. Based on the testing results, please take into account that the results from testing (particularly slug testing) may provide local values for hydraulic conductivity but not necessarily be representative of the average hydraulic conductivity over the area of the landfill. Therefore, please provide multiple modelling simulations informed by the testing results using a realistic range of potential hydraulic conductivity values for the strata beneath the upper terrace surface (i.e. a sensitivity assessment based on varying the hydraulic conductivity that is informed by the testing results).

Updated mounding using the Hantush method (Hantush, 1967) is provided in Appendix 2. As stated in our memo dated 12 November 2024, the actual discharge volumes will depend on the infiltration capacity in the soakage area, rainfall and evaporation rates. Using the Green-Ampt equation and based on the proposed location and aerial extent of the soakage area, the anticipated infiltration capacity of the soakage is on the order of 200 m³/d – 300 m³/d. This generally corresponds to the maximum potential mounding beneath the soakage area (i.e., the mound height that will intersect with the base of the soakage area). To be conservative, the RISC5 model was updated to include a source area with a height of 8 m (assuming seasonal groundwater fluctuations and mounding to the base of the soakage area). Additionally, the RISC5 model was re-run using the maximum (1.0×10^{-5} m/s) and minimum (1.2×10^{-6} m/s) values for hydraulic conductivity. As shown in Appendix 3, concentrations at the boundary are highest in the maximum conductivity scenario; however, even under this scenario, the maximum modelled concentrations at the boundary are either not detectable or orders of magnitude below the human or ecological health-based criteria (0.0069 mg/L, nitrate-N; 0.0027 mg/L total aluminium).

3. Updated contaminant modelling simulations using input parameters more consistent with high parameter concentrations that have been measured in leachate and wastewater.

Whilst we think that it is highly unlikely that blended concentrations of leachate (in the soakage area) will contain concentrations equivalent to the highest parameter concentrations ENGEO detected at the site, the RISC5 model was re-run using the maximum concentrations detected in the water treatment system locations since sampling began. As shown in Appendix 3, the maximum modelled concentrations at the boundary are either not detectable or orders of magnitude below the human or ecological health-based criteria.

4. An ecological survey of the surface water features and aquatic habitat surrounding the site, including Noname Creek, the drain along Taylorville Road, including that part of the drain that begins at spring seep TCD1 and the streams on the southern side of Taylorville Road (excluding the Grey River) and an impact assessment on surface water quality and ecological values resulting from the proposed discharge.

The ecological survey will be provided as a separate memo (Appendix 4)

5. A water quality comparison between water samples collected from the GDC water supply, the Grey River, groundwater upgradient of the landfill north of Taylorville Road and surface water samples collected at TRD2 and TRD4. The samples should include analysis for anions and cations and nitrogen species, in addition to other parameters sampled to date by TRPL. The purpose of the sampling is to inform whether the GDC water supply source is dominantly Grey River water and that there is minimal input to the supply bores from upgradient groundwater and surface drainage.

Water quality samples were collected from the raw (untreated) water at GRE001 (PWT01) and the Grey River approximately 300 – 400 m upstream of GRE001 (GR-01; Figure A5.1). On the day of our sampling visit, all three active bores (Bore 1–Bore 3) were operating. The samples were submitted for laboratory analysis for dissolved metals and ions in solution (Table 5.1). These results were converted to milliequivalents and compared to the most recent (October 24) results for samples collected from monitoring wells along the southern side boundary adjacent to the current operations (MB4d, MB14, MB18; Figure A5.2) and the full boundary extent, including upgradient / cross-gradient of the operations (MB2s/d, MB3, MB4d, MB14, MB18; Figure A5.3).

As shown in Figure A5.2, the ion profile of PWT01 very closely matches that of GR-01, indicating that there is minimal input from waters downgradient of the site operations. When the samples collected from upgradient / cross-gradient wells are included (Figure A5.3) there is a greater distribution of profiles, with more similarity in anions between the upgradient / cross-gradient wells and PWT01. However, the ion profile of PWT-1 is still most closely matched to that of GR-01 and broad similarity is expected, as all samples were collected from the same catchment.

We do note that the profile of MB2d was anomalous relative to the other samples. Water quality results from deep upgradient wells (MB2d and MB19d) are noticeably different from other locations at the site—including shallow wells MB2s and MB19s (Table 5.1)—potentially due to inputs from the underlying Kaiata Mudstone formation. The remaining cross-gradients wells cluster tightly, indicating general homogeneity in the waters above the mudstone contact. As such, MB2d was not considered further in this assessment.

The concentrations of ions in solution indicates that water quality at PWT01 is dominantly influenced by the Grey River, with no demonstrable input from the site. We do note the concentrations of dissolved metals (copper, lead, nickel, zinc) in PWT01 differed significantly from GR-01 and samples collected along the southern site boundary. The cause of this has not been assessed but may be naturally occurring (e.g., concentrations in the Holocene gravels) or may indicate the influence of waters originating at another (offsite) source area.

6. Yield versus drawdown data from the GDC supply bores to confirm the gravel deposits immediately adjacent to the Grey River are of high permeability with a high hydraulic connection to the Grey River.

Based on aquifer testing information provided by GDC (Appendix 6), the hydraulic conductivity in the Holocene river gravels at BH02 (abstraction well for public supply GRE001) is between 3.42×10^{-2} m/s (Boulton method) and 1.58×10^{-2} m/s (Neuman method). Pumping test analysis reports are provided in Appendix 6.

7. XRD sample analysis from the Kaiata mudstone underlying the landfill to confirm clay mineralogy in the mudstone and confirm ENGEO comments that swelling within any fractures that may be present in the mudstone will block the transmission of groundwater via a fracture network.

XRD analysis of two mudstone outcrops (Leaver, 1999) show that it is comprised of a mixture of kaolinite, illite, and vermiculite (Appendix 6). As the mineralogy of the Kaiata Mudstone comprises a mixture of swelling and non-swelling clay minerals, it is inferred that there is negligible flow through defects (e.g., microfractures) due to clay swelling. There is potential for preferential pathways through interbedded layers of non-clay materials (e.g., silts, sands). Groundwater that intersects with these layers will most likely be transported west-southwest, along the dip of the mudstone bedding planes and away from GRE001.

Based on our assessment works to date and in conjunction with the proposed conditions of consent, it is deemed *highly unlikely* that the proposed discharges will have a deleterious effect on human or ecological health.

This memorandum has been prepared for the use of our client, their professional advisers and the relevant Territorial Authorities in relation to the specified project brief described in this memo. No liability is accepted for the use of any part of the report for any other purposes or by any other person or entity. This Limitation should be read in conjunction with the Engineering NZ / ACENZ Standard Terms of Engagement.

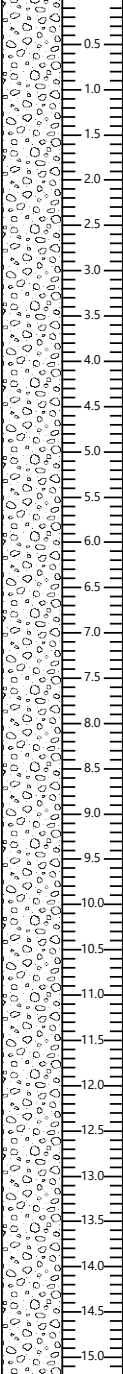
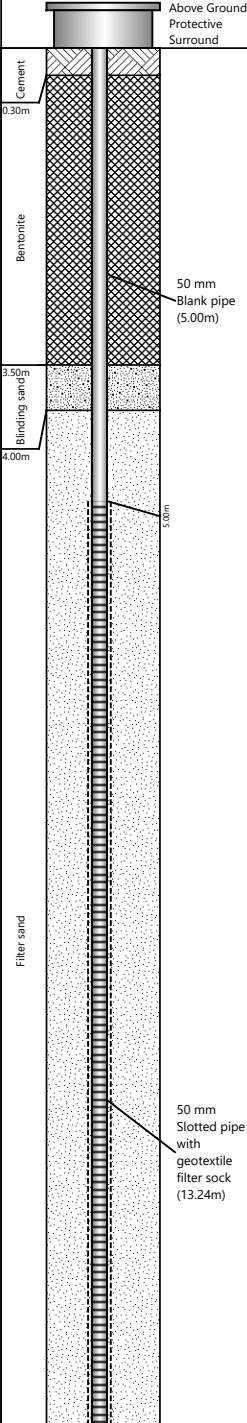
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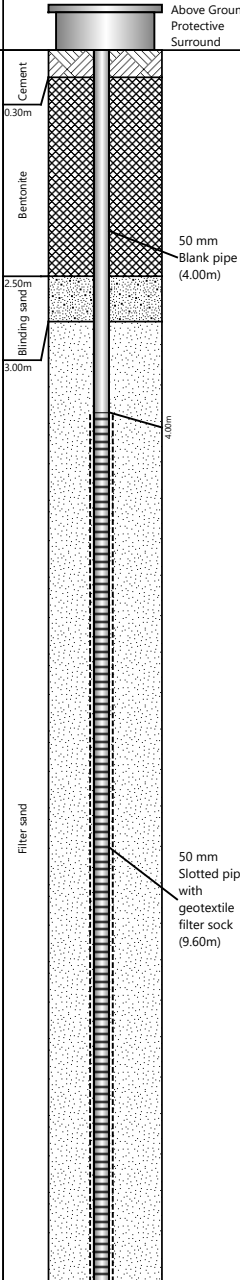
APPENDIX 1

Slug Testing Reports

	Client:	ENGEO	Bore No.:	BH013					
	Project:	287 Taylorville Road, Coal Creek	Job No.:	22638					
Site Location: 287 Taylorville Road, Coal Creek		Date: 3/10/2024 - 3/10/2024							
Grid Reference: 1457182.23m E, 5299473.42m N (NZTM) - Handheld GPS		Rig Operator: E. Diaz Farias							
Elevation: 0.00m		Datum: Ground							
		Equipment: Geoprobe 8140LC - track							
Description	Graphic Log	Depth	Method	Runs and recovery	Water at hole depth	SPT (NZS 4402) N-value Uncorrected	In-Situ Tests	Samples - Permeability tests	Installation & Resources
Sandy fine to coarse GRAVEL with minor cobbles; light brown. Well graded.			114mm OD Sonic core drilling	25 50 75 100% 100% 100% 100% 100% 100% 100% 92% 39% 78% 78%		0 10 20 30 40 50 60			
Remarks									Additional Resources: Core boxes: 7 Water added: 2000 liters Concrete: 3 bags Bentonite: 2 bags Blinding sand: 0.25 bag Filter sand: 6 bags
120 High Street, Southbridge 7602, Canterbury, New Zealand ph: (03) 324 2571 fax: (03) 324 2431 web: www.drilling.co.nz									Hole Depth: 18.24m Page 1 of 2

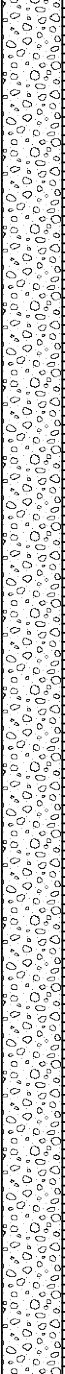
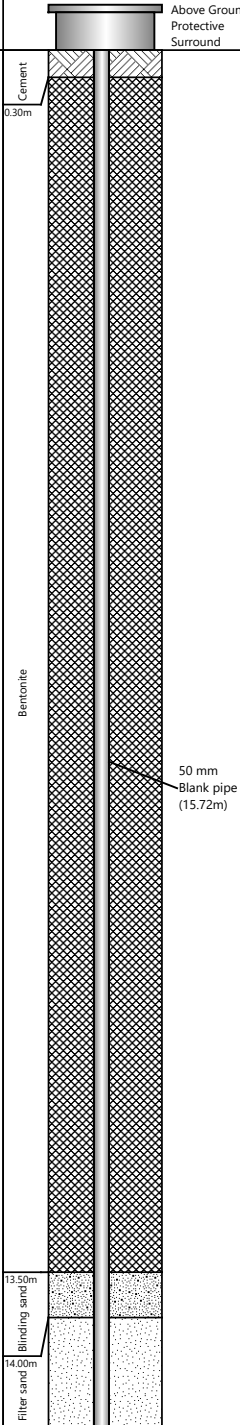
Client:	ENGEO	Bore No.:	BH014
Project:	287 Taylorville Road, Coal Creek	Job No.:	22638

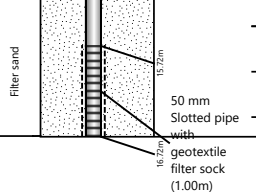
Site Location: 287 Taylorville Road, Coal Creek	Date: 23/9/2024 - 25/9/2024
Grid Reference: 1456883.14m E, 5299346.95m N (NZTM) - Handheld GPS	Rig Operator: E. Diaz Farias
Elevation: 0.00m	Equipment: Geoprobe 8140LC - track
Datum: Ground	

Description	Graphic Log	Depth	Method	Runs and recovery	Water at hole depth	SPT (NZS 4402) N-value Uncorrected	In-Situ Tests	Samples - Permeability tests	Installation & Resources
Sandy fine to coarse GRAVEL; light brown. Well graded.		0.5		25 50 75		0 10 20 30 40 50 60			
WOOD with some rubbish.		1.00m		65%					
Sandy fine to coarse GRAVEL with minor cobbles; light brown. Well graded.		2.50m		59%					
		3.00m		59%					
		4.00m		59%					
		5.00m		100%					
		5.50m		32%					
		6.00m		32%					
		7.00m		32%					
		8.00m		36%					
		8.50m		32%					
		9.00m		32%					
		9.50m		32%					
		10.00m		32%					
Silty fine to coarse GRAVEL; reddish brown. Well graded.		10.30m		100%					
Grey MUDSTONE.		11.40m		100%					
		12.00m		100%					
		12.50m		100%					
		13.00m		100%					
		13.50m		100%					

EOH: 13.68m	5.70m bgl
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Remarks Hammer energy ratio 84.3%	Additional Resources: Core boxes: 5 Water added: 6000 liters Concrete: 3 bags Bentonite: 1.5 bags Blasting sand: 0.25 bag Filter sand: 5 bags
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	Client: ENGEO		Bore No.: BH019						
	Project: 287 Taylorville Road, Coal Creek		Job No.: 22638						
Site Location: 287 Taylorville Road, Coal Creek									
Date: 1/10/2024 - 2/10/2024									
Grid Reference: 1457351.03m E, 5299489.22m N (NZTM) - Handheld GPS									
Rig Operator: E. Diaz Farias									
Elevation: 0.00m									
Datum: Ground									
Equipment: Geoprobe 8140LC - track									
Description	Graphic Log	Depth	Method	Runs and recovery	Water at hole depth	SPT (NZS 4402) N-value Uncorrected	In-Situ Tests	Samples - Permeability tests	Installation & Resources
Sandy fine to coarse GRAVEL with minor cobbles; light brown. Well graded.		0.5 1.0 1.5 2.0 2.5 3.0 3.5 4.0 4.5 5.0 5.5 6.0 6.5 7.0 7.5 8.0 8.5 9.0 9.5 10.0 10.5 11.0 11.5 12.0 12.5 13.0 13.5 14.0 14.5 15.0	114mm OD Sonic core drilling	25		0			
				50		10			
				75		20			
				100%		30			
				100%		40			
				100%		50			
				72%		60			
				100%					
				65%					
				59%					
				39%					
				72%					
				59%					
				65%					
				13.50m		Blinding sand			
14.00m	Filter sand								
Remarks									Additional Resources: Core boxes: 6 Water added: 1500 liters Concrete: 3 bags Bentonite: 1.5 bags Blinding sand: 0.25 bag Filter sand: 6 bags
120 High Street, Southbridge 7602, Canterbury, New Zealand ph: (03) 324 2571 fax: (03) 324 2431 web: www.drilling.co.nz									Hole Depth: 16.72m Page 1 of 2

	Client: ENGEO		Bore No.: BH019						
	Project: 287 Taylorville Road, Coal Creek		Job No.: 22638						
Site Location: 287 Taylorville Road, Coal Creek									
Date: 1/10/2024 - 2/10/2024									
Grid Reference: 1457351.03m E, 5299489.22m N (NZTM) - Handheld GPS									
Rig Operator: E. Diaz Farias									
Elevation: 0.00m									
Datum: Ground									
Equipment: Geoprobe 8140LC - track									
Description	Graphic Log	Depth	Method	Runs and recovery	Water at hole depth	SPT (NZS 4402) N-value Uncorrected	In-Situ Tests	Samples - Permeability tests	Installation & Resources
Sandy fine to coarse GRAVEL with minor cobbles; light brown. Well graded.		15.5	114mm OD Sonic core drilling	25 50 75 78%		0 10 20 30 40 50 60			
Grey MUDSTONE.		16.00m							
EOH: 16.72m									
Remarks									Additional Resources: Core boxes: 6 Water added: 1500 liters Concrete: 3 bags Bentonite: 1.5 bags Blinding sand: 0.25 bag Filter sand: 6 bags
120 High Street, Southbridge 7602, Canterbury, New Zealand ph: (03) 324 2571 fax: (03) 324 2431 web: www.drilling.co.nz									Hole Depth: 16.72m Page 2 of 2

				Slug Test - Water Level Data		Page 1 of 8	
				Project: TRP			
				Number: 26200.000.001			
				Client: TRPL			
Location: Taylorville, Coal Creek			Slug Test: MB19d in			Test Well: MB19d	
Test Conducted by: J. Green / M. Knopick Test Date: 16/10/2024							
Water level at t=0 [m]: 4.23			Static Water Level [m]: 4.91			Water level change at t=0 [m]: -0.68	
	Time [s]	Water Level [m]	WL Change [m]				
1	1	4.2273	-0.6827				
2	2	4.2985	-0.6115				
3	3	4.3191	-0.5909				
4	4	4.329	-0.581				
5	5	4.3895	-0.5205				
6	6	4.403	-0.507				
7	7	4.4656	-0.4444				
8	8	4.5254	-0.3846				
9	9	4.5403	-0.3697				
10	10	4.5503	-0.3597				
11	11	4.5695	-0.3405				
12	12	4.5951	-0.3149				
13	13	4.6093	-0.3007				
14	14	4.5958	-0.3142				
15	15	4.6022	-0.3078				
16	16	4.6178	-0.2922				
17	17	4.6306	-0.2794				
18	18	4.6242	-0.2858				
19	19	4.583	-0.327				
20	20	4.6257	-0.2843				
21	21	4.6392	-0.2708				
22	22	4.6555	-0.2545				
23	23	4.6655	-0.2445				
24	24	4.6726	-0.2374				
25	25	4.6783	-0.2317				
26	26	4.6833	-0.2267				
27	27	4.6882	-0.2218				
28	28	4.6918	-0.2182				
29	29	4.6961	-0.2139				
30	30	4.6939	-0.2161				
31	31	4.6761	-0.2339				
32	32	4.6889	-0.2211				
33	33	4.7025	-0.2075				
34	34	4.7032	-0.2068				
35	35	4.6897	-0.2203				
36	36	4.664	-0.246				
37	37	4.7032	-0.2068				
38	38	4.7188	-0.1912				
39	39	4.7295	-0.1805				
40	40	4.7366	-0.1734				
41	41	4.7423	-0.1677				
42	42	4.7458	-0.1642				
43	43	4.7494	-0.1606				
44	44	4.7522	-0.1578				
45	45	4.7551	-0.1549				
46	46	4.7579	-0.1521				
47	47	4.7608	-0.1492				
48	48	4.7629	-0.1471				
49	49	4.7658	-0.1442				
50	50	4.7679	-0.1421				
51	51	4.77	-0.14				
52	52	4.7722	-0.1378				

				Slug Test - Water Level Data	Page 2 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
53	53	4.7743	-0.1357		
54	54	4.7764	-0.1336		
55	55	4.7786	-0.1314		
56	56	4.7807	-0.1293		
57	57	4.7828	-0.1272		
58	58	4.7842	-0.1258		
59	59	4.7864	-0.1236		
60	60	4.7885	-0.1215		
61	61	4.7899	-0.1201		
62	62	4.7921	-0.1179		
63	63	4.7935	-0.1165		
64	64	4.7956	-0.1144		
65	65	4.797	-0.113		
66	66	4.7992	-0.1108		
67	67	4.8013	-0.1087		
68	68	4.8027	-0.1073		
69	69	4.8041	-0.1059		
70	70	4.8056	-0.1044		
71	71	4.807	-0.103		
72	72	4.8091	-0.1009		
73	73	4.8105	-0.0995		
74	74	4.812	-0.098		
75	75	4.8134	-0.0966		
76	76	4.8148	-0.0952		
77	77	4.8162	-0.0938		
78	78	4.8177	-0.0923		
79	79	4.8191	-0.0909		
80	80	4.8205	-0.0895		
81	81	4.8219	-0.0881		
82	82	4.8205	-0.0895		
83	83	4.8226	-0.0874		
84	84	4.8248	-0.0852		
85	85	4.8262	-0.0838		
86	86	4.8276	-0.0824		
87	87	4.829	-0.081		
88	88	4.8297	-0.0803		
89	89	4.8312	-0.0788		
90	90	4.8319	-0.0781		
91	91	4.8333	-0.0767		
92	92	4.834	-0.076		
93	93	4.8354	-0.0746		
94	94	4.8368	-0.0732		
95	95	4.8376	-0.0724		
96	96	4.839	-0.071		
97	97	4.8397	-0.0703		
98	98	4.8404	-0.0696		
99	99	4.8411	-0.0689		
100	100	4.8418	-0.0682		
101	101	4.8432	-0.0668		
102	102	4.844	-0.066		
103	103	4.8454	-0.0646		
104	104	4.8461	-0.0639		
105	105	4.8468	-0.0632		
106	106	4.8475	-0.0625		
107	107	4.8482	-0.0618		
108	108	4.8489	-0.0611		
109	109	4.8496	-0.0604		

				Slug Test - Water Level Data	Page 3 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
110	110	4.8511	-0.0589		
111	111	4.8518	-0.0582		
112	112	4.8525	-0.0575		
113	113	4.8532	-0.0568		
114	114	4.8539	-0.0561		
115	115	4.8546	-0.0554		
116	116	4.8553	-0.0547		
117	117	4.856	-0.054		
118	118	4.8568	-0.0532		
119	119	4.8575	-0.0525		
120	120	4.8582	-0.0518		
121	121	4.8589	-0.0511		
122	122	4.8589	-0.0511		
123	123	4.8596	-0.0504		
124	124	4.8603	-0.0497		
125	125	4.861	-0.049		
126	126	4.8617	-0.0483		
127	127	4.8624	-0.0476		
128	128	4.8631	-0.0469		
129	129	4.8639	-0.0461		
130	130	4.8639	-0.0461		
131	131	4.8646	-0.0454		
132	132	4.8653	-0.0447		
133	133	4.866	-0.044		
134	134	4.866	-0.044		
135	135	4.8667	-0.0433		
136	136	4.8674	-0.0426		
137	137	4.8674	-0.0426		
138	138	4.8681	-0.0419		
139	139	4.8688	-0.0412		
140	140	4.8695	-0.0405		
141	141	4.8695	-0.0405		
142	142	4.8703	-0.0397		
143	143	4.871	-0.039		
144	144	4.871	-0.039		
145	145	4.8717	-0.0383		
146	146	4.8695	-0.0405		
147	147	4.871	-0.039		
148	148	4.8717	-0.0383		
149	149	4.8731	-0.0369		
150	150	4.8738	-0.0362		
151	151	4.8738	-0.0362		
152	152	4.8745	-0.0355		
153	153	4.8752	-0.0348		
154	154	4.8752	-0.0348		
155	155	4.8759	-0.0341		
156	156	4.8766	-0.0334		
157	157	4.8766	-0.0334		
158	158	4.8766	-0.0334		
159	159	4.8774	-0.0326		
160	160	4.8781	-0.0319		
161	161	4.8781	-0.0319		
162	162	4.8781	-0.0319		
163	163	4.8788	-0.0312		
164	164	4.8788	-0.0312		
165	165	4.8788	-0.0312		
166	166	4.8795	-0.0305		

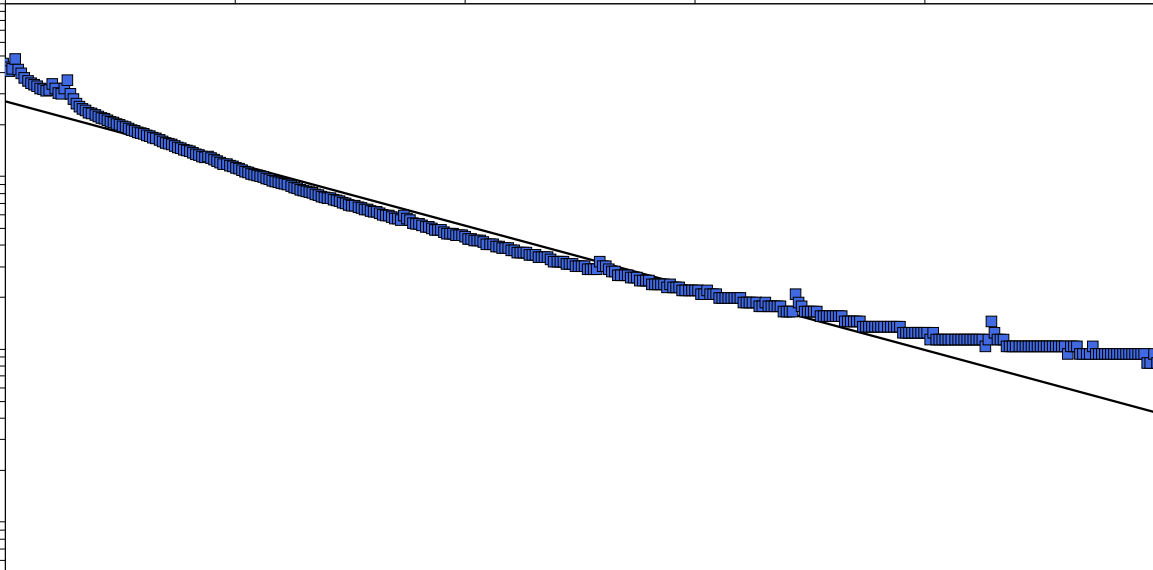
				Slug Test - Water Level Data	Page 4 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
167	167	4.8802	-0.0298		
168	168	4.8802	-0.0298		
169	169	4.8809	-0.0291		
170	170	4.8809	-0.0291		
171	171	4.8809	-0.0291		
172	172	4.8816	-0.0284		
173	173	4.8823	-0.0277		
174	174	4.8823	-0.0277		
175	175	4.8823	-0.0277		
176	176	4.883	-0.027		
177	177	4.883	-0.027		
178	178	4.8838	-0.0262		
179	179	4.8837	-0.0263		
180	180	4.8837	-0.0263		
181	181	4.8845	-0.0255		
182	182	4.8845	-0.0255		
183	183	4.8852	-0.0248		
184	184	4.8852	-0.0248		
185	185	4.8852	-0.0248		
186	186	4.8852	-0.0248		
187	187	4.8859	-0.0241		
188	188	4.8859	-0.0241		
189	189	4.8859	-0.0241		
190	190	4.8866	-0.0234		
191	191	4.8866	-0.0234		
192	192	4.8866	-0.0234		
193	193	4.8866	-0.0234		
194	194	4.8873	-0.0227		
195	195	4.888	-0.022		
196	196	4.888	-0.022		
197	197	4.888	-0.022		
198	198	4.888	-0.022		
199	199	4.8887	-0.0213		
200	200	4.8887	-0.0213		
201	201	4.8887	-0.0213		
202	202	4.8894	-0.0206		
203	203	4.8894	-0.0206		
204	204	4.8894	-0.0206		
205	205	4.8894	-0.0206		
206	206	4.8901	-0.0199		
207	207	4.8901	-0.0199		
208	208	4.8901	-0.0199		
209	209	4.8901	-0.0199		
210	210	4.888	-0.022		
211	211	4.8894	-0.0206		
212	212	4.8894	-0.0206		
213	213	4.8901	-0.0199		
214	214	4.8908	-0.0192		
215	215	4.8908	-0.0192		
216	216	4.8916	-0.0184		
217	217	4.8916	-0.0184		
218	218	4.8916	-0.0184		
219	219	4.8916	-0.0184		
220	220	4.8923	-0.0177		
221	221	4.8923	-0.0177		
222	222	4.8923	-0.0177		
223	223	4.893	-0.017		

				Slug Test - Water Level Data	Page 5 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
224	224	4.893	-0.017		
225	225	4.893	-0.017		
226	226	4.893	-0.017		
227	227	4.8937	-0.0163		
228	228	4.8937	-0.0163		
229	229	4.8937	-0.0163		
230	230	4.8937	-0.0163		
231	231	4.8937	-0.0163		
232	232	4.8944	-0.0156		
233	233	4.8937	-0.0163		
234	234	4.8944	-0.0156		
235	235	4.8944	-0.0156		
236	236	4.8944	-0.0156		
237	237	4.8951	-0.0149		
238	238	4.8951	-0.0149		
239	239	4.8951	-0.0149		
240	240	4.8951	-0.0149		
241	241	4.8951	-0.0149		
242	242	4.8951	-0.0149		
243	243	4.8958	-0.0142		
244	244	4.8958	-0.0142		
245	245	4.8951	-0.0149		
246	246	4.8958	-0.0142		
247	247	4.8958	-0.0142		
248	248	4.8958	-0.0142		
249	249	4.8965	-0.0135		
250	250	4.8965	-0.0135		
251	251	4.8965	-0.0135		
252	252	4.8965	-0.0135		
253	253	4.8965	-0.0135		
254	254	4.8965	-0.0135		
255	255	4.8965	-0.0135		
256	256	4.8965	-0.0135		
257	257	4.8972	-0.0128		
258	258	4.8972	-0.0128		
259	259	4.8972	-0.0128		
260	260	4.8972	-0.0128		
261	261	4.8972	-0.0128		
262	262	4.8979	-0.0121		
263	263	4.8979	-0.0121		
264	264	4.8972	-0.0128		
265	265	4.8979	-0.0121		
266	266	4.8979	-0.0121		
267	267	4.8979	-0.0121		
268	268	4.8979	-0.0121		
269	269	4.8979	-0.0121		
270	270	4.8987	-0.0113		
271	271	4.8987	-0.0113		
272	272	4.8987	-0.0113		
273	273	4.8987	-0.0113		
274	274	4.8958	-0.0142		
275	275	4.8972	-0.0128		
276	276	4.8979	-0.0121		
277	277	4.8987	-0.0113		
278	278	4.8987	-0.0113		
279	279	4.8987	-0.0113		
280	280	4.8987	-0.0113		

				Slug Test - Water Level Data	Page 6 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
281	281	4.8987	-0.0113		
282	282	4.8994	-0.0106		
283	283	4.8994	-0.0106		
284	284	4.8994	-0.0106		
285	285	4.8994	-0.0106		
286	286	4.8994	-0.0106		
287	287	4.8994	-0.0106		
288	288	4.8994	-0.0106		
289	289	4.8994	-0.0106		
290	290	4.9001	-0.0099		
291	291	4.9001	-0.0099		
292	292	4.9001	-0.0099		
293	293	4.9001	-0.0099		
294	294	4.9001	-0.0099		
295	295	4.9001	-0.0099		
296	296	4.9008	-0.0092		
297	297	4.9008	-0.0092		
298	298	4.9008	-0.0092		
299	299	4.9008	-0.0092		
300	300	4.9008	-0.0092		
301	301	4.9008	-0.0092		
302	302	4.9008	-0.0092		
303	303	4.9008	-0.0092		
304	304	4.9008	-0.0092		
305	305	4.9008	-0.0092		
306	306	4.9008	-0.0092		
307	307	4.9008	-0.0092		
308	308	4.9008	-0.0092		
309	309	4.9015	-0.0085		
310	310	4.9015	-0.0085		
311	311	4.9015	-0.0085		
312	312	4.9015	-0.0085		
313	313	4.9015	-0.0085		
314	314	4.9015	-0.0085		
315	315	4.9015	-0.0085		
316	316	4.9015	-0.0085		
317	317	4.9015	-0.0085		
318	318	4.9022	-0.0078		
319	319	4.9015	-0.0085		
320	320	4.9022	-0.0078		
321	321	4.9022	-0.0078		
322	322	4.9022	-0.0078		
323	323	4.9022	-0.0078		
324	324	4.9022	-0.0078		
325	325	4.9022	-0.0078		
326	326	4.9022	-0.0078		
327	327	4.9022	-0.0078		
328	328	4.9022	-0.0078		
329	329	4.9022	-0.0078		
330	330	4.9022	-0.0078		
331	331	4.9022	-0.0078		
332	332	4.9022	-0.0078		
333	333	4.9022	-0.0078		
334	334	4.9022	-0.0078		
335	335	4.9022	-0.0078		
336	336	4.9029	-0.0071		
337	337	4.9022	-0.0078		

				Slug Test - Water Level Data	Page 7 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
338	338	4.9001	-0.0099		
339	339	4.9015	-0.0085		
340	340	4.9022	-0.0078		
341	341	4.9022	-0.0078		
342	342	4.9022	-0.0078		
343	343	4.9029	-0.0071		
344	344	4.9029	-0.0071		
345	345	4.9029	-0.0071		
346	346	4.9029	-0.0071		
347	347	4.9029	-0.0071		
348	348	4.9029	-0.0071		
349	349	4.9029	-0.0071		
350	350	4.9029	-0.0071		
351	351	4.9029	-0.0071		
352	352	4.9029	-0.0071		
353	353	4.9029	-0.0071		
354	354	4.9029	-0.0071		
355	355	4.9029	-0.0071		
356	356	4.9029	-0.0071		
357	357	4.9029	-0.0071		
358	358	4.9029	-0.0071		
359	359	4.9029	-0.0071		
360	360	4.9029	-0.0071		
361	361	4.9029	-0.0071		
362	362	4.9029	-0.0071		
363	363	4.9036	-0.0064		
364	364	4.9029	-0.0071		
365	365	4.9029	-0.0071		
366	366	4.9029	-0.0071		
367	367	4.9036	-0.0064		
368	368	4.9036	-0.0064		
369	369	4.9036	-0.0064		
370	370	4.9036	-0.0064		
371	371	4.9029	-0.0071		
372	372	4.9036	-0.0064		
373	373	4.9036	-0.0064		
374	374	4.9036	-0.0064		
375	375	4.9036	-0.0064		
376	376	4.9036	-0.0064		
377	377	4.9036	-0.0064		
378	378	4.9036	-0.0064		
379	379	4.9036	-0.0064		
380	380	4.9036	-0.0064		
381	381	4.9036	-0.0064		
382	382	4.9036	-0.0064		
383	383	4.9036	-0.0064		
384	384	4.9036	-0.0064		
385	385	4.9036	-0.0064		
386	386	4.9036	-0.0064		
387	387	4.9036	-0.0064		
388	388	4.9036	-0.0064		
389	389	4.9043	-0.0057		
390	390	4.9043	-0.0057		
391	391	4.9036	-0.0064		
392	392	4.9043	-0.0057		
393	393	4.9043	-0.0057		
394	394	4.9036	-0.0064		

				Slug Test - Water Level Data	Page 8 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
395	395	4.9036	-0.0064		
396	396	4.9036	-0.0064		
397	397	4.9036	-0.0064		
398	398	4.9043	-0.0057		
399	399	4.9043	-0.0057		
400	400	4.9043	-0.0057		
401	401	4.9043	-0.0057		
402	402	4.9015	-0.0085		
403	403	4.9029	-0.0071		
404	404	4.9036	-0.0064		
405	405	4.9036	-0.0064		
406	406	4.9043	-0.0057		
407	407	4.9043	-0.0057		
408	408	4.905	-0.005		
409	409	4.9043	-0.0057		
410	410	4.905	-0.005		

			Slug Test Analysis Report		
			Project: TRP		
			Number: 26200.000.001		
			Client: TRPL		
Location: Taylorville, Coal Creek		Slug Test: MB19d in		Test Well: MB19d	
Test Conducted by: J. Green / M. Knopick				Test Date: 16/10/2024	
Analysis Performed by: L. Nugent		MB19d slug in		Analysis Date: 11/11/2024	
Aquifer Thickness: 11.81 m					
Time [s] 15.790.8166241.1316.2391.4 1E0 1E-1 1E-2 1E-3 h/h0 					
Calculation using Bouwer & Rice					
Observation Well		Hydraulic Conductivity			
		[m/s]			
MB19d		9.76 × 10 ⁻⁶			

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB19d in			Test Well: MB19d		
Test Conducted by: J. Green / M. Knopick						Test Date: 16/10/2024		
Aquifer Thickness: 11.81 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB19d slug in	L. Nugent	11/11/2024	Bouwer & Rice	MB19d		9.76 × 10 ⁻⁶	

				Slug Test - Water Level Data		Page 1 of 7	
				Project: TRP			
				Number: 26200.000.001			
				Client: TRPL			
Location: Taylorville, Coal Creek			Slug Test: MB19d out			Test Well: MB19d	
Test Conducted by: J. Green / M. Knopick Test Date: 16/10/2024							
Water level at t=0 [m]: 5.42			Static Water Level [m]: 4.91			Water level change at t=0 [m]: 0.51	
	Time [s]	Water Level [m]	WL Change [m]				
1	1	5.4165	0.5065				
2	2	5.4122	0.5022				
3	3	5.4058	0.4958				
4	4	5.3952	0.4852				
5	5	5.3852	0.4752				
6	6	5.3731	0.4631				
7	7	5.3631	0.4531				
8	8	5.3511	0.4411				
9	9	5.3418	0.4318				
10	10	5.3304	0.4204				
11	11	5.3197	0.4097				
12	12	5.3084	0.3984				
13	13	5.2991	0.3891				
14	14	5.2906	0.3806				
15	15	5.282	0.372				
16	16	5.2742	0.3642				
17	17	5.2664	0.3564				
18	18	5.2593	0.3493				
19	19	5.2522	0.3422				
20	20	5.2451	0.3351				
21	21	5.2386	0.3286				
22	22	5.2322	0.3222				
23	23	5.2258	0.3158				
24	24	5.2194	0.3094				
25	25	5.2138	0.3038				
26	26	5.2081	0.2981				
27	27	5.2024	0.2924				
28	28	5.1938	0.2838				
29	29	5.1896	0.2796				
30	30	5.1853	0.2753				
31	31	5.1803	0.2703				
32	32	5.1753	0.2653				
33	33	5.1704	0.2604				
34	34	5.1661	0.2561				
35	35	5.1611	0.2511				
36	36	5.1568	0.2468				
37	37	5.1519	0.2419				
38	38	5.1476	0.2376				
39	39	5.1433	0.2333				
40	40	5.1391	0.2291				
41	41	5.1348	0.2248				
42	42	5.1312	0.2212				
43	43	5.127	0.217				
44	44	5.1234	0.2134				
45	45	5.1191	0.2091				
46	46	5.1156	0.2056				
47	47	5.112	0.202				
48	48	5.1085	0.1985				
49	49	5.1049	0.1949				
50	50	5.1013	0.1913				
51	51	5.0985	0.1885				
52	52	5.0949	0.1849				

				Slug Test - Water Level Data	Page 2 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
53	53	5.0921	0.1821		
54	54	5.0893	0.1793		
55	55	5.0857	0.1757		
56	56	5.0829	0.1729		
57	57	5.08	0.17		
58	58	5.0772	0.1672		
59	59	5.0743	0.1643		
60	60	5.0715	0.1615		
61	61	5.0686	0.1586		
62	62	5.0658	0.1558		
63	63	5.0636	0.1536		
64	64	5.0608	0.1508		
65	65	5.058	0.148		
66	66	5.0558	0.1458		
67	67	5.053	0.143		
68	68	5.0508	0.1408		
69	69	5.0487	0.1387		
70	70	5.0466	0.1366		
71	71	5.0444	0.1344		
72	72	5.0423	0.1323		
73	73	5.0402	0.1302		
74	74	5.038	0.128		
75	75	5.0359	0.1259		
76	76	5.0338	0.1238		
77	77	5.0316	0.1216		
78	78	5.0302	0.1202		
79	79	5.0281	0.1181		
80	80	5.0259	0.1159		
81	81	5.0245	0.1145		
82	82	5.0224	0.1124		
83	83	5.021	0.111		
84	84	5.0188	0.1088		
85	85	5.0174	0.1074		
86	86	5.016	0.106		
87	87	5.0138	0.1038		
88	88	5.0124	0.1024		
89	89	5.011	0.101		
90	90	5.0089	0.0989		
91	91	5.0074	0.0974		
92	92	5.0039	0.0939		
93	93	5.0032	0.0932		
94	94	5.0025	0.0925		
95	95	5.0018	0.0918		
96	96	5.0003	0.0903		
97	97	4.9989	0.0889		
98	98	4.9975	0.0875		
99	99	4.9961	0.0861		
100	100	4.9946	0.0846		
101	101	4.9939	0.0839		
102	102	4.9925	0.0825		
103	103	4.9918	0.0818		
104	104	4.9904	0.0804		
105	105	4.9889	0.0789		
106	106	4.9875	0.0775		
107	107	4.9868	0.0768		
108	108	4.9854	0.0754		
109	109	4.984	0.074		

				Slug Test - Water Level Data	Page 3 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
110	110	4.9833	0.0733		
111	111	4.9818	0.0718		
112	112	4.9811	0.0711		
113	113	4.9797	0.0697		
114	114	4.979	0.069		
115	115	4.9783	0.0683		
116	116	4.9769	0.0669		
117	117	4.9761	0.0661		
118	118	4.9754	0.0654		
119	119	4.9747	0.0647		
120	120	4.9733	0.0633		
121	121	4.9726	0.0626		
122	122	4.9719	0.0619		
123	123	4.9712	0.0612		
124	124	4.9705	0.0605		
125	125	4.969	0.059		
126	126	4.9683	0.0583		
127	127	4.9676	0.0576		
128	128	4.9662	0.0562		
129	129	4.9655	0.0555		
130	130	4.9648	0.0548		
131	131	4.964	0.054		
132	132	4.9633	0.0533		
133	133	4.9626	0.0526		
134	134	4.9619	0.0519		
135	135	4.9612	0.0512		
136	136	4.9605	0.0505		
137	137	4.9598	0.0498		
138	138	4.9591	0.0491		
139	139	4.9584	0.0484		
140	140	4.9576	0.0476		
141	141	4.9569	0.0469		
142	142	4.9562	0.0462		
143	143	4.9562	0.0462		
144	144	4.9555	0.0455		
145	145	4.9548	0.0448		
146	146	4.9541	0.0441		
147	147	4.9541	0.0441		
148	148	4.9534	0.0434		
149	149	4.9527	0.0427		
150	150	4.952	0.042		
151	151	4.9512	0.0412		
152	152	4.9505	0.0405		
153	153	4.9505	0.0405		
154	154	4.9498	0.0398		
155	155	4.9491	0.0391		
156	156	4.9463	0.0363		
157	157	4.947	0.037		
158	158	4.947	0.037		
159	159	4.9463	0.0363		
160	160	4.9463	0.0363		
161	161	4.9456	0.0356		
162	162	4.9456	0.0356		
163	163	4.9448	0.0348		
164	164	4.9448	0.0348		
165	165	4.9448	0.0348		
166	166	4.9441	0.0341		

				Slug Test - Water Level Data	Page 4 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
167	167	4.9434	0.0334		
168	168	4.9427	0.0327		
169	169	4.942	0.032		
170	170	4.942	0.032		
171	171	4.942	0.032		
172	172	4.9413	0.0313		
173	173	4.9406	0.0306		
174	174	4.9406	0.0306		
175	175	4.9399	0.0299		
176	176	4.9391	0.0291		
177	177	4.9391	0.0291		
178	178	4.9391	0.0291		
179	179	4.9384	0.0284		
180	180	4.9384	0.0284		
181	181	4.9377	0.0277		
182	182	4.9377	0.0277		
183	183	4.9377	0.0277		
184	184	4.937	0.027		
185	185	4.937	0.027		
186	186	4.9363	0.0263		
187	187	4.9363	0.0263		
188	188	4.9363	0.0263		
189	189	4.9356	0.0256		
190	190	4.9349	0.0249		
191	191	4.9349	0.0249		
192	192	4.9342	0.0242		
193	193	4.9342	0.0242		
194	194	4.9342	0.0242		
195	195	4.9335	0.0235		
196	196	4.9335	0.0235		
197	197	4.9327	0.0227		
198	198	4.9327	0.0227		
199	199	4.9327	0.0227		
200	200	4.932	0.022		
201	201	4.932	0.022		
202	202	4.932	0.022		
203	203	4.932	0.022		
204	204	4.9313	0.0213		
205	205	4.9306	0.0206		
206	206	4.9306	0.0206		
207	207	4.9306	0.0206		
208	208	4.9306	0.0206		
209	209	4.9299	0.0199		
210	210	4.9299	0.0199		
211	211	4.9292	0.0192		
212	212	4.9292	0.0192		
213	213	4.9292	0.0192		
214	214	4.9292	0.0192		
215	215	4.9285	0.0185		
216	216	4.9285	0.0185		
217	217	4.9285	0.0185		
218	218	4.9278	0.0178		
219	219	4.9278	0.0178		
220	220	4.9249	0.0149		
221	221	4.9263	0.0163		
222	222	4.9263	0.0163		
223	223	4.9263	0.0163		

				Slug Test - Water Level Data	Page 5 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
224	224	4.9271	0.0171		
225	225	4.9263	0.0163		
226	226	4.9263	0.0163		
227	227	4.9263	0.0163		
228	228	4.9263	0.0163		
229	229	4.9263	0.0163		
230	230	4.9263	0.0163		
231	231	4.9256	0.0156		
232	232	4.9256	0.0156		
233	233	4.9256	0.0156		
234	234	4.9249	0.0149		
235	235	4.9249	0.0149		
236	236	4.9249	0.0149		
237	237	4.9249	0.0149		
238	238	4.9242	0.0142		
239	239	4.9242	0.0142		
240	240	4.9242	0.0142		
241	241	4.9242	0.0142		
242	242	4.9242	0.0142		
243	243	4.9242	0.0142		
244	244	4.9235	0.0135		
245	245	4.9235	0.0135		
246	246	4.9228	0.0128		
247	247	4.9228	0.0128		
248	248	4.9228	0.0128		
249	249	4.9228	0.0128		
250	250	4.9228	0.0128		
251	251	4.9221	0.0121		
252	252	4.9221	0.0121		
253	253	4.9221	0.0121		
254	254	4.9221	0.0121		
255	255	4.9221	0.0121		
256	256	4.9221	0.0121		
257	257	4.9221	0.0121		
258	258	4.9214	0.0114		
259	259	4.9214	0.0114		
260	260	4.9214	0.0114		
261	261	4.9214	0.0114		
262	262	4.9214	0.0114		
263	263	4.9214	0.0114		
264	264	4.9207	0.0107		
265	265	4.9207	0.0107		
266	266	4.9207	0.0107		
267	267	4.9199	0.0099		
268	268	4.9207	0.0107		
269	269	4.9207	0.0107		
270	270	4.9199	0.0099		
271	271	4.9199	0.0099		
272	272	4.9199	0.0099		
273	273	4.9199	0.0099		
274	274	4.9199	0.0099		
275	275	4.9199	0.0099		
276	276	4.9199	0.0099		
277	277	4.9192	0.0092		
278	278	4.9192	0.0092		
279	279	4.9192	0.0092		
280	280	4.9192	0.0092		

				Slug Test - Water Level Data Project: TRP Number: 26200.000.001 Client: TRPL	Page 6 of 7
	Time [s]	Water Level [m]	WL Change [m]		
281	281	4.9185	0.0085		
282	282	4.9185	0.0085		
283	283	4.9185	0.0085		
284	284	4.9157	0.0057		
285	285	4.9171	0.0071		
286	286	4.9178	0.0078		
287	287	4.9178	0.0078		
288	288	4.9185	0.0085		
289	289	4.9185	0.0085		
290	290	4.9185	0.0085		
291	291	4.9185	0.0085		
292	292	4.9185	0.0085		
293	293	4.9185	0.0085		
294	294	4.9178	0.0078		
295	295	4.9178	0.0078		
296	296	4.9178	0.0078		
297	297	4.9178	0.0078		
298	298	4.9178	0.0078		
299	299	4.9178	0.0078		
300	300	4.9178	0.0078		
301	301	4.9171	0.0071		
302	302	4.9171	0.0071		
303	303	4.9171	0.0071		
304	304	4.9171	0.0071		
305	305	4.9171	0.0071		
306	306	4.9171	0.0071		
307	307	4.9171	0.0071		
308	308	4.9171	0.0071		
309	309	4.9171	0.0071		
310	310	4.9171	0.0071		
311	311	4.9164	0.0064		
312	312	4.9164	0.0064		
313	313	4.9164	0.0064		
314	314	4.9164	0.0064		
315	315	4.9171	0.0071		
316	316	4.9164	0.0064		
317	317	4.9164	0.0064		
318	318	4.9164	0.0064		
319	319	4.9164	0.0064		
320	320	4.9164	0.0064		
321	321	4.9164	0.0064		
322	322	4.9164	0.0064		
323	323	4.9164	0.0064		
324	324	4.9157	0.0057		
325	325	4.9164	0.0064		
326	326	4.9164	0.0064		
327	327	4.9164	0.0064		
328	328	4.9157	0.0057		
329	329	4.9157	0.0057		
330	330	4.9157	0.0057		
331	331	4.9157	0.0057		
332	332	4.9157	0.0057		
333	333	4.9157	0.0057		
334	334	4.9157	0.0057		
335	335	4.9157	0.0057		
336	336	4.9157	0.0057		
337	337	4.9157	0.0057		

				Slug Test - Water Level Data		Page 7 of 7
				Project: TRP		
				Number: 26200.000.001		
				Client: TRPL		
	Time [s]	Water Level [m]	WL Change [m]			
338	338	4.9157	0.0057			
339	339	4.9157	0.0057			
340	340	4.915	0.005			
341	341	4.915	0.005			
342	342	4.915	0.005			
343	343	4.915	0.005			
344	344	4.915	0.005			
345	345	4.915	0.005			
346	346	4.915	0.005			
347	347	4.915	0.005			
348	348	4.9121	0.0021			

			Slug Test Analysis Report	
			Project: TRP	
			Number: 26200.000.001	
			Client: TRPL	
Location: Taylorville, Coal Creek		Slug Test: MB19d out		Test Well: MB19d
Test Conducted by: J. Green / M. Knopick			Test Date: 16/10/2024	
Analysis Performed by: L. Nugent		MB19d slug out		Analysis Date: 11/11/2024
Aquifer Thickness: 11.81 m				
Time [s] -8.1 125 258.1 391.2 524.3 657.4 h/h0 1E0 1E-1 1E-2 1E-3				
Calculation using Bouwer & Rice				
Observation Well	Hydraulic Conductivity			
	[m/s]			
MB19d	1.11 × 10 ⁻⁵			

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB19d out			Test Well: MB19d		
Test Conducted by: J. Green / M. Knopick						Test Date: 16/10/2024		
Aquifer Thickness: 11.81 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB19d slug out	L. Nugent	11/11/2024	Bouwer & Rice	MB19d		1.11 × 10 ⁻⁵	

				Slug Test - Water Level Data		Page 1 of 4
				Project: TRP		
				Number: 26200.000.001		
				Client: TRPL		
Location: Taylorville, Coal Creek			Slug Test: MB13 in			Test Well: MB13
Test Conducted by: J. Green / M. Knopick Test Date: 16/10/2024						
Water level at t=0 [m]: 7.58			Static Water Level [m]: 7.93			Water level change at t=0 [m]: -0.35
	Time [s]	Water Level [m]	WL Change [m]			
1	1	7.5785	-0.3515			
2	2	7.6134	-0.3166			
3	3	7.6767	-0.2533			
4	4	7.7415	-0.1885			
5	5	7.7841	-0.1459			
6	6	7.809	-0.121			
7	7	7.8218	-0.1082			
8	8	7.8332	-0.0968			
9	9	7.8389	-0.0911			
10	10	7.8432	-0.0868			
11	11	7.846	-0.084			
12	12	7.8489	-0.0811			
13	13	7.8503	-0.0797			
14	14	7.8517	-0.0783			
15	15	7.8524	-0.0776			
16	16	7.8539	-0.0761			
17	17	7.8553	-0.0747			
18	18	7.8567	-0.0733			
19	19	7.8574	-0.0726			
20	20	7.8588	-0.0712			
21	21	7.8595	-0.0705			
22	22	7.861	-0.069			
23	23	7.861	-0.069			
24	24	7.8617	-0.0683			
25	25	7.8624	-0.0676			
26	26	7.8638	-0.0662			
27	27	7.8652	-0.0648			
28	28	7.8659	-0.0641			
29	29	7.8638	-0.0662			
30	30	7.8659	-0.0641			
31	31	7.8674	-0.0626			
32	32	7.8688	-0.0612			
33	33	7.8702	-0.0598			
34	34	7.8709	-0.0591			
35	35	7.8724	-0.0576			
36	36	7.8731	-0.0569			
37	37	7.8738	-0.0562			
38	38	7.8745	-0.0555			
39	39	7.8752	-0.0548			
40	40	7.8759	-0.0541			
41	41	7.8766	-0.0534			
42	42	7.8773	-0.0527			
43	43	7.8773	-0.0527			
44	44	7.8788	-0.0512			
45	45	7.8795	-0.0505			
46	46	7.8802	-0.0498			
47	47	7.8816	-0.0484			
48	48	7.8823	-0.0477			
49	49	7.883	-0.047			
50	50	7.8837	-0.0463			
51	51	7.8837	-0.0463			
52	52	7.8844	-0.0456			

				Slug Test - Water Level Data	Page 2 of 4
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
53	53	7.8852	-0.0448		
54	54	7.8859	-0.0441		
55	55	7.8866	-0.0434		
56	56	7.8873	-0.0427		
57	57	7.888	-0.042		
58	58	7.8887	-0.0413		
59	59	7.8894	-0.0406		
60	60	7.8894	-0.0406		
61	61	7.8908	-0.0392		
62	62	7.8908	-0.0392		
63	63	7.8916	-0.0384		
64	64	7.8923	-0.0377		
65	65	7.893	-0.037		
66	66	7.893	-0.037		
67	67	7.8937	-0.0363		
68	68	7.8944	-0.0356		
69	69	7.8951	-0.0349		
70	70	7.8958	-0.0342		
71	71	7.8958	-0.0342		
72	72	7.8965	-0.0335		
73	73	7.8972	-0.0328		
74	74	7.8972	-0.0328		
75	75	7.898	-0.032		
76	76	7.8987	-0.0313		
77	77	7.8987	-0.0313		
78	78	7.8994	-0.0306		
79	79	7.9001	-0.0299		
80	80	7.9001	-0.0299		
81	81	7.9008	-0.0292		
82	82	7.9008	-0.0292		
83	83	7.9015	-0.0285		
84	84	7.9022	-0.0278		
85	85	7.9022	-0.0278		
86	86	7.9029	-0.0271		
87	87	7.9037	-0.0263		
88	88	7.9037	-0.0263		
89	89	7.9044	-0.0256		
90	90	7.9044	-0.0256		
91	91	7.9051	-0.0249		
92	92	7.9051	-0.0249		
93	93	7.9029	-0.0271		
94	94	7.9044	-0.0256		
95	95	7.9051	-0.0249		
96	96	7.9058	-0.0242		
97	97	7.9065	-0.0235		
98	98	7.9072	-0.0228		
99	99	7.9079	-0.0221		
100	100	7.9079	-0.0221		
101	101	7.9086	-0.0214		
102	102	7.9086	-0.0214		
103	103	7.9093	-0.0207		
104	104	7.9093	-0.0207		
105	105	7.9093	-0.0207		
106	106	7.9101	-0.0199		
107	107	7.9108	-0.0192		
108	108	7.9108	-0.0192		
109	109	7.9108	-0.0192		

				Slug Test - Water Level Data	Page 3 of 4
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
110	110	7.9115	-0.0185		
111	111	7.9122	-0.0178		
112	112	7.9129	-0.0171		
113	113	7.9136	-0.0164		
114	114	7.9143	-0.0157		
115	115	7.9143	-0.0157		
116	116	7.915	-0.015		
117	117	7.915	-0.015		
118	118	7.9157	-0.0143		
119	119	7.9157	-0.0143		
120	120	7.9157	-0.0143		
121	121	7.9157	-0.0143		
122	122	7.9157	-0.0143		
123	123	7.9172	-0.0128		
124	124	7.9172	-0.0128		
125	125	7.9179	-0.0121		
126	126	7.9186	-0.0114		
127	127	7.9186	-0.0114		
128	128	7.9186	-0.0114		
129	129	7.9193	-0.0107		
130	130	7.9193	-0.0107		
131	131	7.9193	-0.0107		
132	132	7.9193	-0.0107		
133	133	7.92	-0.01		
134	134	7.92	-0.01		
135	135	7.92	-0.01		
136	136	7.9207	-0.0093		
137	137	7.9207	-0.0093		
138	138	7.9207	-0.0093		
139	139	7.9207	-0.0093		
140	140	7.9207	-0.0093		
141	141	7.9207	-0.0093		
142	142	7.9207	-0.0093		
143	143	7.9214	-0.0086		
144	144	7.9214	-0.0086		
145	145	7.9214	-0.0086		
146	146	7.9214	-0.0086		
147	147	7.9214	-0.0086		
148	148	7.9221	-0.0079		
149	149	7.9229	-0.0071		
150	150	7.9229	-0.0071		
151	151	7.9229	-0.0071		
152	152	7.9221	-0.0079		
153	153	7.9229	-0.0071		
154	154	7.9221	-0.0079		
155	155	7.9229	-0.0071		
156	156	7.9236	-0.0064		
157	157	7.9207	-0.0093		
158	158	7.9221	-0.0079		
159	159	7.9229	-0.0071		
160	160	7.9236	-0.0064		
161	161	7.9236	-0.0064		
162	162	7.9236	-0.0064		
163	163	7.9236	-0.0064		
164	164	7.9243	-0.0057		
165	165	7.9243	-0.0057		
166	166	7.9243	-0.0057		

				Slug Test - Water Level Data		Page 4 of 4
				Project: TRP		
				Number: 26200.000.001		
				Client: TRPL		
	Time [s]	Water Level [m]	WL Change [m]			
167	167	7.9243	-0.0057			
168	168	7.9243	-0.0057			
169	169	7.9236	-0.0064			
170	170	7.9243	-0.0057			
171	171	7.9243	-0.0057			
172	172	7.925	-0.005			

			Slug Test Analysis Report		
			Project: TRP		
			Number: 26200.000.001		
			Client: TRPL		
Location: Taylorville, Coal Creek		Slug Test: MB13 in		Test Well: MB13	
Test Conducted by: J. Green / M. Knopick				Test Date: 16/10/2024	
Analysis Performed by: L. Nugent		MB13 slug in		Analysis Date: 11/11/2024	
Aquifer Thickness: 10.31 m					
Time [s] 0 40 80 120 160 200 1E0 1E-1 1E-2 h/h0					
Calculation using Bouwer & Rice					
Observation Well		Hydraulic Conductivity			
		[m/s]			
MB13		1.88 × 10 ⁻⁶			

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB13 in			Test Well: MB13		
Test Conducted by: J. Green / M. Knopick						Test Date: 16/10/2024		
Aquifer Thickness: 10.31 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB13 slug in	L. Nugent	11/11/2024	Bouwer & Rice	MB13		1.88 × 10 ⁻⁶	

				Slug Test - Water Level Data		Page 1 of 8	
				Project: TRP			
				Number: 26200.000.001			
				Client: TRPL			
Location: Taylorville, Coal Creek			Slug Test: MB13 out			Test Well: MB13	
Test Conducted by: MK/ JG				Test Date: 16/10/2024			
Water level at t=0 [m]: 8.85			Static Water Level [m]: 7.93			Water level change at t=0 [m]: 0.92	
	Time [s]	Water Level [m]	WL Change [m]				
1	1	8.8491	0.9191				
2	2	8.551	0.621				
3	3	8.3355	0.4055				
4	4	8.2209	0.2909				
5	5	8.1519	0.2219				
6	6	8.1121	0.1821				
7	7	8.0886	0.1586				
8	8	8.073	0.143				
9	9	8.0623	0.1323				
10	10	8.0552	0.1252				
11	11	8.0502	0.1202				
12	12	8.0466	0.1166				
13	13	8.0438	0.1138				
14	14	8.0417	0.1117				
15	15	8.0424	0.1124				
16	16	8.0409	0.1109				
17	17	8.036	0.106				
18	18	8.031	0.101				
19	19	8.0281	0.0981				
20	20	8.0253	0.0953				
21	21	8.0239	0.0939				
22	22	8.0217	0.0917				
23	23	8.0203	0.0903				
24	24	8.0189	0.0889				
25	25	8.0175	0.0875				
26	26	8.016	0.086				
27	27	8.0146	0.0846				
28	28	8.0132	0.0832				
29	29	8.0118	0.0818				
30	30	8.0111	0.0811				
31	31	8.0096	0.0796				
32	32	8.0082	0.0782				
33	33	8.0075	0.0775				
34	34	8.0061	0.0761				
35	35	8.0047	0.0747				
36	36	8.004	0.074				
37	37	8.0032	0.0732				
38	38	8.0018	0.0718				
39	39	8.0011	0.0711				
40	40	7.9997	0.0697				
41	41	7.999	0.069				
42	42	7.9975	0.0675				
43	43	7.994	0.064				
44	44	7.9947	0.0647				
45	45	7.994	0.064				
46	46	7.9933	0.0633				
47	47	7.9926	0.0626				
48	48	7.9919	0.0619				
49	49	7.9911	0.0611				
50	50	7.9904	0.0604				
51	51	7.9897	0.0597				
52	52	7.9883	0.0583				

				Slug Test - Water Level Data	Page 2 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
53	53	7.9876	0.0576		
54	54	7.9869	0.0569		
55	55	7.9862	0.0562		
56	56	7.9855	0.0555		
57	57	7.9847	0.0547		
58	58	7.984	0.054		
59	59	7.9826	0.0526		
60	60	7.9819	0.0519		
61	61	7.9812	0.0512		
62	62	7.9805	0.0505		
63	63	7.9798	0.0498		
64	64	7.9791	0.0491		
65	65	7.9783	0.0483		
66	66	7.9783	0.0483		
67	67	7.9776	0.0476		
68	68	7.9769	0.0469		
69	69	7.9762	0.0462		
70	70	7.9755	0.0455		
71	71	7.9748	0.0448		
72	72	7.9741	0.0441		
73	73	7.9741	0.0441		
74	74	7.9734	0.0434		
75	75	7.9726	0.0426		
76	76	7.9719	0.0419		
77	77	7.9712	0.0412		
78	78	7.9705	0.0405		
79	79	7.9705	0.0405		
80	80	7.9698	0.0398		
81	81	7.9691	0.0391		
82	82	7.9684	0.0384		
83	83	7.9677	0.0377		
84	84	7.9677	0.0377		
85	85	7.967	0.037		
86	86	7.967	0.037		
87	87	7.9662	0.0362		
88	88	7.9655	0.0355		
89	89	7.9648	0.0348		
90	90	7.9648	0.0348		
91	91	7.9641	0.0341		
92	92	7.9634	0.0334		
93	93	7.9634	0.0334		
94	94	7.9627	0.0327		
95	95	7.9627	0.0327		
96	96	7.962	0.032		
97	97	7.9613	0.0313		
98	98	7.9613	0.0313		
99	99	7.9606	0.0306		
100	100	7.9606	0.0306		
101	101	7.9598	0.0298		
102	102	7.9598	0.0298		
103	103	7.9591	0.0291		
104	104	7.9591	0.0291		
105	105	7.9591	0.0291		
106	106	7.9584	0.0284		
107	107	7.9556	0.0256		
108	108	7.9563	0.0263		
109	109	7.9563	0.0263		

				Slug Test - Water Level Data	Page 3 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
110	110	7.9563	0.0263		
111	111	7.9563	0.0263		
112	112	7.9556	0.0256		
113	113	7.9556	0.0256		
114	114	7.9556	0.0256		
115	115	7.9549	0.0249		
116	116	7.9549	0.0249		
117	117	7.9549	0.0249		
118	118	7.9549	0.0249		
119	119	7.9542	0.0242		
120	120	7.9542	0.0242		
121	121	7.9542	0.0242		
122	122	7.9534	0.0234		
123	123	7.9534	0.0234		
124	124	7.9527	0.0227		
125	125	7.9527	0.0227		
126	126	7.9527	0.0227		
127	127	7.952	0.022		
128	128	7.9513	0.0213		
129	129	7.9513	0.0213		
130	130	7.9513	0.0213		
131	131	7.9506	0.0206		
132	132	7.9506	0.0206		
133	133	7.9506	0.0206		
134	134	7.9499	0.0199		
135	135	7.9499	0.0199		
136	136	7.9499	0.0199		
137	137	7.9492	0.0192		
138	138	7.9492	0.0192		
139	139	7.9492	0.0192		
140	140	7.9492	0.0192		
141	141	7.9492	0.0192		
142	142	7.9485	0.0185		
143	143	7.9485	0.0185		
144	144	7.9485	0.0185		
145	145	7.9485	0.0185		
146	146	7.9477	0.0177		
147	147	7.9477	0.0177		
148	148	7.9477	0.0177		
149	149	7.9477	0.0177		
150	150	7.947	0.017		
151	151	7.947	0.017		
152	152	7.947	0.017		
153	153	7.947	0.017		
154	154	7.9463	0.0163		
155	155	7.9463	0.0163		
156	156	7.9463	0.0163		
157	157	7.9463	0.0163		
158	158	7.9463	0.0163		
159	159	7.9463	0.0163		
160	160	7.9456	0.0156		
161	161	7.9456	0.0156		
162	162	7.9456	0.0156		
163	163	7.9456	0.0156		
164	164	7.9456	0.0156		
165	165	7.9456	0.0156		
166	166	7.9456	0.0156		

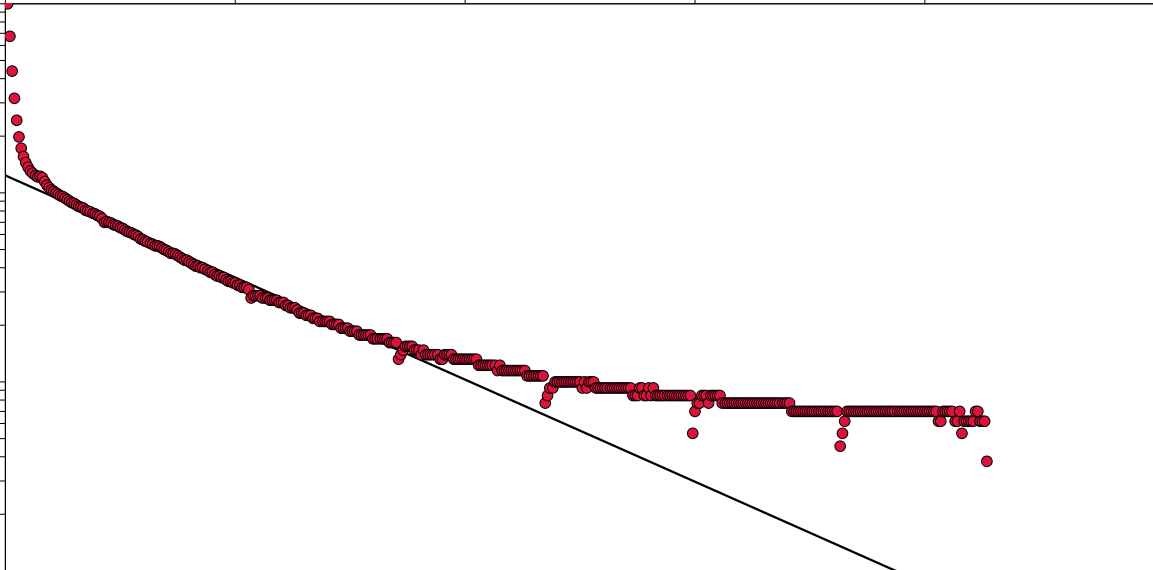
				Slug Test - Water Level Data	Page 4 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
167	167	7.9449	0.0149		
168	168	7.9449	0.0149		
169	169	7.9449	0.0149		
170	170	7.9449	0.0149		
171	171	7.9421	0.0121		
172	172	7.9428	0.0128		
173	173	7.9435	0.0135		
174	174	7.9442	0.0142		
175	175	7.9442	0.0142		
176	176	7.9442	0.0142		
177	177	7.9442	0.0142		
178	178	7.9435	0.0135		
179	179	7.9435	0.0135		
180	180	7.9435	0.0135		
181	181	7.9428	0.0128		
182	182	7.9435	0.0135		
183	183	7.9428	0.0128		
184	184	7.9428	0.0128		
185	185	7.9428	0.0128		
186	186	7.9428	0.0128		
187	187	7.9428	0.0128		
188	188	7.9428	0.0128		
189	189	7.9421	0.0121		
190	190	7.9421	0.0121		
191	191	7.9428	0.0128		
192	192	7.9428	0.0128		
193	193	7.9428	0.0128		
194	194	7.9428	0.0128		
195	195	7.9421	0.0121		
196	196	7.9421	0.0121		
197	197	7.9421	0.0121		
198	198	7.9421	0.0121		
199	199	7.9421	0.0121		
200	200	7.9421	0.0121		
201	201	7.9421	0.0121		
202	202	7.9421	0.0121		
203	203	7.9421	0.0121		
204	204	7.9421	0.0121		
205	205	7.9421	0.0121		
206	206	7.9413	0.0113		
207	207	7.9413	0.0113		
208	208	7.9413	0.0113		
209	209	7.9413	0.0113		
210	210	7.9413	0.0113		
211	211	7.9413	0.0113		
212	212	7.9413	0.0113		
213	213	7.9413	0.0113		
214	214	7.9406	0.0106		
215	215	7.9413	0.0113		
216	216	7.9406	0.0106		
217	217	7.9406	0.0106		
218	218	7.9406	0.0106		
219	219	7.9406	0.0106		
220	220	7.9406	0.0106		
221	221	7.9406	0.0106		
222	222	7.9406	0.0106		
223	223	7.9406	0.0106		

				Slug Test - Water Level Data Project: TRP Number: 26200.000.001 Client: TRPL Page 5 of 8
	Time [s]	Water Level [m]	WL Change [m]	
224	224	7.9406	0.0106	
225	225	7.9406	0.0106	
226	226	7.9406	0.0106	
227	227	7.9399	0.0099	
228	228	7.9399	0.0099	
229	229	7.9399	0.0099	
230	230	7.9399	0.0099	
231	231	7.9399	0.0099	
232	232	7.9399	0.0099	
233	233	7.9399	0.0099	
234	234	7.9399	0.0099	
235	235	7.9371	0.0071	
236	236	7.9378	0.0078	
237	237	7.9385	0.0085	
238	238	7.9385	0.0085	
239	239	7.9392	0.0092	
240	240	7.9392	0.0092	
241	241	7.9392	0.0092	
242	242	7.9392	0.0092	
243	243	7.9392	0.0092	
244	244	7.9392	0.0092	
245	245	7.9392	0.0092	
246	246	7.9392	0.0092	
247	247	7.9392	0.0092	
248	248	7.9392	0.0092	
249	249	7.9392	0.0092	
250	250	7.9392	0.0092	
251	251	7.9385	0.0085	
252	252	7.9392	0.0092	
253	253	7.9385	0.0085	
254	254	7.9392	0.0092	
255	255	7.9392	0.0092	
256	256	7.9392	0.0092	
257	257	7.9385	0.0085	
258	258	7.9385	0.0085	
259	259	7.9385	0.0085	
260	260	7.9385	0.0085	
261	261	7.9385	0.0085	
262	262	7.9385	0.0085	
263	263	7.9385	0.0085	
264	264	7.9385	0.0085	
265	265	7.9385	0.0085	
266	266	7.9385	0.0085	
267	267	7.9385	0.0085	
268	268	7.9385	0.0085	
269	269	7.9385	0.0085	
270	270	7.9385	0.0085	
271	271	7.9385	0.0085	
272	272	7.9385	0.0085	
273	273	7.9378	0.0078	
274	274	7.9378	0.0078	
275	275	7.9378	0.0078	
276	276	7.9385	0.0085	
277	277	7.9385	0.0085	
278	278	7.9378	0.0078	
279	279	7.9378	0.0078	
280	280	7.9385	0.0085	

				Slug Test - Water Level Data	Page 6 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
281	281	7.9378	0.0078		
282	282	7.9385	0.0085		
283	283	7.9378	0.0078		
284	284	7.9378	0.0078		
285	285	7.9378	0.0078		
286	286	7.9378	0.0078		
287	287	7.9378	0.0078		
288	288	7.9378	0.0078		
289	289	7.9378	0.0078		
290	290	7.9378	0.0078		
291	291	7.9378	0.0078		
292	292	7.9378	0.0078		
293	293	7.9378	0.0078		
294	294	7.9378	0.0078		
295	295	7.9378	0.0078		
296	296	7.9378	0.0078		
297	297	7.9378	0.0078		
298	298	7.9378	0.0078		
299	299	7.9349	0.0049		
300	300	7.9364	0.0064		
301	301	7.9371	0.0071		
302	302	7.9371	0.0071		
303	303	7.9378	0.0078		
304	304	7.9378	0.0078		
305	305	7.9378	0.0078		
306	306	7.9371	0.0071		
307	307	7.9378	0.0078		
308	308	7.9378	0.0078		
309	309	7.9378	0.0078		
310	310	7.9378	0.0078		
311	311	7.9378	0.0078		
312	312	7.9371	0.0071		
313	313	7.9371	0.0071		
314	314	7.9371	0.0071		
315	315	7.9371	0.0071		
316	316	7.9371	0.0071		
317	317	7.9371	0.0071		
318	318	7.9371	0.0071		
319	319	7.9371	0.0071		
320	320	7.9371	0.0071		
321	321	7.9371	0.0071		
322	322	7.9371	0.0071		
323	323	7.9371	0.0071		
324	324	7.9371	0.0071		
325	325	7.9371	0.0071		
326	326	7.9371	0.0071		
327	327	7.9371	0.0071		
328	328	7.9371	0.0071		
329	329	7.9371	0.0071		
330	330	7.9371	0.0071		
331	331	7.9371	0.0071		
332	332	7.9371	0.0071		
333	333	7.9371	0.0071		
334	334	7.9371	0.0071		
335	335	7.9371	0.0071		
336	336	7.9371	0.0071		
337	337	7.9371	0.0071		

				Slug Test - Water Level Data	Page 7 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
338	338	7.9371	0.0071		
339	339	7.9371	0.0071		
340	340	7.9371	0.0071		
341	341	7.9371	0.0071		
342	342	7.9364	0.0064		
343	343	7.9364	0.0064		
344	344	7.9364	0.0064		
345	345	7.9364	0.0064		
346	346	7.9364	0.0064		
347	347	7.9364	0.0064		
348	348	7.9364	0.0064		
349	349	7.9364	0.0064		
350	350	7.9364	0.0064		
351	351	7.9364	0.0064		
352	352	7.9364	0.0064		
353	353	7.9364	0.0064		
354	354	7.9364	0.0064		
355	355	7.9364	0.0064		
356	356	7.9364	0.0064		
357	357	7.9364	0.0064		
358	358	7.9364	0.0064		
359	359	7.9364	0.0064		
360	360	7.9364	0.0064		
361	361	7.9364	0.0064		
362	362	7.9364	0.0064		
363	363	7.9342	0.0042		
364	364	7.9349	0.0049		
365	365	7.9357	0.0057		
366	366	7.9364	0.0064		
367	367	7.9364	0.0064		
368	368	7.9364	0.0064		
369	369	7.9364	0.0064		
370	370	7.9364	0.0064		
371	371	7.9364	0.0064		
372	372	7.9364	0.0064		
373	373	7.9364	0.0064		
374	374	7.9364	0.0064		
375	375	7.9364	0.0064		
376	376	7.9364	0.0064		
377	377	7.9364	0.0064		
378	378	7.9364	0.0064		
379	379	7.9364	0.0064		
380	380	7.9364	0.0064		
381	381	7.9364	0.0064		
382	382	7.9364	0.0064		
383	383	7.9364	0.0064		
384	384	7.9364	0.0064		
385	385	7.9364	0.0064		
386	386	7.9364	0.0064		
387	387	7.9364	0.0064		
388	388	7.9364	0.0064		
389	389	7.9364	0.0064		
390	390	7.9364	0.0064		
391	391	7.9364	0.0064		
392	392	7.9364	0.0064		
393	393	7.9364	0.0064		
394	394	7.9364	0.0064		

				Slug Test - Water Level Data	Page 8 of 8
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
395	395	7.9364	0.0064		
396	396	7.9364	0.0064		
397	397	7.9364	0.0064		
398	398	7.9364	0.0064		
399	399	7.9364	0.0064		
400	400	7.9364	0.0064		
401	401	7.9364	0.0064		
402	402	7.9364	0.0064		
403	403	7.9364	0.0064		
404	404	7.9364	0.0064		
405	405	7.9364	0.0064		
406	406	7.9357	0.0057		
407	407	7.9357	0.0057		
408	408	7.9364	0.0064		
409	409	7.9364	0.0064		
410	410	7.9364	0.0064		
411	411	7.9364	0.0064		
412	412	7.9364	0.0064		
413	413	7.9357	0.0057		
414	414	7.9357	0.0057		
415	415	7.9364	0.0064		
416	416	7.9349	0.0049		
417	417	7.9357	0.0057		
418	418	7.9357	0.0057		
419	419	7.9357	0.0057		
420	420	7.9357	0.0057		
421	421	7.9357	0.0057		
422	422	7.9364	0.0064		
423	423	7.9364	0.0064		
424	424	7.9357	0.0057		
425	425	7.9357	0.0057		
426	426	7.9357	0.0057		
427	427	7.9335	0.0035		

			Slug Test Analysis Report		
			Project: TRP		
			Number: 26200.000.001		
			Client: TRPL		
Location: Taylorville, Coal Creek		Slug Test: MB13 out		Test Well: MB13	
Test Conducted by: MK/ JG				Test Date: 16/10/2024	
Analysis Performed by: L. Nugent		MB13 slug out		Analysis Date: 11/11/2024	
Aquifer Thickness: 10.81 m					
Time [s] 0 100 200 300 400 500 1E0 1E-1 1E-2 1E-3 h/h0 					
Calculation using Bouwer & Rice					
Observation Well		Hydraulic Conductivity			
		[m/s]			
MB13		1.51 × 10 ⁻⁶			

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB13 out			Test Well: MB13		
Test Conducted by: MK/ JG						Test Date: 16/10/2024		
Aquifer Thickness: 10.81 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB13 slug out	L. Nugent	11/11/2024	Bouwer & Rice	MB13		1.51 × 10 ⁻⁶	

				Slug Test - Water Level Data		Page 1 of 5	
				Project: TRP			
				Number: 26200.000.001			
				Client: TRPL			
Location: Taylorville, Coal Creek			Slug Test: MB14 in			Test Well: MB14	
Test Conducted by: J. Green			Test Date: 16/01/2025				
Water level at t=0 [m]: 7.86			Static Water Level [m]: 8.70			Water level change at t=0 [m]: -0.84	
	Time [s]	Water Level [m]	WL Change [m]				
1	1	7.855	-0.845				
2	2	7.867	-0.833				
3	3	7.907	-0.793				
4	4	7.958	-0.742				
5	5	8.012	-0.688				
6	6	8.061	-0.639				
7	7	8.106	-0.594				
8	8	8.148	-0.552				
9	9	8.183	-0.517				
10	10	8.211	-0.489				
11	11	8.234	-0.466				
12	12	8.253	-0.447				
13	13	8.267	-0.433				
14	14	8.278	-0.422				
15	15	8.287	-0.413				
16	16	8.295	-0.405				
17	17	8.302	-0.398				
18	18	8.309	-0.391				
19	19	8.316	-0.384				
20	20	8.322	-0.378				
21	21	8.328	-0.372				
22	22	8.334	-0.366				
23	23	8.34	-0.36				
24	24	8.345	-0.355				
25	25	8.351	-0.349				
26	26	8.356	-0.344				
27	27	8.36	-0.34				
28	28	8.365	-0.335				
29	29	8.369	-0.331				
30	30	8.37	-0.33				
31	31	8.376	-0.324				
32	32	8.381	-0.319				
33	33	8.386	-0.314				
34	34	8.39	-0.31				
35	35	8.394	-0.306				
36	36	8.399	-0.301				
37	37	8.404	-0.296				
38	38	8.408	-0.292				
39	39	8.413	-0.287				
40	40	8.417	-0.283				
41	41	8.421	-0.279				
42	42	8.425	-0.275				
43	43	8.429	-0.271				
44	44	8.434	-0.266				
45	45	8.438	-0.262				
46	46	8.442	-0.258				
47	47	8.447	-0.253				
48	48	8.451	-0.249				
49	49	8.456	-0.244				
50	50	8.46	-0.24				
51	51	8.465	-0.235				
52	52	8.469	-0.231				

				Slug Test - Water Level Data	Page 2 of 5
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
53	53	8.473	-0.227		
54	54	8.477	-0.223		
55	55	8.48	-0.22		
56	56	8.484	-0.216		
57	57	8.487	-0.213		
58	58	8.491	-0.209		
59	59	8.494	-0.206		
60	60	8.497	-0.203		
61	61	8.50	-0.20		
62	62	8.504	-0.196		
63	63	8.508	-0.192		
64	64	8.511	-0.189		
65	65	8.515	-0.185		
66	66	8.518	-0.182		
67	67	8.52	-0.18		
68	68	8.523	-0.177		
69	69	8.526	-0.174		
70	70	8.528	-0.172		
71	71	8.531	-0.169		
72	72	8.534	-0.166		
73	73	8.537	-0.163		
74	74	8.54	-0.16		
75	75	8.543	-0.157		
76	76	8.545	-0.155		
77	77	8.548	-0.152		
78	78	8.551	-0.149		
79	79	8.554	-0.146		
80	80	8.555	-0.145		
81	81	8.559	-0.141		
82	82	8.561	-0.139		
83	83	8.564	-0.136		
84	84	8.566	-0.134		
85	85	8.568	-0.132		
86	86	8.57	-0.13		
87	87	8.572	-0.128		
88	88	8.574	-0.126		
89	89	8.575	-0.125		
90	90	8.577	-0.123		
91	91	8.579	-0.121		
92	92	8.58	-0.12		
93	93	8.582	-0.118		
94	94	8.582	-0.118		
95	95	8.585	-0.115		
96	96	8.587	-0.113		
97	97	8.589	-0.111		
98	98	8.591	-0.109		
99	99	8.593	-0.107		
100	100	8.595	-0.105		
101	101	8.596	-0.104		
102	102	8.598	-0.102		
103	103	8.60	-0.10		
104	104	8.602	-0.098		
105	105	8.603	-0.097		
106	106	8.605	-0.095		
107	107	8.607	-0.093		
108	108	8.609	-0.091		
109	109	8.611	-0.089		

				Slug Test - Water Level Data	Page 3 of 5
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
110	110	8.613	-0.087		
111	111	8.615	-0.085		
112	112	8.616	-0.084		
113	113	8.618	-0.082		
114	114	8.619	-0.081		
115	115	8.621	-0.079		
116	116	8.622	-0.078		
117	117	8.623	-0.077		
118	118	8.625	-0.075		
119	119	8.626	-0.074		
120	120	8.627	-0.073		
121	121	8.628	-0.072		
122	122	8.63	-0.07		
123	123	8.631	-0.069		
124	124	8.632	-0.068		
125	125	8.633	-0.067		
126	126	8.634	-0.066		
127	127	8.635	-0.065		
128	128	8.636	-0.064		
129	129	8.637	-0.063		
130	130	8.638	-0.062		
131	131	8.639	-0.061		
132	132	8.64	-0.06		
133	133	8.641	-0.059		
134	134	8.641	-0.059		
135	135	8.642	-0.058		
136	136	8.643	-0.057		
137	137	8.643	-0.057		
138	138	8.644	-0.056		
139	139	8.644	-0.056		
140	140	8.645	-0.055		
141	141	8.645	-0.055		
142	142	8.646	-0.054		
143	143	8.647	-0.053		
144	144	8.647	-0.053		
145	145	8.648	-0.052		
146	146	8.649	-0.051		
147	147	8.65	-0.05		
148	148	8.65	-0.05		
149	149	8.651	-0.049		
150	150	8.651	-0.049		
151	151	8.652	-0.048		
152	152	8.653	-0.047		
153	153	8.653	-0.047		
154	154	8.654	-0.046		
155	155	8.654	-0.046		
156	156	8.655	-0.045		
157	157	8.656	-0.044		
158	158	8.653	-0.047		
159	159	8.655	-0.045		
160	160	8.655	-0.045		
161	161	8.656	-0.044		
162	162	8.657	-0.043		
163	163	8.658	-0.042		
164	164	8.658	-0.042		
165	165	8.659	-0.041		
166	166	8.659	-0.041		

				Slug Test - Water Level Data	Page 4 of 5
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
167	167	8.66	-0.04		
168	168	8.66	-0.04		
169	169	8.66	-0.04		
170	170	8.661	-0.039		
171	171	8.661	-0.039		
172	172	8.661	-0.039		
173	173	8.662	-0.038		
174	174	8.662	-0.038		
175	175	8.663	-0.037		
176	176	8.663	-0.037		
177	177	8.663	-0.037		
178	178	8.663	-0.037		
179	179	8.664	-0.036		
180	180	8.664	-0.036		
181	181	8.665	-0.035		
182	182	8.665	-0.035		
183	183	8.666	-0.034		
184	184	8.666	-0.034		
185	185	8.666	-0.034		
186	186	8.667	-0.033		
187	187	8.667	-0.033		
188	188	8.667	-0.033		
189	189	8.668	-0.032		
190	190	8.668	-0.032		
191	191	8.669	-0.031		
192	192	8.669	-0.031		
193	193	8.669	-0.031		
194	194	8.67	-0.03		
195	195	8.67	-0.03		
196	196	8.67	-0.03		
197	197	8.671	-0.029		
198	198	8.671	-0.029		
199	199	8.671	-0.029		
200	200	8.671	-0.029		
201	201	8.671	-0.029		
202	202	8.672	-0.028		
203	203	8.672	-0.028		
204	204	8.673	-0.027		
205	205	8.673	-0.027		
206	206	8.673	-0.027		
207	207	8.674	-0.026		
208	208	8.674	-0.026		
209	209	8.674	-0.026		
210	210	8.674	-0.026		
211	211	8.675	-0.025		
212	212	8.675	-0.025		
213	213	8.675	-0.025		
214	214	8.675	-0.025		
215	215	8.676	-0.024		
216	216	8.676	-0.024		
217	217	8.676	-0.024		
218	218	8.676	-0.024		
219	219	8.677	-0.023		
220	220	8.677	-0.023		
221	221	8.677	-0.023		
222	222	8.675	-0.025		
223	223	8.677	-0.023		

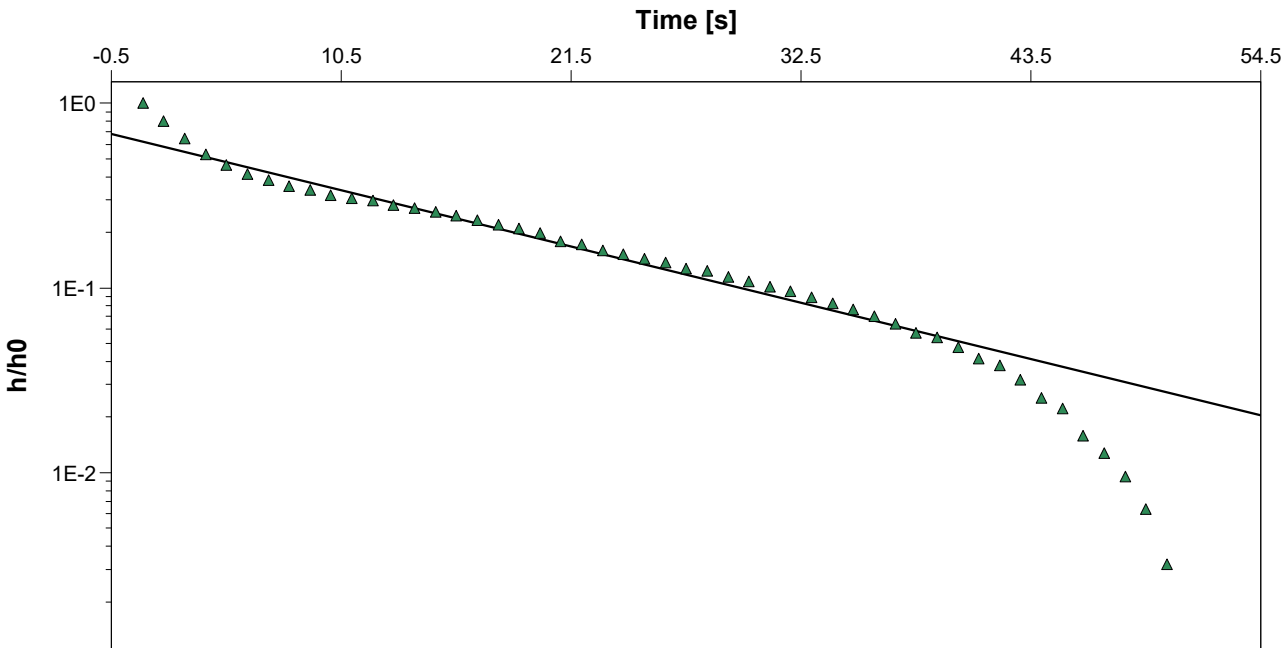
Client: TRPL

	Time [s]	Water Level [m]	WL Change [m]
224	224	8.678	-0.022
225	225	8.678	-0.022
226	226	8.679	-0.021

			Slug Test Analysis Report		
			Project: TRP		
			Number: 26200.000.001		
			Client: TRPL		
Location: Taylorville, Coal Creek		Slug Test: MB14 in		Test Well: MB14	
Test Conducted by: J. Green				Test Date: 16/01/2025	
Analysis Performed by: L Nugent		MB14 slug in		Analysis Date: 21/01/2025	
Aquifer Thickness: 4.92 m					
Time [s] </					

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB14 in			Test Well: MB14		
Test Conducted by: J. Green						Test Date: 16/01/2025		
Aquifer Thickness: 4.92 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB14 slug in	L Nugent	21/01/2025	Bouwer & Rice	MB14		2.33 × 10 ⁻⁶	

				Slug Test - Water Level Data		Page 1 of 1	
				Project: TRP			
				Number: 26200.000.001			
				Client: TRPL			
Location: Taylorville, Coal Creek			Slug Test: MB14 out			Test Well: MB14	
Test Conducted by:			Test Date: 16/01/2025				
Water level at t=0 [m]: 9.01			Static Water Level [m]: 8.70			Water level change at t=0 [m]: 0.31	
	Time [s]	Water Level [m]	WL Change [m]				
1	1	9.014	0.314				
2	2	8.952	0.252				
3	3	8.903	0.203				
4	4	8.866	0.166				
5	5	8.846	0.146				
6	6	8.83	0.13				
7	7	8.821	0.121				
8	8	8.812	0.112				
9	9	8.807	0.107				
10	10	8.80	0.10				
11	11	8.796	0.096				
12	12	8.793	0.093				
13	13	8.788	0.088				
14	14	8.785	0.085				
15	15	8.781	0.081				
16	16	8.777	0.077				
17	17	8.773	0.073				
18	18	8.769	0.069				
19	19	8.766	0.066				
20	20	8.762	0.062				
21	21	8.756	0.056				
22	22	8.754	0.054				
23	23	8.75	0.05				
24	24	8.748	0.048				
25	25	8.745	0.045				
26	26	8.743	0.043				
27	27	8.74	0.04				
28	28	8.739	0.039				
29	29	8.736	0.036				
30	30	8.734	0.034				
31	31	8.732	0.032				
32	32	8.73	0.03				
33	33	8.728	0.028				
34	34	8.726	0.026				
35	35	8.724	0.024				
36	36	8.722	0.022				
37	37	8.72	0.02				
38	38	8.718	0.018				
39	39	8.717	0.017				
40	40	8.715	0.015				
41	41	8.713	0.013				
42	42	8.712	0.012				
43	43	8.71	0.01				
44	44	8.708	0.008				
45	45	8.707	0.007				
46	46	8.705	0.005				
47	47	8.704	0.004				
48	48	8.703	0.003				
49	49	8.702	0.002				
50	50	8.701	0.001				
51	51	8.70	0.00				

			Slug Test Analysis Report		
			Project: TRP		
			Number: 26200.000.001		
			Client: TRPL		
Location: Taylorville, Coal Creek		Slug Test: MB14 out		Test Well: MB14	
Test Conducted by:				Test Date: 16/01/2025	
Analysis Performed by: L. Nugent		MB14 slug out		Analysis Date: 21/01/2025	
Aquifer Thickness: 4.92 m					
Time [s] 					
Calculation using Bouwer & Rice					
Observation Well		Hydraulic Conductivity [m/s]			
MB14		9.93 × 10 ⁻⁶			

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB14 out			Test Well: MB14		
Test Conducted by:						Test Date: 16/01/2025		
Aquifer Thickness: 4.92 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB14 slug out	L. Nugent	21/01/2025	Bouwer & Rice	MB14		9.93 × 10 ⁻⁶	

				Slug Test - Water Level Data		Page 1 of 7	
				Project: TRP			
				Number: 26200.000.001			
				Client: TRPL			
Location: Taylorville, Coal Creek			Slug Test: MB14 in.2			Test Well: MB14	
Test Conducted by: J. Green			Test Date: 16/01/2025				
Water level at t=0 [m]: 8.11			Static Water Level [m]: 8.70			Water level change at t=0 [m]: -0.59	
	Time [s]	Water Level [m]	WL Change [m]				
1	1	8.106	-0.594				
2	2	8.171	-0.529				
3	3	8.251	-0.449				
4	4	8.297	-0.403				
5	5	8.319	-0.381				
6	6	8.343	-0.357				
7	7	8.356	-0.344				
8	8	8.365	-0.335				
9	9	8.373	-0.327				
10	10	8.38	-0.32				
11	11	8.387	-0.313				
12	12	8.391	-0.309				
13	13	8.398	-0.302				
14	14	8.403	-0.297				
15	15	8.407	-0.293				
16	16	8.412	-0.288				
17	17	8.417	-0.283				
18	18	8.421	-0.279				
19	19	8.425	-0.275				
20	20	8.429	-0.271				
21	21	8.433	-0.267				
22	22	8.437	-0.263				
23	23	8.442	-0.258				
24	24	8.445	-0.255				
25	25	8.449	-0.251				
26	26	8.453	-0.247				
27	27	8.457	-0.243				
28	28	8.46	-0.24				
29	29	8.463	-0.237				
30	30	8.466	-0.234				
31	31	8.468	-0.232				
32	32	8.471	-0.229				
33	33	8.474	-0.226				
34	34	8.477	-0.223				
35	35	8.479	-0.221				
36	36	8.482	-0.218				
37	37	8.484	-0.216				
38	38	8.466	-0.234				
39	39	8.482	-0.218				
40	40	8.487	-0.213				
41	41	8.491	-0.209				
42	42	8.495	-0.205				
43	43	8.497	-0.203				
44	44	8.50	-0.20				
45	45	8.502	-0.198				
46	46	8.505	-0.195				
47	47	8.507	-0.193				
48	48	8.509	-0.191				
49	49	8.511	-0.189				
50	50	8.513	-0.187				
51	51	8.515	-0.185				
52	52	8.517	-0.183				

				Slug Test - Water Level Data	Page 2 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
53	53	8.519	-0.181		
54	54	8.521	-0.179		
55	55	8.523	-0.177		
56	56	8.525	-0.175		
57	57	8.528	-0.172		
58	58	8.53	-0.17		
59	59	8.531	-0.169		
60	60	8.533	-0.167		
61	61	8.535	-0.165		
62	62	8.537	-0.163		
63	63	8.538	-0.162		
64	64	8.54	-0.16		
65	65	8.539	-0.161		
66	66	8.542	-0.158		
67	67	8.544	-0.156		
68	68	8.546	-0.154		
69	69	8.547	-0.153		
70	70	8.549	-0.151		
71	71	8.551	-0.149		
72	72	8.552	-0.148		
73	73	8.554	-0.146		
74	74	8.555	-0.145		
75	75	8.556	-0.144		
76	76	8.557	-0.143		
77	77	8.558	-0.142		
78	78	8.559	-0.141		
79	79	8.561	-0.139		
80	80	8.562	-0.138		
81	81	8.563	-0.137		
82	82	8.564	-0.136		
83	83	8.565	-0.135		
84	84	8.567	-0.133		
85	85	8.568	-0.132		
86	86	8.569	-0.131		
87	87	8.57	-0.13		
88	88	8.572	-0.128		
89	89	8.573	-0.127		
90	90	8.574	-0.126		
91	91	8.575	-0.125		
92	92	8.577	-0.123		
93	93	8.578	-0.122		
94	94	8.579	-0.121		
95	95	8.577	-0.123		
96	96	8.577	-0.123		
97	97	8.579	-0.121		
98	98	8.583	-0.117		
99	99	8.585	-0.115		
100	100	8.587	-0.113		
101	101	8.588	-0.112		
102	102	8.589	-0.111		
103	103	8.591	-0.109		
104	104	8.592	-0.108		
105	105	8.594	-0.106		
106	106	8.595	-0.105		
107	107	8.596	-0.104		
108	108	8.597	-0.103		
109	109	8.598	-0.102		

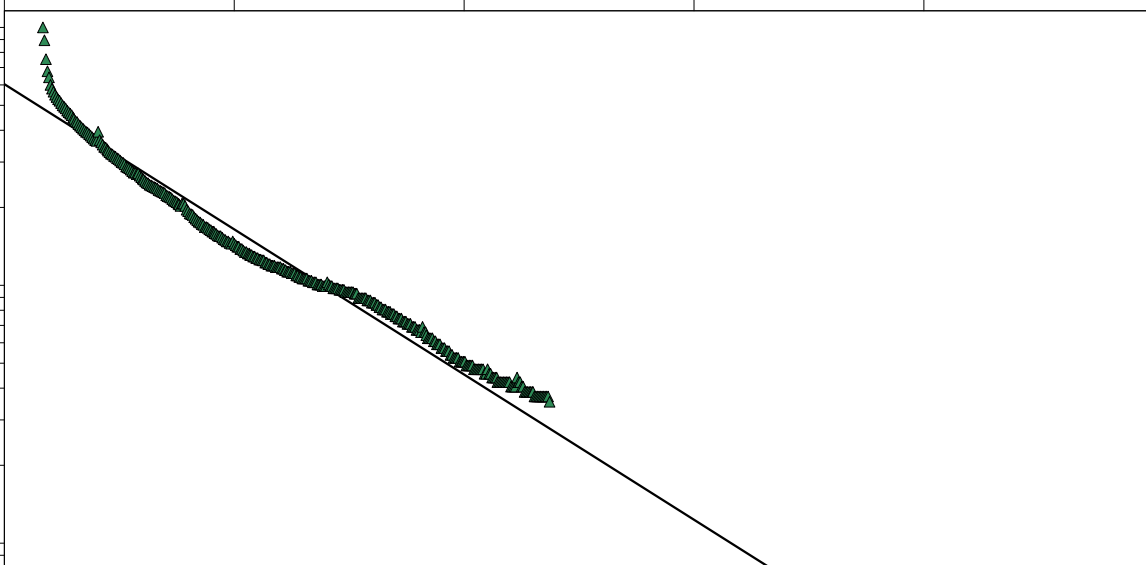
				Slug Test - Water Level Data	Page 3 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
110	110	8.60	-0.10		
111	111	8.60	-0.10		
112	112	8.601	-0.099		
113	113	8.602	-0.098		
114	114	8.603	-0.097		
115	115	8.604	-0.096		
116	116	8.604	-0.096		
117	117	8.605	-0.095		
118	118	8.606	-0.094		
119	119	8.607	-0.093		
120	120	8.608	-0.092		
121	121	8.608	-0.092		
122	122	8.609	-0.091		
123	123	8.61	-0.09		
124	124	8.611	-0.089		
125	125	8.612	-0.088		
126	126	8.612	-0.088		
127	127	8.613	-0.087		
128	128	8.614	-0.086		
129	129	8.612	-0.088		
130	130	8.614	-0.086		
131	131	8.615	-0.085		
132	132	8.616	-0.084		
133	133	8.616	-0.084		
134	134	8.618	-0.082		
135	135	8.618	-0.082		
136	136	8.619	-0.081		
137	137	8.62	-0.08		
138	138	8.62	-0.08		
139	139	8.621	-0.079		
140	140	8.621	-0.079		
141	141	8.622	-0.078		
142	142	8.622	-0.078		
143	143	8.623	-0.077		
144	144	8.623	-0.077		
145	145	8.624	-0.076		
146	146	8.624	-0.076		
147	147	8.625	-0.075		
148	148	8.625	-0.075		
149	149	8.626	-0.074		
150	150	8.626	-0.074		
151	151	8.627	-0.073		
152	152	8.627	-0.073		
153	153	8.628	-0.072		
154	154	8.628	-0.072		
155	155	8.629	-0.071		
156	156	8.629	-0.071		
157	157	8.629	-0.071		
158	158	8.63	-0.07		
159	159	8.63	-0.07		
160	160	8.63	-0.07		
161	161	8.63	-0.07		
162	162	8.631	-0.069		
163	163	8.631	-0.069		
164	164	8.632	-0.068		
165	165	8.632	-0.068		
166	166	8.633	-0.067		

				Slug Test - Water Level Data	Page 4 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
167	167	8.633	-0.067		
168	168	8.633	-0.067		
169	169	8.634	-0.066		
170	170	8.634	-0.066		
171	171	8.634	-0.066		
172	172	8.635	-0.065		
173	173	8.635	-0.065		
174	174	8.636	-0.064		
175	175	8.636	-0.064		
176	176	8.637	-0.063		
177	177	8.637	-0.063		
178	178	8.637	-0.063		
179	179	8.637	-0.063		
180	180	8.638	-0.062		
181	181	8.638	-0.062		
182	182	8.638	-0.062		
183	183	8.639	-0.061		
184	184	8.639	-0.061		
185	185	8.639	-0.061		
186	186	8.64	-0.06		
187	187	8.64	-0.06		
188	188	8.64	-0.06		
189	189	8.64	-0.06		
190	190	8.641	-0.059		
191	191	8.641	-0.059		
192	192	8.641	-0.059		
193	193	8.639	-0.061		
194	194	8.64	-0.06		
195	195	8.641	-0.059		
196	196	8.641	-0.059		
197	197	8.642	-0.058		
198	198	8.642	-0.058		
199	199	8.642	-0.058		
200	200	8.642	-0.058		
201	201	8.643	-0.057		
202	202	8.643	-0.057		
203	203	8.643	-0.057		
204	204	8.643	-0.057		
205	205	8.644	-0.056		
206	206	8.644	-0.056		
207	207	8.644	-0.056		
208	208	8.644	-0.056		
209	209	8.644	-0.056		
210	210	8.644	-0.056		
211	211	8.645	-0.055		
212	212	8.645	-0.055		
213	213	8.645	-0.055		
214	214	8.647	-0.053		
215	215	8.647	-0.053		
216	216	8.647	-0.053		
217	217	8.647	-0.053		
218	218	8.647	-0.053		
219	219	8.647	-0.053		
220	220	8.648	-0.052		
221	221	8.648	-0.052		
222	222	8.648	-0.052		
223	223	8.649	-0.051		

				Slug Test - Water Level Data	Page 5 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
224	224	8.649	-0.051		
225	225	8.649	-0.051		
226	226	8.65	-0.05		
227	227	8.65	-0.05		
228	228	8.651	-0.049		
229	229	8.651	-0.049		
230	230	8.652	-0.048		
231	231	8.652	-0.048		
232	232	8.652	-0.048		
233	233	8.653	-0.047		
234	234	8.653	-0.047		
235	235	8.653	-0.047		
236	236	8.654	-0.046		
237	237	8.654	-0.046		
238	238	8.654	-0.046		
239	239	8.655	-0.045		
240	240	8.655	-0.045		
241	241	8.656	-0.044		
242	242	8.656	-0.044		
243	243	8.656	-0.044		
244	244	8.657	-0.043		
245	245	8.657	-0.043		
246	246	8.657	-0.043		
247	247	8.658	-0.042		
248	248	8.658	-0.042		
249	249	8.658	-0.042		
250	250	8.659	-0.041		
251	251	8.659	-0.041		
252	252	8.659	-0.041		
253	253	8.66	-0.04		
254	254	8.66	-0.04		
255	255	8.66	-0.04		
256	256	8.661	-0.039		
257	257	8.659	-0.041		
258	258	8.661	-0.039		
259	259	8.661	-0.039		
260	260	8.662	-0.038		
261	261	8.663	-0.037		
262	262	8.663	-0.037		
263	263	8.663	-0.037		
264	264	8.663	-0.037		
265	265	8.664	-0.036		
266	266	8.664	-0.036		
267	267	8.665	-0.035		
268	268	8.665	-0.035		
269	269	8.665	-0.035		
270	270	8.666	-0.034		
271	271	8.666	-0.034		
272	272	8.666	-0.034		
273	273	8.667	-0.033		
274	274	8.667	-0.033		
275	275	8.667	-0.033		
276	276	8.668	-0.032		
277	277	8.668	-0.032		
278	278	8.669	-0.031		
279	279	8.669	-0.031		
280	280	8.669	-0.031		

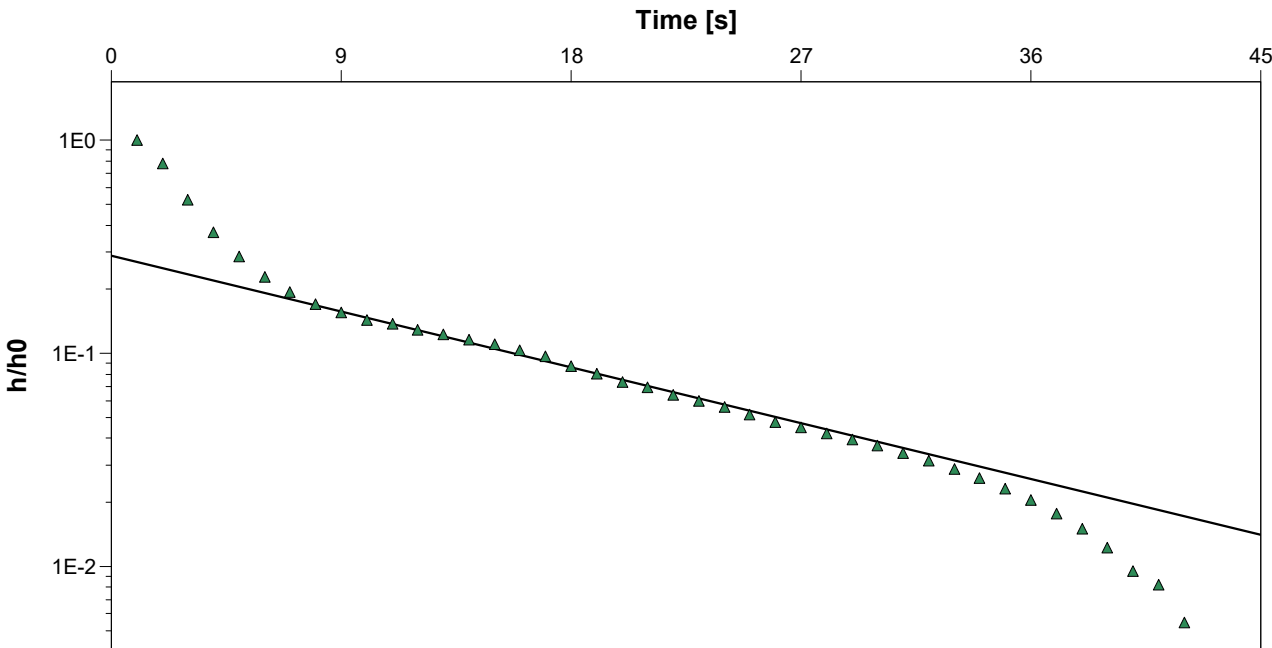
				Slug Test - Water Level Data	Page 6 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
281	281	8.669	-0.031		
282	282	8.67	-0.03		
283	283	8.67	-0.03		
284	284	8.67	-0.03		
285	285	8.67	-0.03		
286	286	8.67	-0.03		
287	287	8.671	-0.029		
288	288	8.671	-0.029		
289	289	8.671	-0.029		
290	290	8.671	-0.029		
291	291	8.671	-0.029		
292	292	8.672	-0.028		
293	293	8.672	-0.028		
294	294	8.672	-0.028		
295	295	8.672	-0.028		
296	296	8.672	-0.028		
297	297	8.672	-0.028		
298	298	8.672	-0.028		
299	299	8.673	-0.027		
300	300	8.673	-0.027		
301	301	8.672	-0.028		
302	302	8.673	-0.027		
303	303	8.673	-0.027		
304	304	8.674	-0.026		
305	305	8.674	-0.026		
306	306	8.674	-0.026		
307	307	8.674	-0.026		
308	308	8.675	-0.025		
309	309	8.675	-0.025		
310	310	8.675	-0.025		
311	311	8.675	-0.025		
312	312	8.675	-0.025		
313	313	8.675	-0.025		
314	314	8.675	-0.025		
315	315	8.675	-0.025		
316	316	8.675	-0.025		
317	317	8.676	-0.024		
318	318	8.676	-0.024		
319	319	8.676	-0.024		
320	320	8.676	-0.024		
321	321	8.674	-0.026		
322	322	8.675	-0.025		
323	323	8.675	-0.025		
324	324	8.676	-0.024		
325	325	8.676	-0.024		
326	326	8.677	-0.023		
327	327	8.677	-0.023		
328	328	8.677	-0.023		
329	329	8.677	-0.023		
330	330	8.677	-0.023		
331	331	8.677	-0.023		
332	332	8.677	-0.023		
333	333	8.678	-0.022		
334	334	8.678	-0.022		
335	335	8.678	-0.022		
336	336	8.678	-0.022		
337	337	8.678	-0.022		

				Slug Test - Water Level Data	Page 7 of 7
				Project: TRP	
				Number: 26200.000.001	
				Client: TRPL	
	Time [s]	Water Level [m]	WL Change [m]		
338	338	8.678	-0.022		
339	339	8.678	-0.022		
340	340	8.678	-0.022		
341	341	8.678	-0.022		
342	342	8.678	-0.022		
343	343	8.679	-0.021		

			Slug Test Analysis Report	
			Project: TRP	
			Number: 26200.000.001	
			Client: TRPL	
Location: Taylorville, Coal Creek		Slug Test: MB14 in.2		Test Well: MB14
Test Conducted by: J. Green			Test Date: 16/01/2025	
Analysis Performed by: L. Nugent		MB14 slug in (2)		Analysis Date: 21/01/2025
Aquifer Thickness: 6.80 m				
Time [s] -25.2 130 285.3 440.5 595.8 751 h/h0 1E0 1E-1 1E-2 				
Calculation using Bouwer & Rice				
Observation Well		Hydraulic Conductivity		
		[m/s]		
MB14		1.30 × 10 ⁻⁶		

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB14 in.2			Test Well: MB14		
Test Conducted by: J. Green						Test Date: 16/01/2025		
Aquifer Thickness: 6.80 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB14 slug in (2)	L. Nugent	21/01/2025	Bouwer & Rice	MB14		1.30 × 10 ⁻⁶	

				Slug Test - Water Level Data		Page 1 of 1	
				Project: TRP			
				Number: 26200.000.001			
				Client: TRPL			
Location: Taylorville, Coal Creek			Slug Test: MB14 out.2			Test Well: MB14	
Test Conducted by: J. Green			Test Date: 16/01/2025				
Water level at t=0 [m]: 9.43			Static Water Level [m]: 8.70			Water level change at t=0 [m]: 0.73	
	Time [s]	Water Level [m]	WL Change [m]				
1	1	9.433	0.733				
2	2	9.271	0.571				
3	3	9.085	0.385				
4	4	8.972	0.272				
5	5	8.909	0.209				
6	6	8.867	0.167				
7	7	8.842	0.142				
8	8	8.825	0.125				
9	9	8.814	0.114				
10	10	8.805	0.105				
11	11	8.801	0.101				
12	12	8.795	0.095				
13	13	8.79	0.09				
14	14	8.785	0.085				
15	15	8.781	0.081				
16	16	8.776	0.076				
17	17	8.771	0.071				
18	18	8.764	0.064				
19	19	8.759	0.059				
20	20	8.754	0.054				
21	21	8.751	0.051				
22	22	8.747	0.047				
23	23	8.744	0.044				
24	24	8.741	0.041				
25	25	8.738	0.038				
26	26	8.735	0.035				
27	27	8.733	0.033				
28	28	8.731	0.031				
29	29	8.729	0.029				
30	30	8.727	0.027				
31	31	8.725	0.025				
32	32	8.723	0.023				
33	33	8.721	0.021				
34	34	8.719	0.019				
35	35	8.717	0.017				
36	36	8.715	0.015				
37	37	8.713	0.013				
38	38	8.711	0.011				
39	39	8.709	0.009				
40	40	8.707	0.007				
41	41	8.706	0.006				
42	42	8.704	0.004				
43	43	8.702	0.002				
44	44	8.701	0.001				

			Slug Test Analysis Report		
			Project: TRP		
			Number: 26200.000.001		
			Client: TRPL		
Location: Taylorville, Coal Creek		Slug Test: MB14 out.2		Test Well: MB14	
Test Conducted by: J. Green				Test Date: 16/01/2025	
Analysis Performed by: L. Nugent		MB14 slug out (2)		Analysis Date: 21/01/2025	
Aquifer Thickness: 6.80 m					
Time [s] 					
Calculation using Bouwer & Rice					
Observation Well		Hydraulic Conductivity [m/s]			
MB14		1.04 × 10 ⁻⁵			

				Slug Test - Analyses Report				
				Project: TRP				
				Number: 26200.000.001				
				Client: TRPL				
Location: Taylorville, Coal Creek			Slug Test: MB14 out.2			Test Well: MB14		
Test Conducted by: J. Green						Test Date: 16/01/2025		
Aquifer Thickness: 6.80 m								
	Analysis Name	Analysis Performed	Analysis Date	Method name	Well	T [m²/s]	K [m/s]	S
1	MB14 slug out (2)	L. Nugent	21/01/2025	Bouwer & Rice	MB14		1.04 × 10 ⁻⁵	

APPENDIX 2

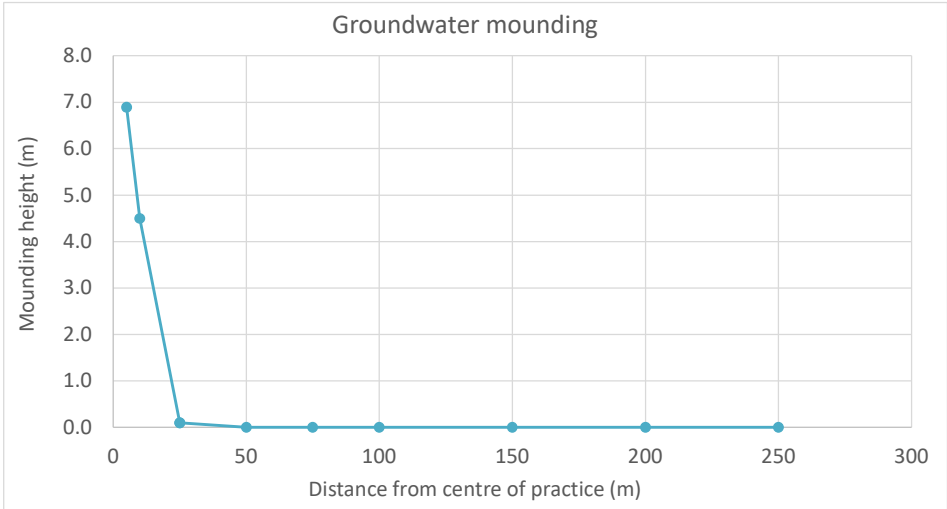
Updated Hantush Model

This spreadsheet will calculate the height of a groundwater mound beneath an stormwater infiltration BMP.
More information can be found in the U.S. Geological Survey Scientific Investigations Report 2010-5102
"Simulation of groundwater mounding beneath hypothetical stormwater infiltration basins".

[Go to the USGS report](#)

Input Values		
Recharge (infiltration) rate (m/day)	2.000	R
Specific yield, S_y (dimensionless)	0.190	S_y
Horizontal hydraulic conductivity, K_h (m/day)	0.860000	K_h
1/2 length of basin (x direction, in m)	10.0	x
1/2 width of basin (y direction, in m)	10.0	y
Duration of infiltration period (days)	1	t
Initial thickness of saturated zone (m)	8.0	$h_i(0)$

CALCULATE



Distance from center of infiltration BMP (m)	Ground-water mounding (m)
5	6.89
10	4.50
25	0.10
25	0.10
50	0.00
75	0.00
100	0.00
150	0.00
200	0.00
250	0.00

IF YOU CHANGE THE SCALE OF THE
DISTANCE MEASUREMENTS
(OR ANY OTHER PARAMETER),
YOU MUST RECALCULATE!!

APPENDIX 3

Updated RISC5 Model

Summary of Input Values Used in Fate and Transport Model

Model Description:

Source media: Groundwater (dissolved phase concentration)

Saturated zone model (dissolved phase source)

Aquifer Properties		
Effective porosity	cm3/cm3	3.0E-01
Fraction organic carbon	g oc/g soil	1.0E-03
Hydraulic conductivity	m/d	8.6E-01
Soil bulk density	g/cm3	1.6E+00
Hydraulic gradient	m/m	1.8E-02

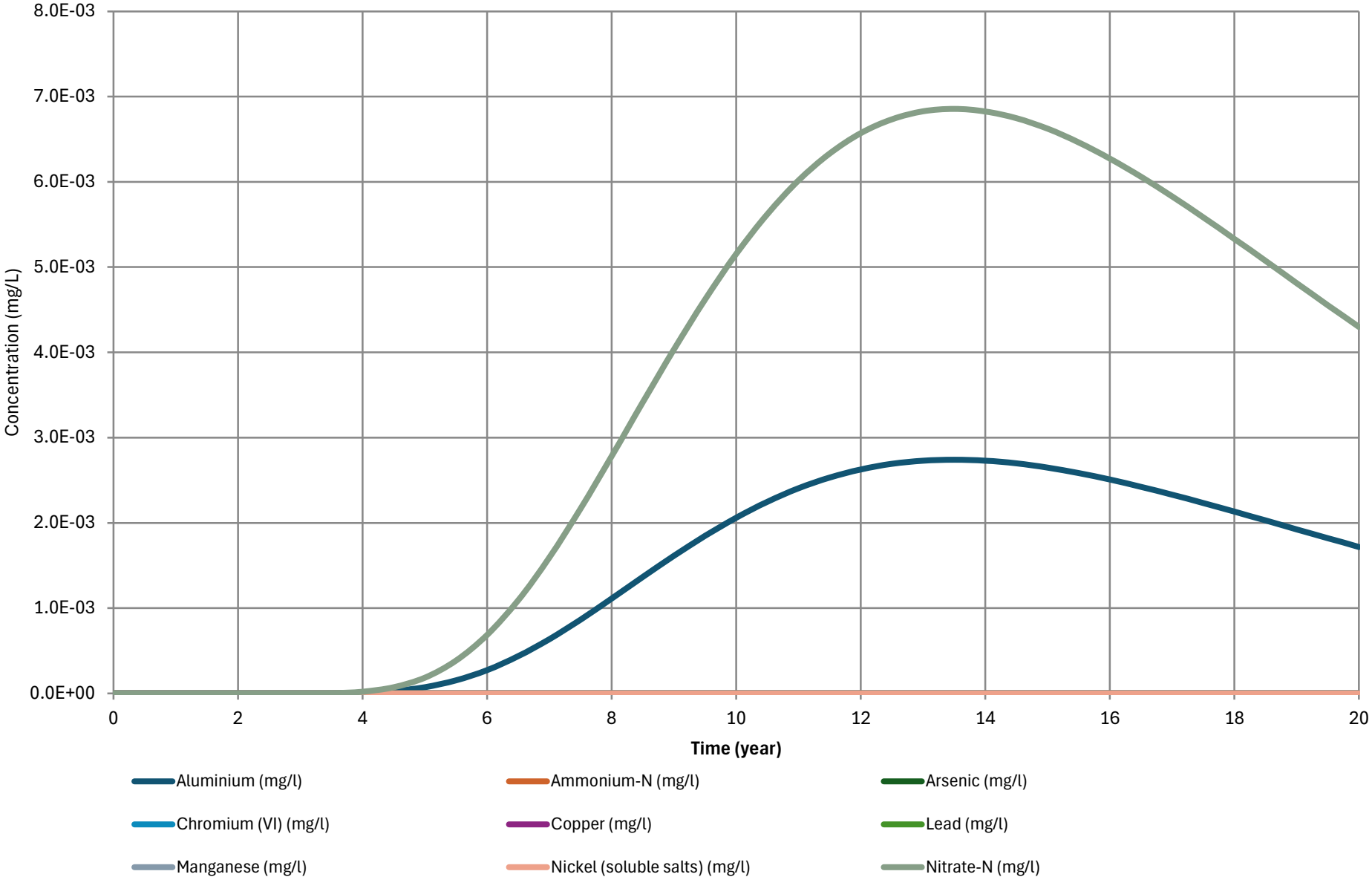
Groundwater Source Geometry		
***Pulse Source. Length of pulse:	yr	5.0E-01
Total thickness of source	m	8.0E+00
Length of source	m	1.0E+01
Width of source	m	1.0E+01

Receptor Well Location		
Distance downgradient	m	3.1E+02
Distance cross-gradient	m	0.0E+00
Depth to top of well screen	m	0.0E+00
Depth to bottom of well screen	m	5.0E+00
Number of vertical points used to calculate conc.	-	5.0E+00
Longitudinal dispersivity code calculated. See output file.		
Transverse dispersivity code calculated. See output file.		
Vertical dispersivity code calculated. See output file.		

Source for Groundwater Model [mg/l]		
Total Aluminium	mg/l	2.6E+00
Ammonium-N	mg/l	9.8E+01
Total Arsenic	mg/l	9.8E-01
Total Chromium	mg/l	2.2E-01
Total Copper	mg/l	1.9E-01
Total Lead	mg/l	1.3E-02
Total Manganese	mg/l	6.7E+00
Total Nickel	mg/l	1.6E-01
Nitrate-N	mg/l	6.5E+00

[illegible]

Soakage Area to 310 m Downgradient Well (0.86 m/d)



[illegible]

Summary of Input Values Used in Fate and Transport Model

Model Description:

Source media: Groundwater (dissolved phase concentration)

Saturated zone model (dissolved phase source)

Aquifer Properties		
Effective porosity	cm3/cm3	3.0E-01
Fraction organic carbon	g oc/g soil	1.0E-03
Hydraulic conductivity	m/d	1.1E-01
Soil bulk density	g/cm3	1.6E+00
Hydraulic gradient	m/m	1.8E-02

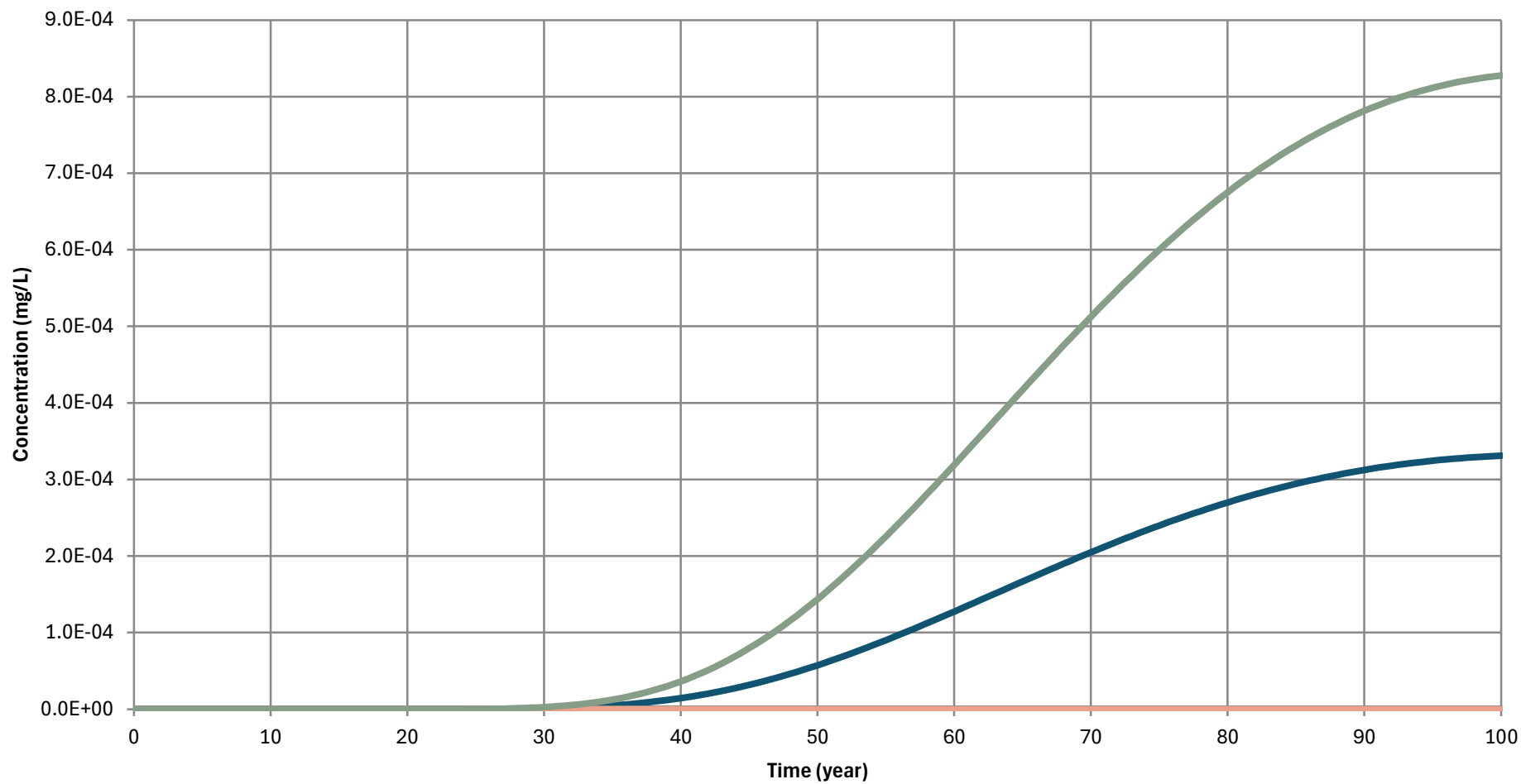
Groundwater Source Geometry		
***Pulse Source. Length of pulse:	yr	5.0E-01
Total thickness of source	m	8.0E+00
Length of source	m	1.0E+01
Width of source	m	1.0E+01

Receptor Well Location		
Distance downgradient	m	3.1E+02
Distance cross-gradient	m	0.0E+00
Depth to top of well screen	m	0.0E+00
Depth to bottom of well screen	m	5.0E+00
Number of vertical points used to calculate conc.	-	5.0E+00
Longitudinal dispersivity code calculated. See output file.		
Transverse dispersivity code calculated. See output file.		
Vertical dispersivity code calculated. See output file.		

Dissolved Source for Groundwater Model [mg/l]		
Total Aluminium	mg/l	2.6E+00
Ammonium-N	mg/l	9.8E+01
Total Arsenic	mg/l	9.8E-01
Total Chromium	mg/l	2.2E-01
Total Copper	mg/l	1.9E-01
Total Lead	mg/l	1.3E-02
Total Manganese	mg/l	6.7E+00
Total Nickel	mg/l	1.6E-01
Nitrate-N	mg/l	6.5E+00

[illegible]

Soakage Area to 310 m Downgradient Well (0.11 m/d)



Aluminium (mg/l)	Ammonium-N (mg/l)	Arsenic (mg/l)
Chromium (VI) (mg/l)	Copper (mg/l)	Lead (mg/l)
Manganese (mg/l)	Nickel (soluble salts) (mg/l)	Nitrate-N (mg/l)

Time	Aluminium	Ammonium-N	Arsenic	Chromium (VI)	Copper	Lead	Manganese	Nickel (soluble salts)	Nitrate-N
(year)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	(mg/l)	(mg/l)
0	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
1	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
2	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
3	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
4	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
5	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
6	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
7	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
8	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
9	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
10	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
11	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
12	2.8E-14	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	7.0E-14
13	3.4E-13	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	8.4E-13
14	2.4E-12	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	6.0E-12
15	1.3E-11	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	3.3E-11
16	5.7E-11	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.4E-10
17	2.1E-10	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	5.2E-10
18	6.5E-10	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.6E-09
19	1.8E-09	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	4.5E-09
20	4.4E-09	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.1E-08
21	1.0E-08	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	2.5E-08
22	2.1E-08	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	5.2E-08
23	4.1E-08	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.0E-07
24	7.4E-08	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.9E-07
25	1.3E-07	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	3.2E-07
26	2.2E-07	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	5.4E-07
27	3.4E-07	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	8.6E-07
28	5.3E-07	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.3E-06
29	7.9E-07	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	2.0E-06
30	1.1E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	2.8E-06
31	1.6E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	4.0E-06
32	2.2E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	5.5E-06
33	3.0E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	7.4E-06
34	3.9E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	9.7E-06
35	5.0E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.3E-05
36	6.4E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.6E-05
37	8.0E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	2.0E-05
38	9.9E-06	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	2.5E-05
39	1.2E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	3.0E-05
40	1.5E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	3.6E-05
41	1.7E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	4.3E-05
42	2.0E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	5.1E-05
43	2.4E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	6.0E-05
44	2.8E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	6.9E-05
45	3.2E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	8.0E-05
46	3.6E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	9.1E-05
47	4.1E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.0E-04
48	4.6E-05	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	1.2E-04

[illegible]

APPENDIX 4

Ecological Survey

APPENDIX 5

Water Quality Comparison

Table A5.1: Excerpt of Sample Results

	GRE001	Grey River	Upgradient / Cross-Gradient of Operations						Downgradient of Operations				Upgradient	
	PWT01	GR-01	MB2s		MB2d		MB3		MB4d		MB14	MB18	MB19s	MB19d
	16/01/25	16/01/25	30/07/24	14/10/24	30/07/24	14/10/24	30/07/24	14/10/24	30/07/24	16/10/24	14/10/24	10/11/24	15/10/24	15/10/24
Sum of Anions	0.63	0.67	0.84	0.84	3.9	3.7	0.74	0.77	1.8	1.12	11	17.3	2.9	9.7
Sum of Cations	0.63	0.65	0.89	0.89	3.9	3.9	0.8	0.8	1.91	1.16	11.8	18.7	3.2	10.5
An/Cat balance (%)	0%	2%	-3%	-3%	0%	-3%	-4%	-2%	-3%	-2%	-4%	-4%	-5%	-4%
pH	6.4	7.2	6.1	5.6	6.8	6.7	6	5.7	6.5	6.2	7.3	6.5	6.2	7.1
Total Alkalinity	20	24	19.2	18.7	176	169	16.2	16	43	24	350	330	54	290
Carbonate	< 1.0	< 1.0	-	< 1.0	-	< 1.0	-	< 1.0	-	< 1.0	< 1.0	< 1.0	< 1.0	< 1.0
Bicarbonate	24	29	23	23	210	210	19.8	19.6	53	29	430	400	66	360
Total Hardness	21	22	32	32	50	50	28	28	76	43	470	690	113	330
Electrical conductivity (mS/m)	6.6	6.6	9.8	9.9	35.8	34.9	8.9	9.1	19.9	12.7	100.7	157.9	30.5	89.7
Dissolved calcium	6.7	7.2	11.1	11.3	13.3	13.2	9.4	9.6	28	15.8	173	240	34	102
Dissolved magnesium	0.98	1.08	1.1	1.04	4.1	4.1	1.04	1.03	1.64	1	9.4	23	6.6	18.5
Dissolved potassium	0.84	0.77	0.89	0.86	1.59	1.65	0.69	0.69	0.79	0.94	16.4	22	2.2	8.3
Dissolved sodium	4.2	4.3	4.9	4.6	64	64	4.9	4.8	8.4	5.9	27	62	18.6	71
Chloride	4.4	4.5	7.8	7.8	11.1	11.8	7	7.8	7	6.7	24	33	10	15.2
Total-N	0.2	0.15		1.2		0.71		1.54		1.53	6.3	21	1.97	10.7
Total Ammoniacal-N	< 0.010	< 0.010	< 0.010	< 0.010	0.25	0.27	< 0.010	< 0.010	0.039	< 0.010	4.1	18.7	0.5	3.0
Nitrite-N	< 0.002	< 0.002	< 0.002	< 0.002	< 0.002	0.003	< 0.002	< 0.002	0.034	< 0.002	< 0.02	< 0.002	0.009	< 0.002
Nitrate-N	0.17	0.091	1.23	1.12	< 0.002	0.007	1.42	1.43	1.92	1.41	< 0.02	< 0.002	0.36	< 0.002
Nitrite+Nitrate-N	0.17	0.091	1.23	1.12	< 0.002	0.01	1.42	1.43	1.95	1.41	< 0.02	< 0.002	0.37	< 0.002
TKN	< 0.10	< 0.10		< 0.10		0.7		0.1		0.11	6.3	21	1.6	10.7
Sulphate	4.3	2.4	6.9	7.9	0.6	0.9	5.8	6.3	29	17	158	470	71	165
Dissolved arsenic	< 0.0010	< 0.0010	< 0.0010	< 0.0010	0.0075	0.017	< 0.0010	< 0.0010	< 0.0010	< 0.0010	0.026	0.12	0.0049	0.014
Dissolved cadmium	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005
Dissolved chromium	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	0.013	0.0047	0.0017	0.0055
Dissolved copper	0.193	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	0.0013	0.0007
Dissolved lead	0.006	< 0.00010	< 0.00010	< 0.00010	< 0.00010	< 0.00010	< 0.00010	< 0.00010	< 0.00010	< 0.00010	< 0.00010	0.0003	0.00014	0.00036
Dissolved nickel	0.001	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	0.0019	0.0039	0.002	0.0044
Dissolved zinc	0.15	< 0.0010	0.0034	0.0021	< 0.0010	0.0014	0.0033	0.0018	0.0018	0.0014	0.0091	0.0087	0.0027	0.008

Notes:

All values provided in units of milligrams per liter (mg/L) unless otherwise specified.

< Below the laboratory reporting limit (<RL)



- Legend**
- GRE001
 - Grey River Sample Location
 - Upgradient Monitoring Well
 - Crossgradient Sample Locations
 - Downgradient Sample Locations
 - GRE001 Abstraction Wells
 - Approximate Site Boundary

0 50 m 100 m
LINZ CC BY 4.0 © Imagery Basemap contributors

ENGEO

Produced by **Datanest.earth**

Title: Ion Assessment Sample Locations		
Client: Taylorville Resource Park Ltd		Size: A4
Project: Taylorville Resource Park	Drawn: LN	Figure: A5.1
Date: 28-01-2025	Checked: DR	
Proj No: 26200.000.001	Scale: 1:5000	Version: 1.0

Figure A5.2 Extended Durov Class Scatter

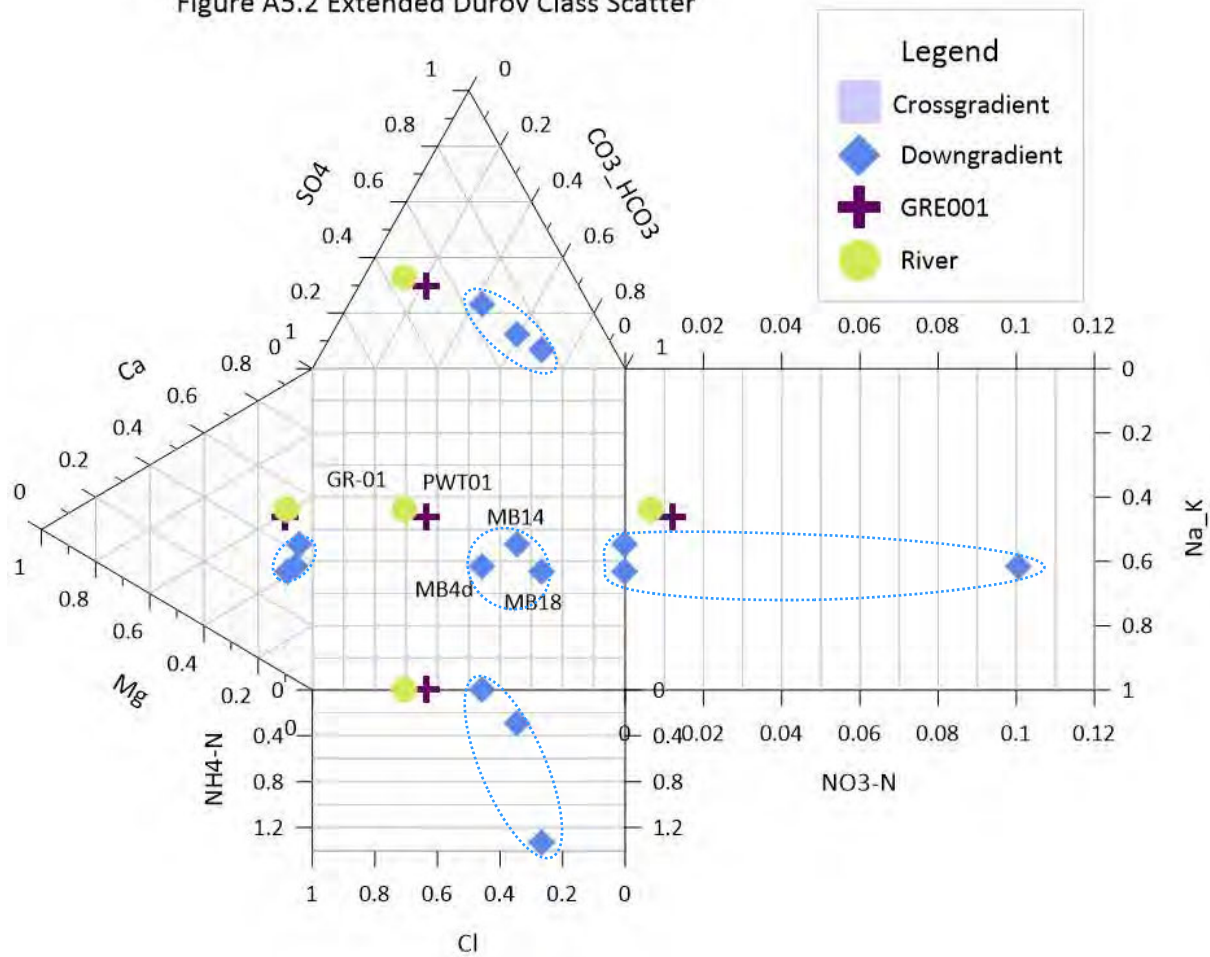
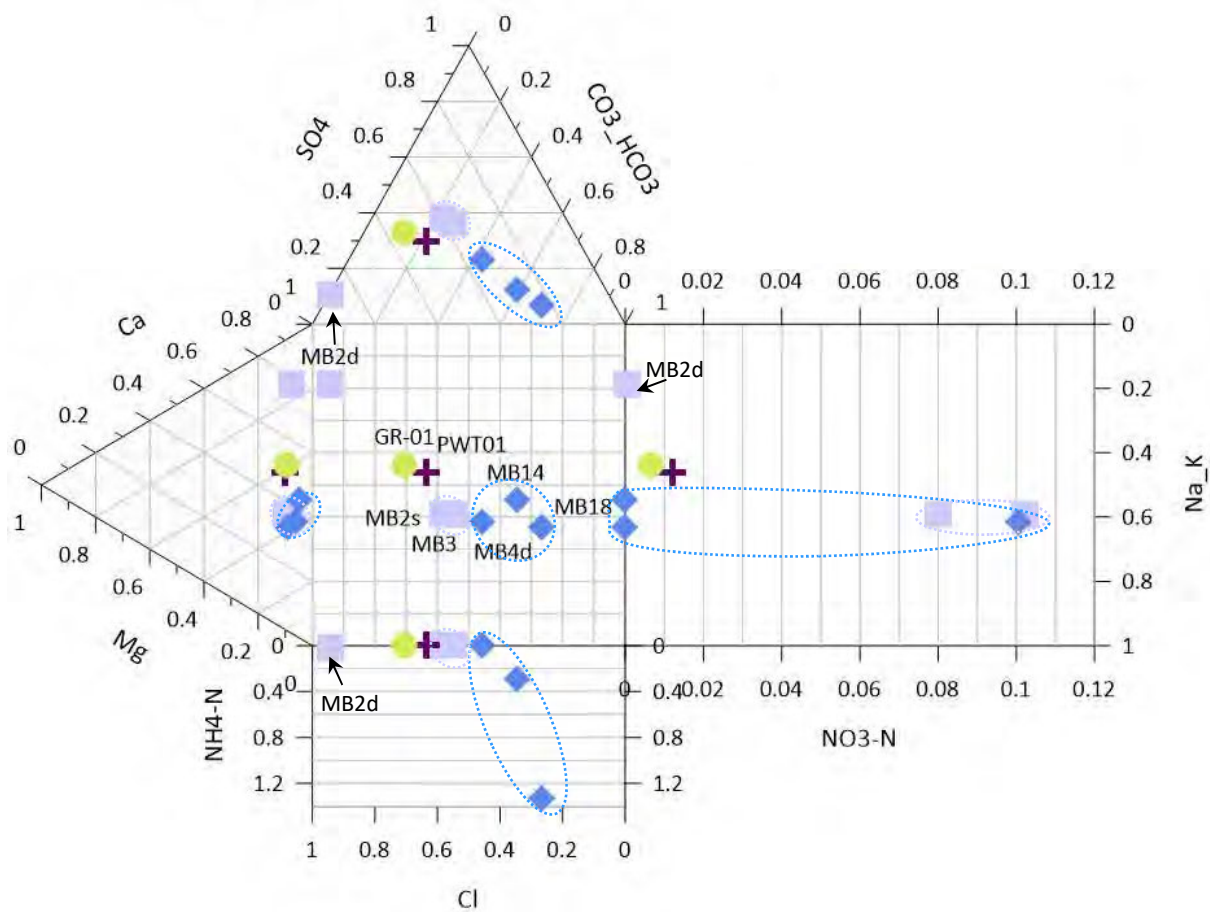


Figure A5.3 Extended Durov Class Scatter





APPENDIX 6

Aquifer Testing Information

McMILLAN Drilling

FINAL DELIVERABLES FOR GREY DISTRICT COUNCIL

Coal Creek WTP Bore Installation (2021-22-26a) - Drilling
and testing of one new production bore



Handover Documentation

Coal Creek WTP Bore Installation (2021-22-26a)

☒ Producer Statement

BH01 (decommissioned)

☒ Final Bore Log

BH02 (production)

☒ Final Bore Log

☒ Step Rate Pumping Test - 9/05/2023

- Graphs
- Manual readings - pumped bore

☒ Constant Discharge Pumping Test - 9/05/2023 - 10/05/2023

- Graphs
- Manual readings - pumped bore

☒ Mill Test Certificate

- DN300 casing - stainless steel

Separate Files

☒ Sentinel Data includes drawdowns and flowrates (vented level transducer)

SIXTH SCHEDULE**FORM OF PRODUCER STATEMENT - CONSTRUCTION**

ISSUED BY:McMillan Drilling Ltd (Contractor)

TO:Grey District Council (Principal)

IN RESPECT OF:..... Drilling of one New Well

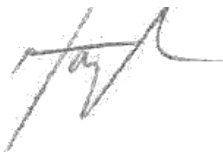
AT: Coal Creek, Greymouth

McMillan Drilling Ltd (Contractor) has contracted to Grey District Council (Principal) to carry out and complete works in accordance with the contract documents titled **GDC 2021-22-26a Coal Creek WTP Bore Installation - Drilling Investigation**.

- BH01 - 10.8m deep (decommissioned)
- BH02 - 10.5m deep (production)

I Iain Haycock a duly authorised representative of McMillan Drilling Ltd believe on reasonable grounds that McMillan Drilling Ltd has carried out and completed:

- ☒ All works as specified in the contract and all works were carried out in accordance with NZS 4411:2001 (Environmental Standard for Drilling of Soil and Rock) and that the wells have been adequately sealed to prevent ingress of contaminants from the surface.
- ☐ Part only as specified in the attached particulars of the contract works in accordance with the Contract.



(Signature of Authorized Agent on behalf of)

Date: 11/05/2023

McMillan Drilling Ltd
(Contractor)

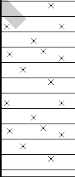
120 High Street Southbridge, 7602
(Address)



Client:	Grey District Council	Bore No.:	BH01
Project:	Drilling of New 12" (300mm) Well	Job No.:	20874

Site Location: Coal Creek, GREYMOUTH
Grid Reference: 1457012m E, 5298980m N (NZTM) Map or aerial photograph
Lot Number:

Date: 2/12/2022
Rig Operator: A. Harkerss
Method / Rig: Cable Tool

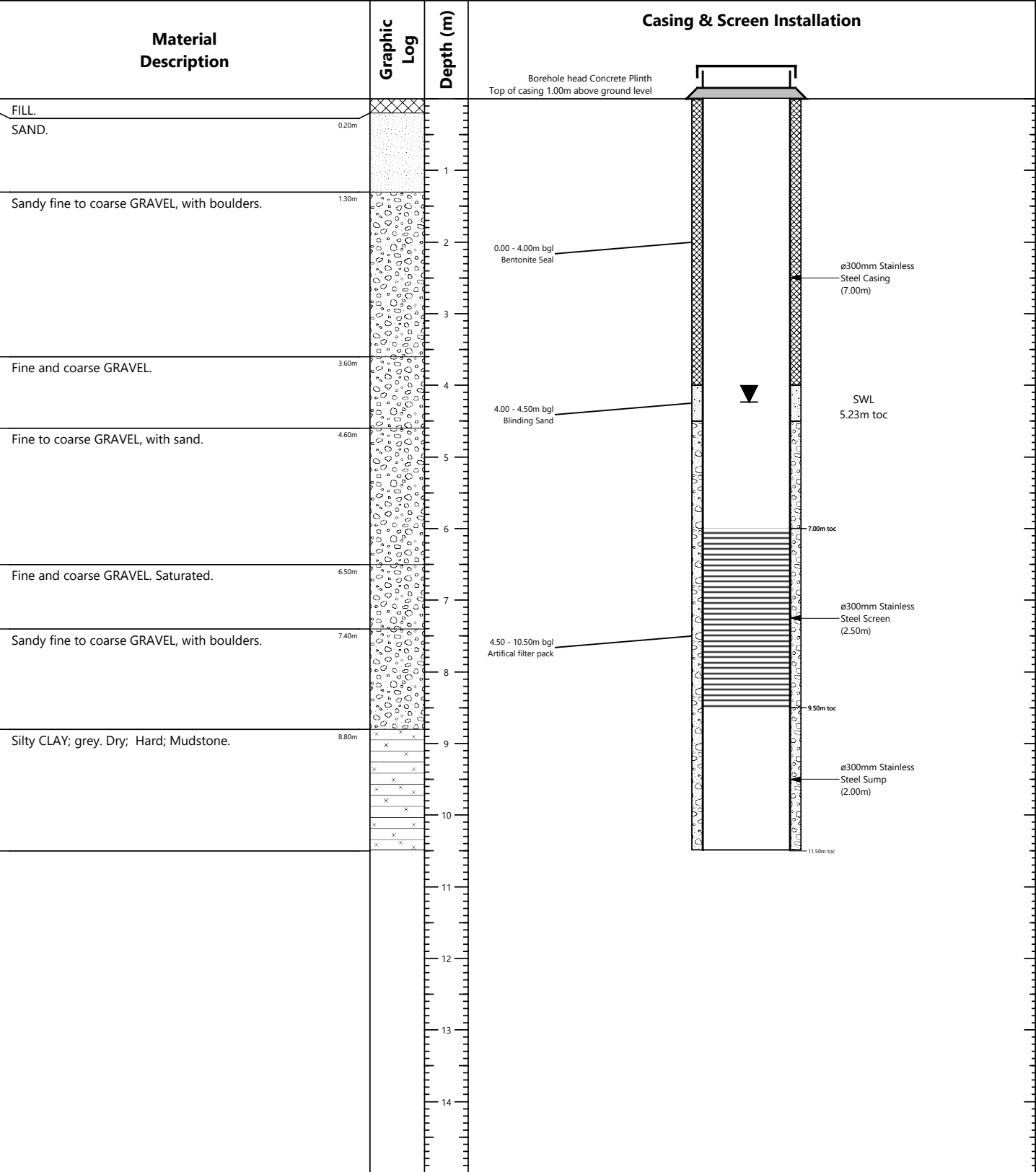
Material Description	Graphic Log	Depth (m)	Casing & Screen Installation
FILL.			
CLAY; brown.		0.40m	
Sandy fine to coarse GRAVEL.		0.80m	
Coarse SAND, with gravel and boulders; grey.		2.90m	
Sandy fine to coarse GRAVEL; grey. Some staining.		8.00m	
Clayey GRAVEL; brown.		8.60m	
Silty CLAY; grey. Dry; Mudstone.		8.70m	
		11	
		12	
		13	
		14	

Remarks: Insufficient water found - all casing removed, backfilled & sealed to surface.	Development Duration (hr):	Flow Rate (gpm) (l/sec)	Draw Down (m)	Duration (hr)
Bore Diameter: 450mm Total Bore Depth: -	Static Water Level: - Total Depth Drilled: 10.80m bgl	Pumping Duration (hr):		



Client:	Grey District Council	Bore No.:	BH02
	Project: Drilling of New 12" (300mm) Well	Job No.:	20874

Site Location: Coal Creek, GREYMOUTH	Date: 11/5/2023
Grid Reference: 1457266m E, 5298901m N (NZTM) Map or aerial photograph	Rig Operator: A. Harkerss, Z. Luxton
Lot Number:	Method / Rig: Cable Tool



Remarks: Production Bore Bore completed with DN300 table E flange, blank and dressing kit. 24hr constant rate test influenced by rising river levels - resultant draw down not reported (refer electronic data & graphs)	Development Duration (hr): 17.25	Flow Rate		Draw Down	Duration
	Pumping Duration (hr): 28	(gpm)	(l/sec)	(m)	(hr)
Bore Diameter: 450mm	Static Water Level: 5.23m btoc	528	40.0	0.41	1.00
Total Bore Depth: 11.50m toc	Total Depth Drilled: 10.50m bgl	660	50.0	0.61	1.00
		792	60.0	0.79	1.00
		924	70.0	1.03	1.00



Client:

Grey District Council

Bore No.:

BH02 - Production

Project:

Coal Creek WTP Bore Installation (2021-22-26a)

Job No.:

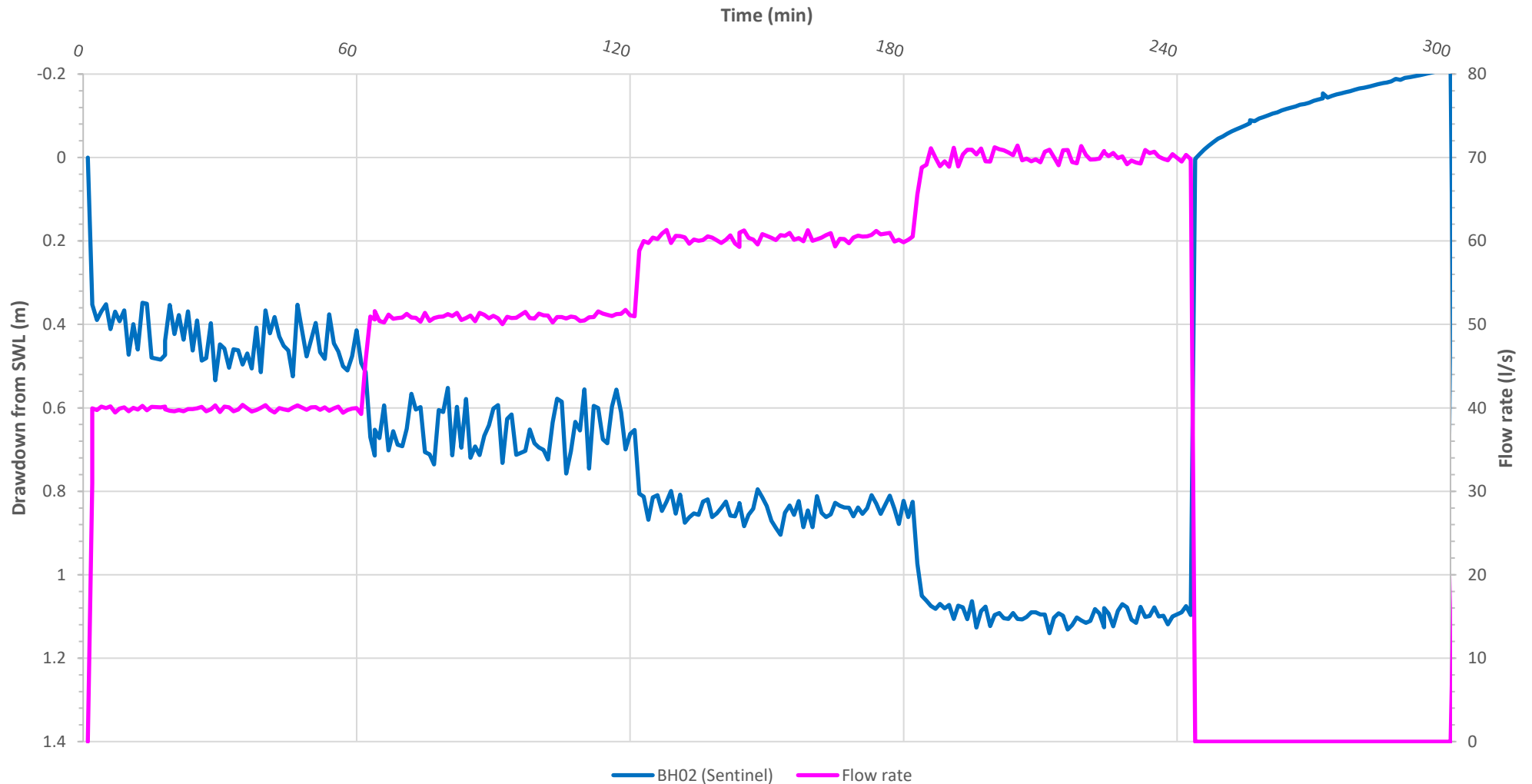
20874

Stepped Drawdown and Sand Test

Site Location: Coal Creek, GREYMOUTH

Static water level: 4.77m btoc

Date: 9/5/2023

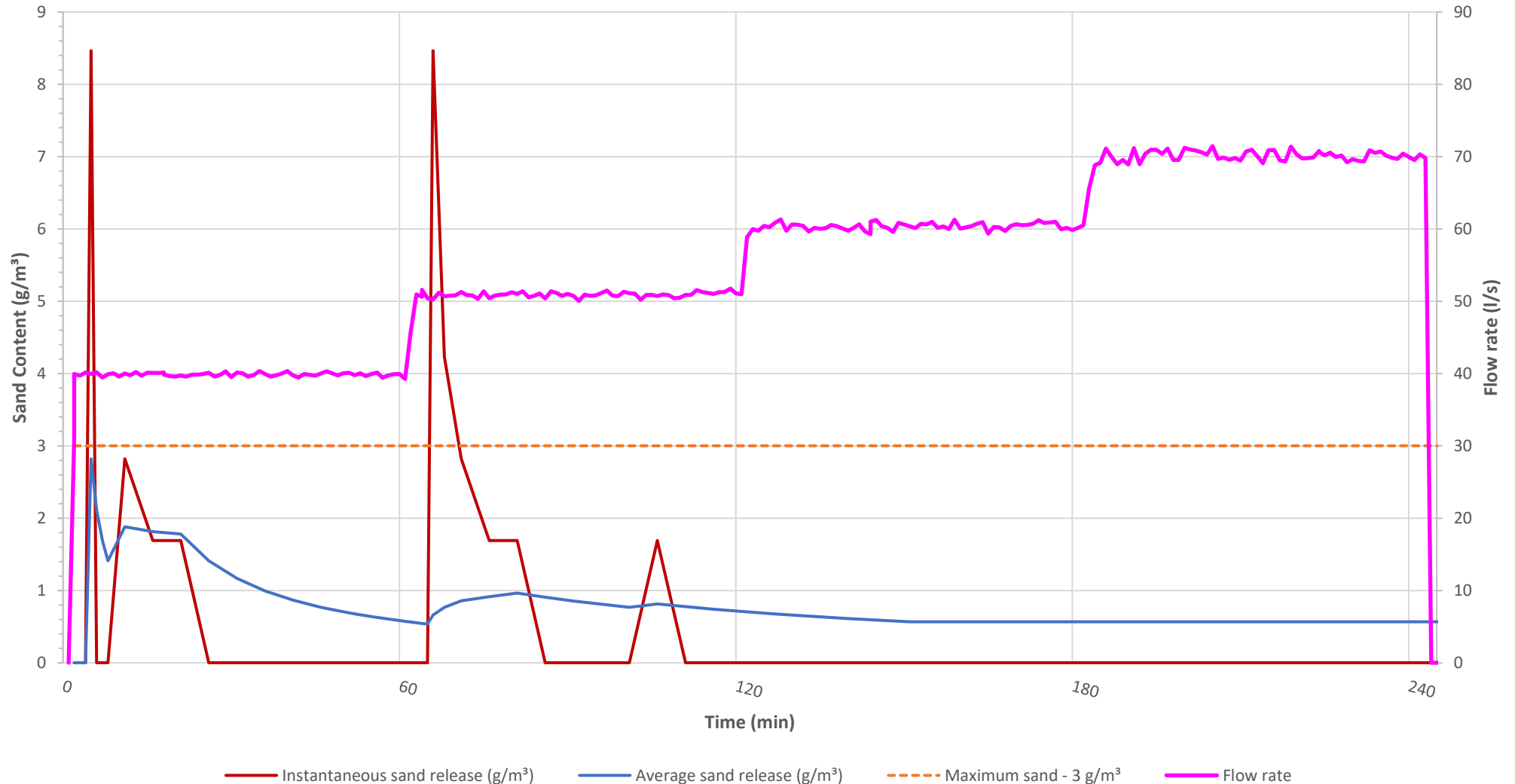


Client:	Grey District Council	Bore No.:	BH02 - Production
Project:	Coal Creek WTP Bore Installation (2021-22-26a)	Job No.:	20874

Sand release test

Site Location: Coal Creek, GREYMOUTH

Date: 9/5/2023



	Client: Grey District Council	Bore No.: BH02 - Production
	Project: Coal Creek WTP Bore Installation (2021-22-26a)	Job No.: 20874

Stepped Drawdown and Sand Test

Site Location: Coal Creek, GREYMOUTH

Start date: 9/5/2023

Water level: 4.77m btoc (SWL)

Finish date: 9/5/2023

Top of casing: 1.00m agl

Well depth: 10.50m toc

Step / min	Date	Time	Water depth (m)	Drawdown (m)	Pumping rate (l/s)	Sand (ml)	Comments
1 / 1	9/5/2023	8:01	5.15	0.38	40.09	0.000	
1 / 2	9/5/2023	8:02	5.15	0.38	40.07	0.000	
1 / 3	9/5/2023	8:03	5.15	0.38	40.07	0.000	
1 / 4	9/5/2023	8:04	5.15	0.38	40.00	0.010	
1 / 5	9/5/2023	8:05	5.15	0.38	40.30	0.010	
1 / 6	9/5/2023	8:06	5.15	0.38	40.01	0.010	
1 / 7	9/5/2023	8:07	5.15	0.38	40.00	0.010	
1 / 10	9/5/2023	8:10	5.15	0.38	40.30	0.010	
1 / 15	9/5/2023	8:15	5.16	0.39	40.20	0.010	
1 / 20	9/5/2023	8:20	5.17	0.40	40.10	0.010	
1 / 25	9/5/2023	8:25	5.17	0.40	40.00	0.010	
1 / 30	9/5/2023	8:30	5.17	0.40	40.40	0.010	
1 / 35	9/5/2023	8:35	5.17	0.40	40.10	0.010	
1 / 40	9/5/2023	8:40	5.18	0.41	40.00	0.010	
1 / 45	9/5/2023	8:45	5.18	0.41	40.10	0.010	
1 / 50	9/5/2023	8:50	5.18	0.41	39.90	0.010	
1 / 55	9/5/2023	8:55	5.18	0.41	39.90	0.010	
1 / 60	9/5/2023	9:00	5.18	0.41	40.00	0.010	

	Client: Grey District Council					Bore No.: BH02 - Production	
	Project: Coal Creek WTP Bore Installation (2021-22-26a)					Job No.: 20874	
Stepped Drawdown and Sand Test							
Site Location: Coal Creek, GREYMOUTH					Start date: 9/5/2023		
Water level: 4.77m btoc (SWL)					Finish date: 9/5/2023		
Top of casing: 1.00m agl					Well depth: 10.50m toc		
Step / min	Date	Time	Water depth (m)	Drawdown (m)	Pumping rate (l/s)	Sand (ml)	Comments
2 / 1	9/5/2023	9:01	5.35	0.58	49.80	0.000	
2 / 2	9/5/2023	9:02	5.36	0.59	51.80	0.000	
2 / 3	9/5/2023	9:03	5.36	0.59	50.90	0.000	
2 / 4	9/5/2023	9:04	5.37	0.60	50.70	0.000	
2 / 5	9/5/2023	9:05	5.37	0.60	50.70	0.010	
2 / 7	9/5/2023	9:07	5.37	0.60	50.90	0.010	
2 / 10	9/5/2023	9:10	5.37	0.60	51.30	0.010	
2 / 15	9/5/2023	9:15	5.37	0.60	51.00	0.010	
2 / 20	9/5/2023	9:20	5.37	0.60	51.20	0.010	
2 / 25	9/5/2023	9:25	5.37	0.60	51.00	0.010	
2 / 30	9/5/2023	9:30	5.38	0.61	50.90	0.010	
2 / 35	9/5/2023	9:35	5.38	0.61	51.10	0.010	
2 / 40	9/5/2023	9:40	5.38	0.61	51.30	0.010	
2 / 45	9/5/2023	9:45	5.38	0.61	51.80	0.020	
2 / 50	9/5/2023	9:50	5.38	0.61	51.00	0.020	
2 / 55	9/5/2023	9:55	5.38	0.61	51.80	0.020	
2 / 60	9/5/2023	10:00	5.38	0.61	50.50	0.020	

	Client: Grey District Council					Bore No.: BH02 - Production	
	Project: Coal Creek WTP Bore Installation (2021-22-26a)					Job No.: 20874	
Stepped Drawdown and Sand Test							
Site Location: Coal Creek, GREYMOUTH					Start date: 9/5/2023		
Water level: 4.77m btoc (SWL)					Finish date: 9/5/2023		
Top of casing: 1.00m agl					Well depth: 10.50m toc		
Step / min	Date	Time	Water depth (m)	Drawdown (m)	Pumping rate (l/s)	Sand (ml)	Comments
3 / 1	9/5/2023	10:01	5.54	0.77	60.10	0.000	
3 / 2	9/5/2023	10:02	5.55	0.78	60.60	0.000	
3 / 3	9/5/2023	10:03	5.55	0.78	60.50	0.000	
3 / 4	9/5/2023	10:04	5.55	0.78	60.30	0.000	
3 / 5	9/5/2023	10:05	5.55	0.78	61.10	0.000	
3 / 7	9/5/2023	10:07	5.55	0.78	59.90	0.000	
3 / 10	9/5/2023	10:10	5.56	0.79	60.60	0.000	
3 / 15	9/5/2023	10:15	5.56	0.79	60.90	0.000	
3 / 20	9/5/2023	10:20	5.56	0.79	61.30	0.000	
3 / 25	9/5/2023	10:25	5.56	0.79	60.80	0.000	
3 / 30	9/5/2023	10:30	5.56	0.79	61.00	0.000	
3 / 35	9/5/2023	10:35	5.56	0.79	60.50		
3 / 40	9/5/2023	10:40	5.56	0.79	61.00		
3 / 45	9/5/2023	10:45	5.56	0.79	60.70		
3 / 50	9/5/2023	10:50	5.56	0.79	60.90		
3 / 55	9/5/2023	10:55	5.56	0.79	60.90		
3 / 60	9/5/2023	11:00	5.56	0.79	60.90		

	Client: Grey District Council					Bore No.: BH02 - Production	
	Project: Coal Creek WTP Bore Installation (2021-22-26a)					Job No.: 20874	
Stepped Drawdown and Sand Test							
Site Location: Coal Creek, GREYMOUTH					Start date: 9/5/2023		
Water level: 4.77m btoc (SWL)					Finish date: 9/5/2023		
Top of casing: 1.00m agl					Well depth: 10.50m toc		
Step / min	Date	Time	Water depth (m)	Drawdown (m)	Pumping rate (l/s)	Sand (ml)	Comments
4 / 1	9/5/2023	11:02	5.74	0.97	70.00		
4 / 2	9/5/2023	11:03	5.76	0.99	69.70		
4 / 3	9/5/2023	11:04	5.77	1.00	70.00		
4 / 4	9/5/2023	11:05	5.77	1.00	70.30		
4 / 5	9/5/2023	11:06	5.77	1.00	69.70		
4 / 7	9/5/2023	11:08	5.78	1.01	70.10		
4 / 10	9/5/2023	11:11	5.78	1.01	70.70		
4 / 15	9/5/2023	11:16	5.78	1.01	70.40		
4 / 20	9/5/2023	11:21	5.79	1.02	70.40		
4 / 25	9/5/2023	11:26	5.80	1.03	70.90		
4 / 30	9/5/2023	11:31	5.80	1.03	71.30		
4 / 35	9/5/2023	11:36	5.80	1.03	71.00		
4 / 40	9/5/2023	11:41	5.80	1.03	70.60		
4 / 45	9/5/2023	11:46	5.80	1.03	70.60		
4 / 50	9/5/2023	11:51	5.80	1.03	71.10		
4 / 55	9/5/2023	11:56	5.80	1.03	69.80		
4 / 60	9/5/2023	12:01	5.80	1.03	70.50		

	Client: Grey District Council	Bore No.: BH02 - Production
	Project: Coal Creek WTP Bore Installation (2021-22-26a)	Job No.: 20874

Stepped Drawdown and Sand Test

Site Location: Coal Creek, GREYMOUTH	Start date: 9/5/2023
Water level: 4.77m btoc (SWL)	Finish date: 9/5/2023
Top of casing: 1.00m agl	Well depth: 10.50m toc

Step / min	Date	Time	Water depth (m)	Drawdown (m)	Pumping rate (l/s)	Sand (ml)	Comments
Rec / 1	9/5/2023	12:03	4.78	0.01			
Rec / 2	9/5/2023	12:04	4.75	-0.02			
Rec / 3	9/5/2023	12:05	4.76	-0.01			
Rec / 4	9/5/2023	12:06	4.74	-0.03			
Rec / 5	9/5/2023	12:07	4.72	-0.05			
Rec / 7	9/5/2023	12:09	4.70	-0.07			
Rec / 10	9/5/2023	12:12	4.69	-0.08			
Rec / 15	9/5/2023	12:17	4.68	-0.09			
Rec / 20	9/5/2023	12:22	4.67	-0.10			
Rec / 25	9/5/2023	12:27	4.65	-0.12			
Rec / 30	9/5/2023	12:32	4.64	-0.13			
Rec / 35	9/5/2023	12:37	4.62	-0.15			
Rec / 40	9/5/2023	12:42	4.61	-0.16			
Rec / 45	9/5/2023	12:47	4.60	-0.17			
Rec / 50	9/5/2023	12:52	4.60	-0.17			
Rec / 55	9/5/2023	12:57	4.60	-0.17			
Rec / 60	9/5/2023	13:02	4.59	-0.18			

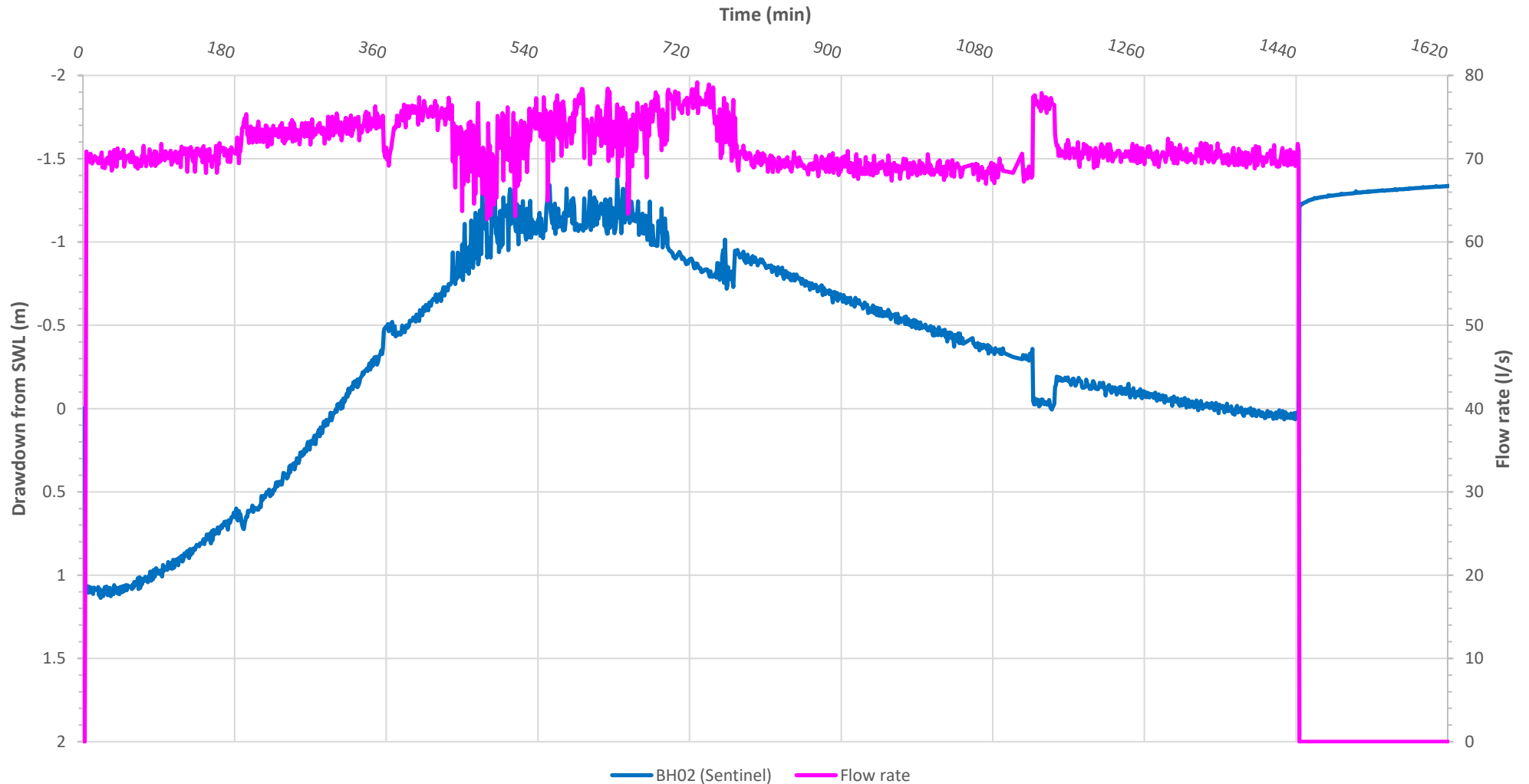
Client:	Grey District Council	Bore No.:	BH02 - Production
Project:	Coal Creek WTP Bore Installation (2021-22-26a)	Job No.:	20874

Constant Rate and Sand Test

Site Location: Coal Creek, GREYMOUTH

Static water level: 4.60m btoc

Date: 9/5/2023

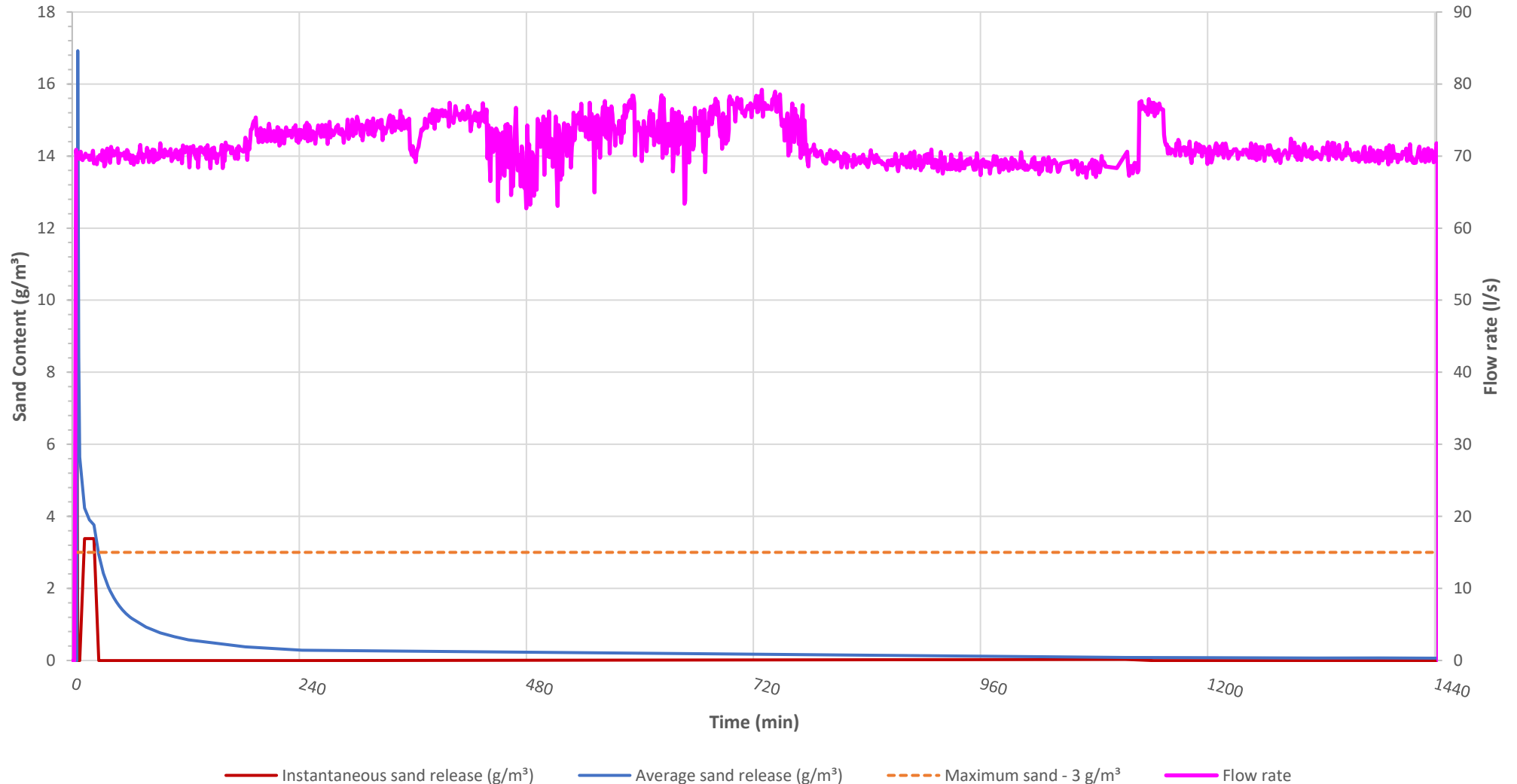


Client:	Grey District Council	Bore No.:	BH02 - Production
Project:	Coal Creek WTP Bore Installation (2021-22-26a)	Job No.:	20874

Sand release test

Site Location: Coal Creek, GREYMOUTH

Date: 9/5/2023



	Client: Grey District Council					Bore No.: BH02 - Production	
	Project: Coal Creek WTP Bore Installation (2021-22-26a)					Job No.: 20874	
Constant Rate and Sand Test							
Site Location: Coal Creek, GREYMOUTH					Start date: 9/5/2023		
Water level: 4.60m btoc (SWL)					Finish date: 11/5/2023		
Top of casing: 1.00m agl					Well depth: 10.50m toc		
Step / min	Date	Time	Water depth (m)	Drawdown (m)	Pumping rate (l/s)	Sand (ml)	Comments
Constant rate	9/5/2023	13:01	5.58	0.98	70.8	0.010	
Constant rate	9/5/2023	13:02	5.58	0.98	70.70	0.020	
Constant rate	9/5/2023	13:03	5.59	0.99	70.00	0.020	
Constant rate	9/5/2023	13:04	5.59	0.99	70.10	0.020	
Constant rate	9/5/2023	13:05	5.60	1.00	70.90	0.020	
Constant rate	9/5/2023	13:10	5.60	1.00	70.40	0.020	
Constant rate	9/5/2023	13:15	5.60	1.00	70.60	0.020	
Constant rate	9/5/2023	13:20	5.60	1.00	70.20	0.020	
Constant rate	9/5/2023	13:25	5.60	1.00	70.80	0.020	
Constant rate	9/5/2023	13:30	5.60	1.00	69.90	0.020	
Constant rate	9/5/2023	13:35	5.60	1.00	70.40	0.020	
Constant rate	9/5/2023	13:40	5.60	1.00	70.10	0.020	
Constant rate	9/5/2023	13:45	5.60	1.00	70.40	0.020	
Constant rate	9/5/2023	13:50	5.60	1.00	70.60	0.020	
Constant rate	9/5/2023	13:55	5.60	1.00	70.00	0.020	
Constant rate	9/5/2023	14:00	5.59	0.99	70.20	0.020	
Constant rate	9/5/2023	14:15	5.55	0.95	71.40	0.020	Heavy rain - water in river rising
Constant rate	9/5/2023	14:30	5.50	0.90	70.90	0.020	
Constant rate	9/5/2023	14:45	5.45	0.85	70.60	0.020	
Constant rate	9/5/2023	15:00	5.40	0.80	70.30	0.020	
Constant rate	9/5/2023	16:00	5.20	0.60	71.40	0.020	
Constant rate	9/5/2023	17:00	4.94	0.34	72.60	0.020	
Constant rate	10/5/2023	7:30	4.31	-0.29	69.10	0.050	
Constant rate	10/5/2023	8:00	4.58	-0.02	75.10	0.050	Flowrate adjusted
Constant rate	10/5/2023	9:00	4.47	-0.13	69.90	0.050	
Constant rate	10/5/2023	10:00	4.51	-0.09	71.50	0.050	
Constant rate	10/5/2023	11:00	4.58	-0.02	71.10	0.050	
Constant rate	10/5/2023	12:00	4.61	0.01	71.00	0.050	
Constant rate	10/5/2023	13:00	4.65	0.05	69.90	0.050	

	Client: Grey District Council					Bore No.: BH02 - Production	
	Project: Coal Creek WTP Bore Installation (2021-22-26a)					Job No.: 20874	
Constant Rate and Sand Test							
Site Location: Coal Creek, GREYMOUTH					Start date: 9/5/2023		
Water level: 4.60m btoc (SWL)					Finish date: 11/5/2023		
Top of casing: 1.00m agl					Well depth: 10.50m toc		
Step / min	Date	Time	Water depth (m)	Drawdown (m)	Pumping rate (l/s)	Sand (ml)	Comments
Rec / 1	10/5/2023	13:01	3.50	-1.10			
Rec / 2	10/5/2023	13:02	3.45	-1.15			
Rec / 3	10/5/2023	13:03	3.44	-1.16			
Rec / 4	10/5/2023	13:04	3.44	-1.16			
Rec / 5	10/5/2023	13:05	3.44	-1.16			
Rec / 7	10/5/2023	13:07	3.45	-1.15			
Rec / 10	10/5/2023	13:10	3.45	-1.15			
Rec / 15	10/5/2023	13:15	3.43	-1.17			
Rec / 20	10/5/2023	13:20	3.44	-1.16			
Rec / 25	10/5/2023	13:25	3.43	-1.17			
Rec / 30	10/5/2023	13:30	3.40	-1.20			
Rec / 35	10/5/2023	13:35	3.40	-1.20			
Rec / 40	10/5/2023	13:40	3.40	-1.20			
Rec / 45	10/5/2023	13:45	3.39	-1.21			
Rec / 50	10/5/2023	13:50	3.39	-1.21			
Rec / 55	10/5/2023	13:55	3.38	-1.22			
Rec / 60	10/5/2023	14:00	3.38	-1.22			
Rec / 90	10/5/2023	14:30	3.37	-1.23			
Rec / 120	10/5/2023	15:00	3.36	-1.24			
Rec / 150	10/5/2023	15:30	3.34	-1.26			
Rec / 180	10/5/2023	16:00	3.34	-1.26			
Rec / 1140	11/5/2023	8:00	3.45	-1.15			

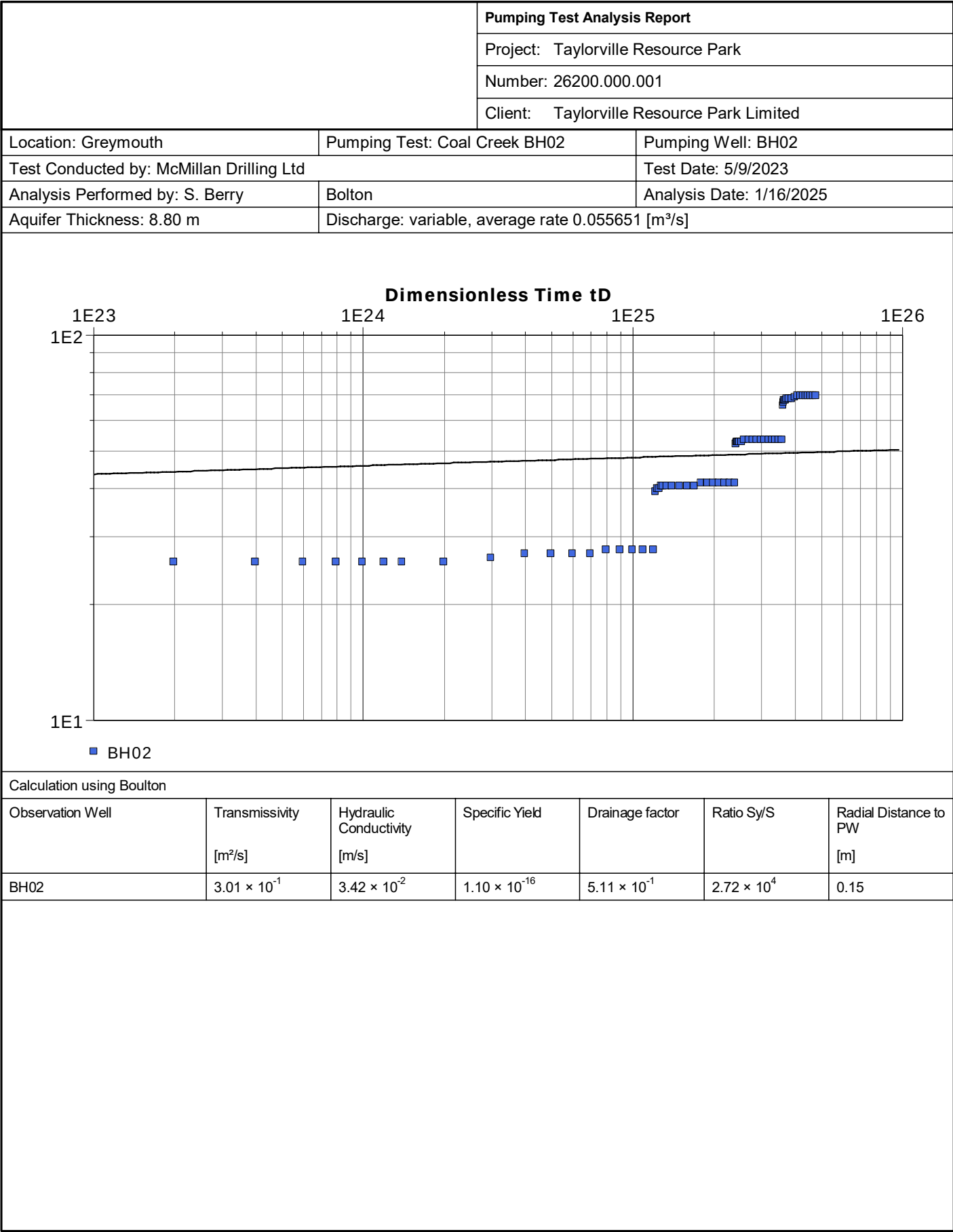


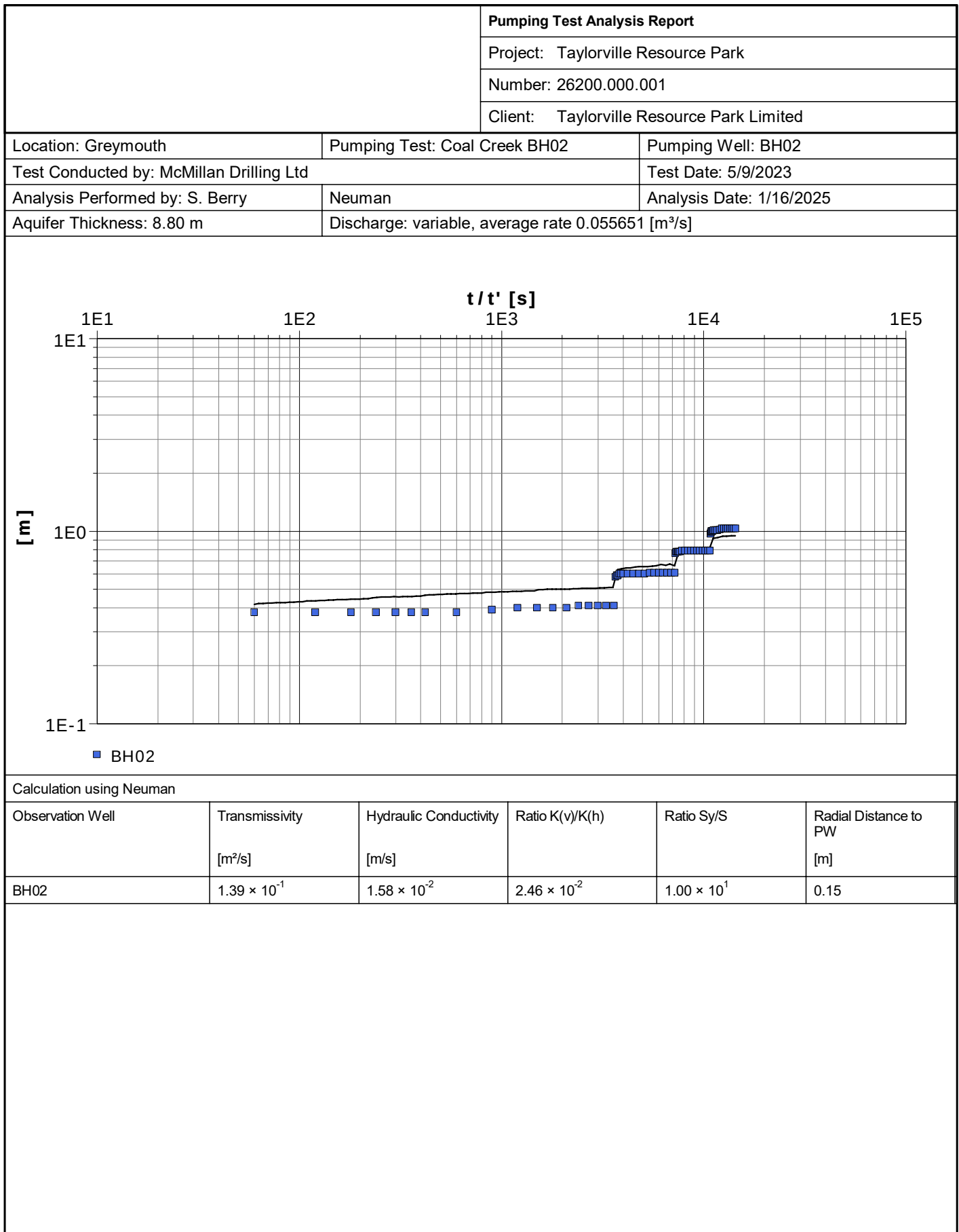
Customer	STAINLESS PRODUCTS LTD			Customer No.	FHK076	Certificate No.	211220P0123
Product	STAINLESS STEEL WELDED PIPE			Commercial Invoice No.	J21110271	Date of Issue	2021/12/20
Specification	ASTM A312/A312M-21			PO No.	6853	Order No.	EP21100179-1
Steel Grade	TP304/304L	Finish	PICKLED	P.I.No.	PI 2110-114	Weight No.	C21120150

Item No.	Heat No.	Product Description		Chemical Composition(%)								Tension Test			*01 HR	*02 HT	*03 HY	*04 ET	*05 F1
				Quantity Pcs	C	Si	Mn	P	S	Ni	Cr	0.2% Y.S.	T.S.	E.L.					
		Dimension	X10 ⁻³		X10 ⁻²	X10 ⁻²	X10 ⁻³	X10 ⁻³	X10 ⁻²	X10 ⁻²	MPa	%	HRB	°C	MPa				
			Specification		Min.														
			Max.	35	100	200	45	30	1100	2000									
1	VCB111262	12"(323.9mm) x SCH-10S(4.57mm) x 6M	1	20	37	140	29	1	801	1830	307	650	58	86	1060	3.0	-	OK	
2	VCB111696	12"(323.9mm) x SCH-10S(4.57mm) x 6M	3	24	43	141	32	1	801	1803	305	632	56	85	1060	3.0	-	OK	
3	VCB111910	12"(323.9mm) x SCH-10S(4.57mm) x 6M	1	20	37	140	29	1	801	1830	307	650	58	86	1060	3.0	-	OK	
4	PCB110416	1"(33.4mm) x SCH-40S(3.38mm) x 6M	35	23	42	141	29	2	802	1805	304	655	56	86	1060	-	OK	OK	
5	PCB110479	1"(33.4mm) x SCH-40S(3.38mm) x 6M	5	26	39	139	37	1	801	1820	292	625	56	86	1060	-	OK	OK	
6	PCB100754	1 1/4"(42.2mm) x SCH-40S(3.56mm) x 6M	15	20	39	138	28	2	802	1825	286	643	59	86	1060	-	OK	OK	
7	PCB110792	1 1/2"(48.3mm) x SCH-40S(3.68mm) x 6M	15	25	35	138	33	2	801	1811	285	629	56	85	1060	-	OK	OK	
8	PCB110262	3"(88.9mm) x SCH-40S(5.49mm) x 6M	15	22	35	137	32	1	801	1820	293	623	59	84	1060	-	OK	OK	
9	PCB110605	6"(168.3mm) x SCH-40S(7.11mm) x 6M	2	25	39	129	30	1	801	1831	331	654	54	87	1060	9.0	-	OK	
TOTAL:			92																

Remark	Note
	Hardness according to NACE MR0175 / ISO 15156-3:2015 : Satisfactory.
	Visual Inspection(Surface) and Dimension Check : OK

Surveyor By	We hereby certify that material described herein has been manufactured and tested with satisfactory results in accordance with the requirement of the above material specification. The material described above has been found free of radiation by raw material supplier.	1.ISO 9001:2015 (No.01 100 056299) 2.PED / AD2000 (No.01 202 TWN/Q-05 0105) 3.ISO 14001:2015 (No.01 104 822 096385) 4.ISO 45001:2018 (No.01 113 822 096385)	Manager of Quality Assurance Department
N/A			<i>Russell Wu</i>





APPENDIX 7

XRD Analysis

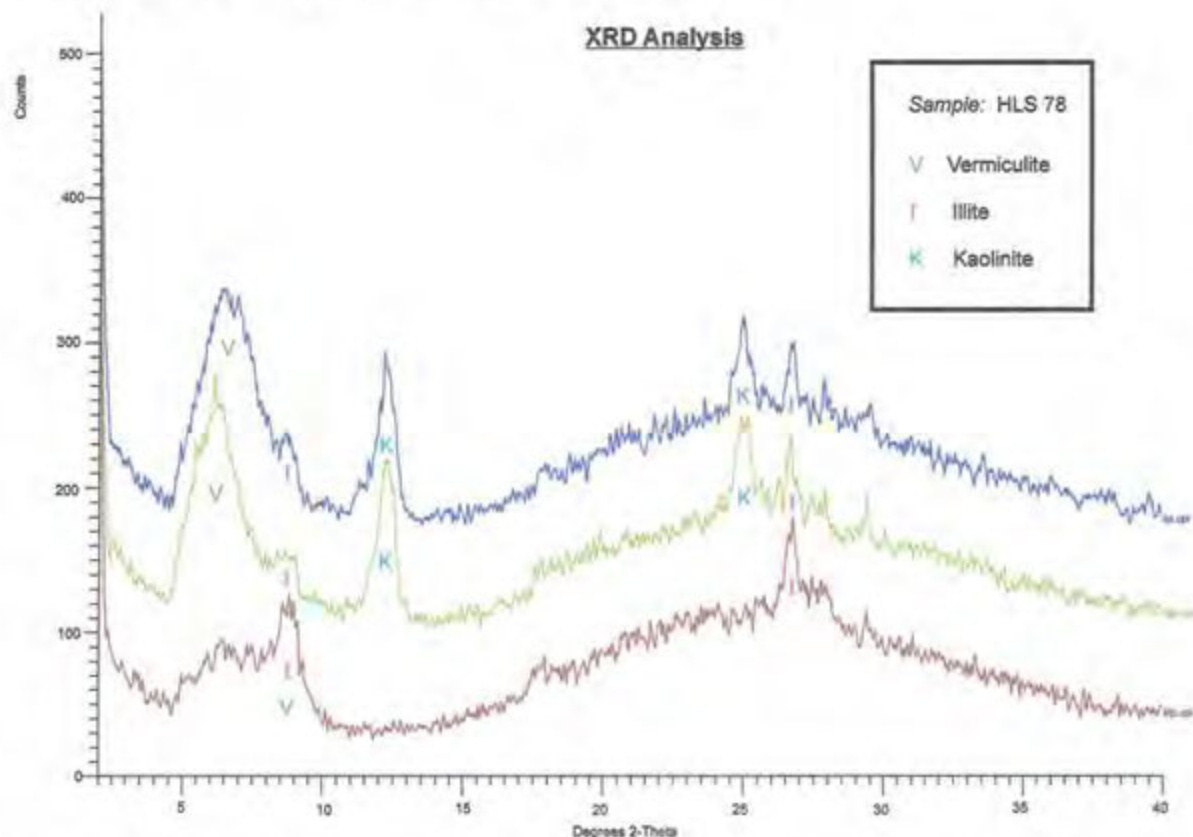


Figure 4.12: XRD analysis of Kaiata Mudstone sample HLS 78 from the Punakaiki River Gorge. Another sample, HLS 34 at Mile River, gave the same results at a lesser concentration.

TO: Taylorville Resource Park
COPY TO: Jimmy Whitmore (Engeo)
FROM: Laura Drummond

Date: 7 February 2025
Job No: 67662

TAYLORVILLE RESOURCE PARK S92 RESPONSE - ECOLOGY

Bioresearches were engaged by Engeo, on behalf of their client Taylorville Resource Park (TRP), to undertake an ecological impact assessment for the discharge of water associated with landfill operations, for the temporary discharge of leachate contaminated water to land (RC-2024-0122).

This memorandum addresses the questions raised under Section 92 (S92) of the RMA by Pattle Delmore Partners (PDP), on behalf of West Coast Regional Council.

The S92 ecology related question is copied in blue italics below, with the responses in black:

4. [West Coast Regional Council requests] An ecological survey of the surface water features and aquatic habitat surrounding the site, including Noname Creek, the drain along Taylorville Road, including that part of the drain that begins at spring seep TCD1 and the streams on the southern side of Taylorville Road (excluding the Grey River) and an impact assessment on surface water quality and ecological values resulting from the proposed discharge.

The Ecological Impact Assessment (EclIA) assessed the ecological values of the roadside drain, Noname Creek, and the tributary of the Grey River (Grey Tributary) which receives surface water from the surrounding environment (including the roadside drain, any stormwater runoff etc). This assessment generally follows the EclIA Guidelines for use in New Zealand published by the Environmental Institute of Australia and New Zealand (EIANZ) (Roper-Lindsay *et al.*, 2017¹). The EclIA Guidelines provide a standardised matrix framework that allows ecological effects assessments to be clear, transparent, and consistent. The EclIAG framework is generally used in Ecological Impact Assessments in New Zealand as good practice, and a detailed analysis of this methodology is presented Appendix A.

¹ Roper-Lindsay, J., Fuller S.A., Hooson, S., Sanders, M.D., Ussher, G.T. (2018). Ecological impact assessment. EIANZ guidelines for use in New Zealand: terrestrial and freshwater ecosystems. 2nd edition.

Ecological Values

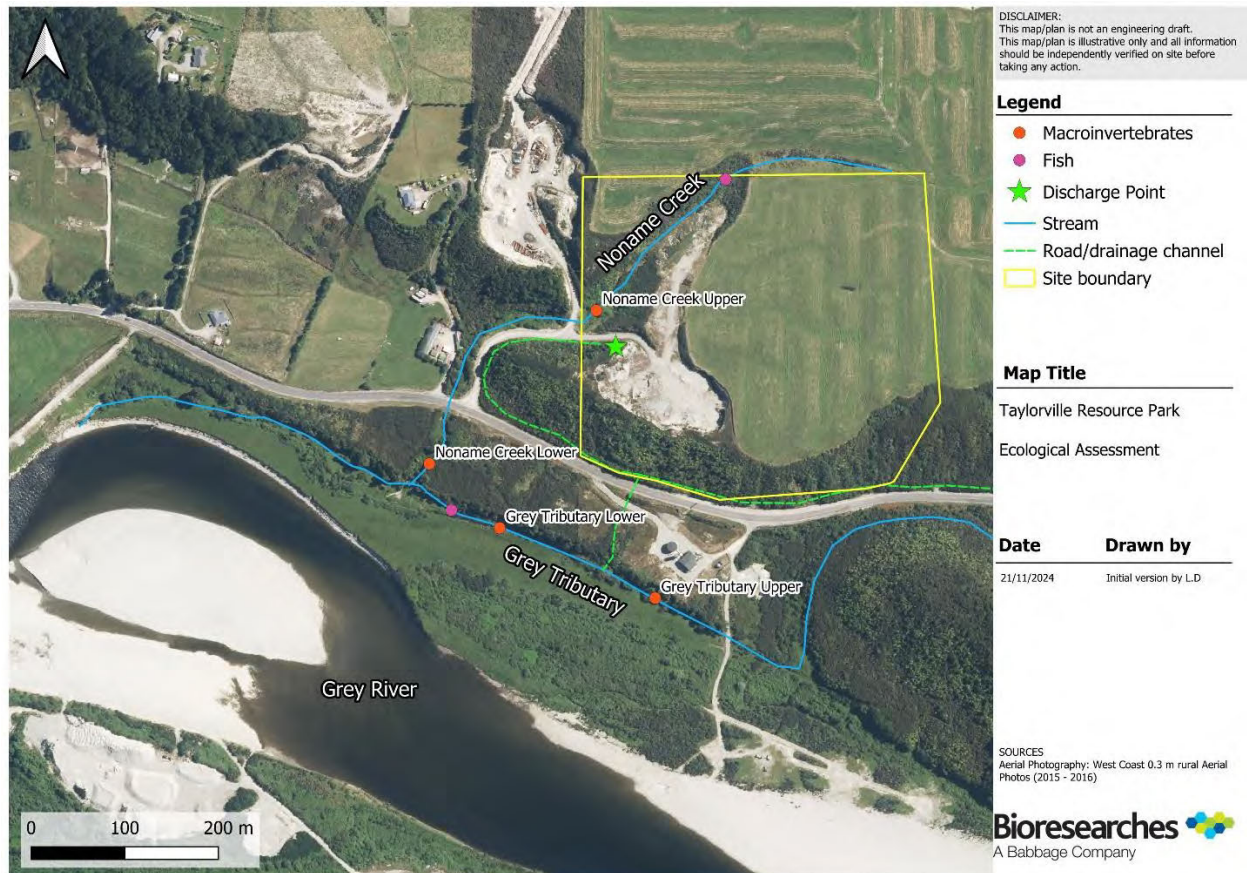


Figure 1. Map of the site and surrounding river systems

Noname Creek

Noname Creek was considered to be of **low** ecological value.

The creek originates on the northern boundary of the site and flows in a west to south-west direction for approximately 590 m before entering a culvert under Taylorville Road. It then continues for a further 91 m before discharging into the Grey Tributary.

Noname Creek consisted of a narrow (average 0.40 m), shallow (average 0.10 m) waterway. The substrate is primarily hard bottomed, dominated by gravels, cobbles and boulders, although occasional patches of soft sediment are present. When disturbed, sediment plumes were present within the upper portion, above the Taylorville Road culvert. Hydrological heterogeneity throughout the reach was good, and consisted of fast and slow runs, riffles and pools. No natural inland wetlands were associated with Noname Creek.

Riparian vegetation was variable, with the upper Noname Creek consisting of predominantly exotic vegetation and pasture. The lower portion of Noname Creek, below the culvert was largely native with patē, pigeon wood (*Hedycarya arborea*), kanono (*Coprosma grandifolia*), kiokio (*Parablechnum novae-zelandiae*), blackberry, and kōtukutuku (*Fuchsia excorticata*). With the exception of the upper 150 m of Noname Creek, the waterway was well vegetated with good shading and bank stability. The upper portion of Noname Creek, present along the site boundary, was less vegetation and predominantly contained gorse, blackberry, and willow (*Salix* sp) within the riparian yard. Despite this, bank stability appeared to be

good and riparian yard functions such as overhanging vegetation and organic matter inputs were abundant.



Photo 1. Noname Upper



Photo 2. Noname Upper

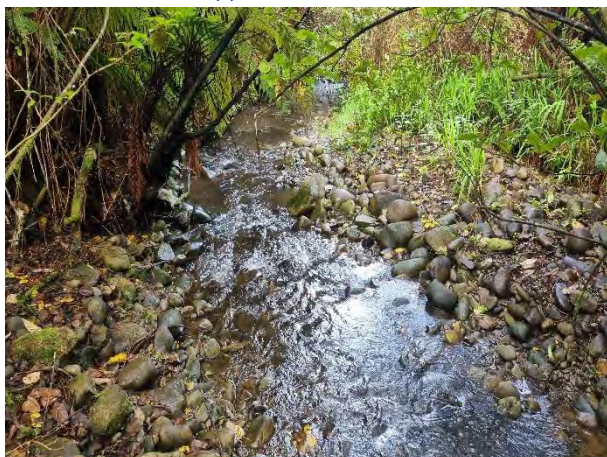


Photo 3. Noname Lower

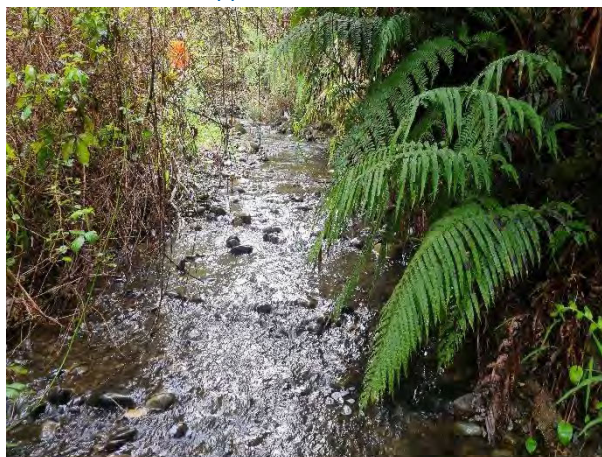


Photo 4. Noname Lower

Aquatic habitat abundance and diversity in Noname Creek was good, with riffles, pools, runs and overhanging vegetation present throughout. It is likely the elevation and culvert separating the upper and lower reaches, which may act as a partial barrier to fish passage. In the absence of fish barriers, it is expected bully species (*Gobiomorphus* sp) and eels could be present.

Table 1. Ecological value and justification of Noname Creek under assessment matters of the EIANZ guidelines.

Matters	Value and Justification
Representativeness	Low Modified water body on the upper reaches through agricultural land use. More naturalised downstream reach. Water flow shallow with a run-riffle-run sequences over cobbles and gravels.
Rarity/distinctiveness	Low May support threatened species on the lower extent, low likelihood of fish species colonising upper reach. Typical for stream within the West Coast region.
Diversity and pattern	Low Good variety of in-stream habitats, however shallow in depth. Available habitat is not complex and limited, and is likely to only provide habitat for pollutant tolerant macroinvertebrates. culvert creates at least a partial barrier to fish passage, limiting is

Ecological context	Moderate First order stream with permanent hard substrates throughout and loading of fine sediments from surrounding agricultural land use. Lower riparian yard indigenous, upper riparian yard exotic.
Overall Ecological Value	Low

Grey Tributary

The Grey Tributary was considered to be of **low** ecological value.

The Grey Tributary is located south of the site and classified as a permanent stream. The proposed discharge will eventually enter the Grey Tributary via a roadside drain that collects water from the site and adjoining roadside drains. From this drain connection point, water flows for an additional 615 m in the tributary before entering the Grey River.

The Grey Tributary consisted of a wide (8 m – 13 m) flow path which was relatively shallow throughout, with depth ranging between 0.4 m to 0.7 m, excluding deep pools. The flow path was predominantly soft bottomed, with a layer of thick, but firm mud covering the stream bed. This substrate was not considered to be excessively silty or subject to high sediment loading, with the boot depressions in the substrate approximately 0.06 m in depth. Flow through the tributary was slow with a low variation in hydrology, only consisting of slow runs and deep pools. Approximately 300 m downstream of the drain connection to the tributary, the stream channel narrows (<1 m), with flows becoming swifter and the prevalence of stony substrates increased. The channel morphology of the Grey Tributary was variable. The true right bank was relatively steep sloping, and vegetated with a mix of exotic and native vegetation. The true left bank of the stream was flat sloping with high connectivity to the floodplain.

The riparian vegetation on the true right bank was exotic and native, consisting of species such as of willow (*Salix* sp.), kamāhi (*Pterophylla racemosa*), wild broom (*Cytisus scoparius*), karamū (*Coprosma lucida*), round-leaved coprosma (*Coprosma rotundifolia*), kahikatea (*Dacrycarpus dacrydioides*), gorse (*Ulex europaeus*), patē (*Schefflera digita*), blackberry (*Rubus fruticosus*), and pōhuehue (*Muehlenbeckia complexa*). The true left bank supported a riverine wetland is supported on the immediate floodplain, extending 1 m – 2 m upslope, with the hydric vegetation predominantly exotic; soft rush (*Juncus effusus*), buttercup (*Ranunculus* sp.), wet-tolerant grasses, and fan flowered rush (*Juncus sarophorus*). The base of the Grey River supported macrophyte growth with pond weed (*Potamogeton* sp.).

*Photo 5. Grey Tributary Upper**Photo 6. Grey Tributary Upper**Photo 7. Grey Tributary Lower**Photo 8. Grey Tributary Lower*

Aquatic habitat within the Grey River was limited to slow runs, pools, macrophytes, woody debris, overhanging vegetation and floodplain vegetation. The macroinvertebrate and fish communities within the Grey Tributary would likely reflect this. Fish passage from the Grey River to the Grey Tributary is limited due to the topography of the two waterways, limiting the potential diversity of fish species capable of accessing the Grey Tributary, and therefore Noname Creek, to species with climbing abilities.

Table 2. Ecological value and justification of the Grey Tributary under assessment matters of the EIANZ guidelines.

Matters	Value and Justification
Representativeness	Low Permanent stream with wetland habitats on margins. Water slow flowing and predominantly soft bottomed, lacking diverse habitats, and hydrological variation.
Rarity/distinctiveness	Moderate Likely to support at risk species, such as longfin eel. Soft bottomed and silty substrates reduce the quality and complexity of aquatic habitats. Exotic wetland present on the margins and floodplain.
Diversity and pattern	Low Likely low diversity of macroinvertebrate and fish communities due to low diversity and complexity of aquatic habitats. Low diversity and complexity of aquatic habitats

Ecological context	Low Second order stream with low diversity and abundance on in-stream habitats, predominantly being macrophytes, rushes and woody debris. Indigenous and exotic, pasture riparian yard limiting functions such as filtration and shade.
Overall Ecological Value	Low

Roadside Drain

The roadside drain was considered to be of **negligible** ecological value

The roadside drain meets the definition of a 'drain' under the Regional Land and Water Plan (2014), as it is an artificial channel, likely constructed to collect and transport water from the roading. The drain is absent in historic aerial images, and does not incorporate any natural bed, as no river or creek is present in the location of the roadside drain in historic aerials.

The drain, at the point of TCD1 discharge (Section 1), consists of a narrow, shallow drain within 2 m of the gravel road. The base of the drain appears to be bedrock with areas of filamentous algae and water cress (*Nasturtium officinale*) within the channel. Further downstream, where the drain runs eastwards (Section 2) adjacent to Taylorville Road, patches of budding-club rush (*Isolepis prolifera*) and duckweed (*Lemna minor*) was observed. As with Section 1, water depth was overtly shallow, and the channel is located within 2 m of the road and narrow. Section 3 of the drain, which runs westwards from the spring, is also situated close to the road, has very shallow water depth and the bed is covered in fine silts and soft sediments. Overall, the drain (Section 1, Section 2 and Section 3), contains very little hydrological heterogeneity, with shallow depth, and very little to stagnant flow.

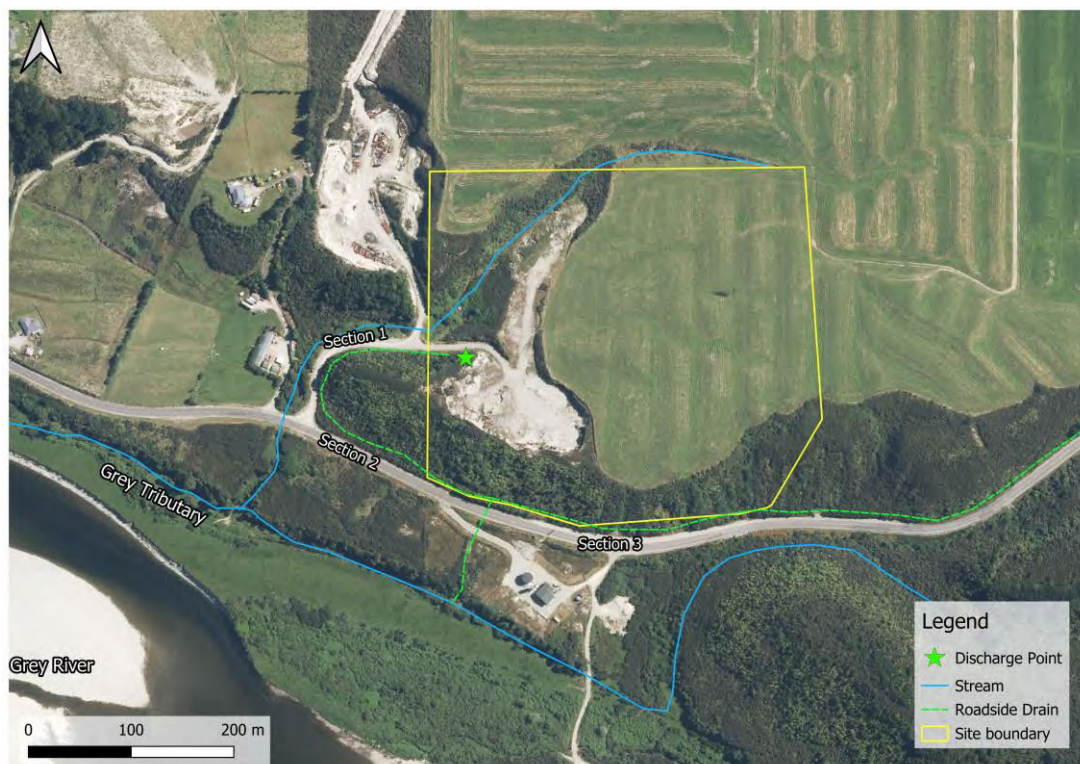


Figure 2. Map of the roadside drain and descriptive sections.

The roadside drain receives water from multiple sources, including daylighted groundwater, overland flow paths, and surface water run-off from the surrounding landscape and roading. Water quality analysis undertaken by Engeo² show elevated levels of nitrate-N and conductivity throughout the drain, independent of the site operations.

The drain lacks complex habitats, such as woody debris, cobbles/gravels and interstitial spaces, overhanging vegetation and due to the poor water quality, the macroinvertebrate populations are expected to be of low diversity and abundance, and would consist of pollutant tolerant taxa throughout both Drain 1 and Drain 2. Barriers to fish passage are present where the drain discharges into the Grey Tributary. Native fish populations are not expected to be present in the drain.

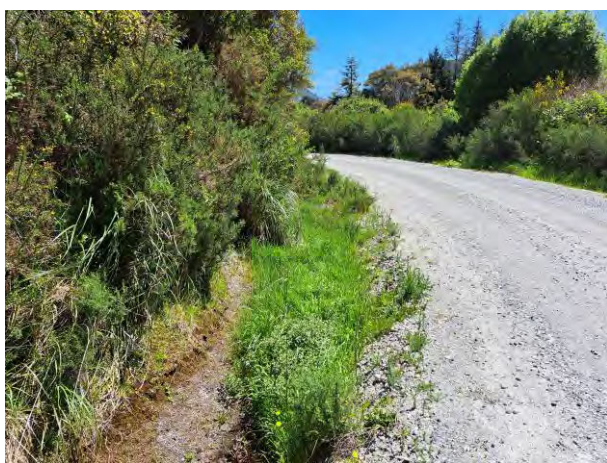


Photo 9. Section 1



Photo 10. Section 2



Photo 11. Section 3



Photo 12. Discharge point and polishing area

Table 3. Ecological value and justification of the roadside drain under assessment matters of the EIANZ guidelines.

Matters	Value and Justification
Representativeness	Very Low Waterbody is an artificial drain constructed to receive water run-off from surrounding roading. Water still standing or hosts trickle flow, with macrophytes and long filamentous periphyton growing in channel. Minimal aquatic habitat.

² Engeo (2024). Review of PDP (11-Oct-24) "Impacts of Taylorville Resource Park Limited Landfill on Grey District Council Coal Creek Water Supply Intake". Memo date: 6 November 2024.

Rarity/distinctiveness	Very Low Roadside drain is highly unlikely to support threatened fauna as it is not a natural waterway and has very limited connectivity to the wider catchment.
Diversity and pattern	Very Low Drain channel is not a natural waterway and consists of a uniform channel width and depth. Available habitat is not complex and limited, and is likely to only provide habitat for pollutant tolerant macroinvertebrates.
Ecological context	Very Low Channel is an artificial roadside drain, lacking aquatic habitat and diversity with the exception of macrophyte growth. Receives water from multiple sources, including containments from agricultural and roading, reducing water quality.
Overall Ecological Value	Negligible

Ecological Impact Assessment

The discharge of water from the site (both temporary to land (RC-2024-0122) and via the hummus filter (RC2021.0097)), has the potential to alter the water quality and chemistry within the roadside drain. PDP has acknowledged leachate concentrations are significantly diluted as water moves away from the landfill, and will be below the Drinking Water Standards health-based MAV criteria, therefore no changes in water quality to the Grey Tributary should occur as a result of the discharge to land.

The discharge of water via the roadside drain, which will enter the Grey Tributary ~100 m downstream of the discharge. Water discharged from the TRP is captured via a roadside drain and diverted away from Noname Creek. As such, no future treated leachate or contaminant originating from TRP should enter Noname Creek.

In regards to the roadside drain, hydraulic modelling undertaken by Engeo³ indicates any contaminants will be generally diluted and dispersed through an aquifer to negligible concentrations prior to discharge and interception within the roadside drain. As such, the discharge of water is unlikely to result in significant, or noticeable changes to the ecological values of the drain as a result of altered water quality and water chemistry. Considering the dilution factor within the drain, no excess in contaminants as a result of the TRP discharge should enter the Grey Tributary.

Whilst site operations may have a minor influence on the water quality within the roadside drains, multiple sources of water enter the channel which are heavily influenced by surrounding land use, independent of TRP. Existing land-use includes agricultural/rural, including aerial spraying of fertiliser, and vehicle use and the associated roading run-off from Taylorville Road. As such, influences to water quality in the roadside cannot be solely attributed to the operations within the site.

The discharge of water to land, which will then enter the roadside drain will not result in the displacement of indigenous fauna (macroinvertebrates and fish), as any organisms, if present, in the drain will be pollutant tolerant macroinvertebrate taxa.

³ Engeo (2024). Preliminary Hydrogeological Assessment: 287 Taylorville Road, Coal Creek, Greymouth.

The discharge to the roadside drain was considered to be of **low** magnitude and overall **very low** level of effect to the roadside drain and Grey Tributary as it will have a very slight shift from the existing baseline conditions, and negligible impact on ecological values and communities of the drain. There should be no effect to Noname Creek as discharged water should not enter this system.

Regards,

Laura Drummond, M.Sc. (Hons), MEIANZ
Coastal & Freshwater Ecologist

M +64 27 254 9685 |

Bioresearches 
A Babbage Company

Babbage Consultants Limited

Appendix A. Ecological Impact Assessment Methodology

Table 4. Criteria for assigning value to habitat/species for assessment.

Value	Determining Factors
Very High	Nationally Threatened species found in the 'zone of influence' (ZOI) either permanently or seasonally. Area rates 'High' for at least three of the assessment matters of Representativeness, Rarity/distinctiveness, Diversity and Pattern, and Ecological Context. Likely to be nationally important and recognised as such.
High	Species listed as At Risk – Declining found in the ZOI either permanently or seasonally. Area rates 'High' for two of the assessment matters, and 'Moderate' and 'Low' for the remainder OR area rates 'High' for one of the assessment matters and 'Moderate' for the remainder. Likely to be regionally significant and recognised as such.
Moderate	Species listed as At Risk – Relict, Naturally Uncommon, Recovering found in the ZOI either permanently or seasonally. Locally uncommon or distinctive species. Area rates 'High' for one of the assessment matters, 'Moderate' or 'Low' for the remainder OR area rates as 'Moderate' for at least two of the assessment matters and 'Low' or 'Very Low' for the remainder. Likely to be important at the level of the Ecological District.
Low	Nationally and locally common indigenous species. Area rates 'Low' or 'Very Low' for majority of assessment matters, and 'Moderate' for one. Limited ecological value other than as local habitat for tolerant native species.
Negligible	Exotic species including pests, species having recreational value. Area rates 'Very Low' for three assessment matters and 'Moderate', 'Low' or 'Very Low' for the remainder.

Table 5. Criteria for describing the magnitude of effects (EIANZ 2018)

Magnitude	Description
Very High	Total loss of, or a very major alteration to, key elements/features of the existing baseline conditions, such that the post-development character, composition and/or attributes will be fundamentally changed and may be lost from the site altogether; AND/OR Loss of a very high proportion of the known population or range of the element/feature.
High	Major loss of major alteration to key elements/features of the existing baseline conditions such that the post-development character, composition and/or attributes will be fundamentally changed; AND/OR Loss of a high proportion of the known population or range of the element/feature.
Moderate	Loss or alteration to one or more key elements/features of the existing baseline conditions, such that the post-development character, composition and/or attributes will be partially changed; AND/OR Loss of a moderate proportion of the known population or range of the element/feature.

Low	Minor shift away from existing baseline conditions. Change arising from the loss/alteration will be discernible, but underlying character, composition and/or attributes of the existing baseline condition will be similar to pre-development circumstances and patterns; AND/OR Having minor effect on the known population or range of the element/feature.
Negligible	Very slight change from the existing baseline condition. Change barely distinguishable, approximating to the 'no change' situation; AND/OR Having negligible effect on the known population or range of the element/feature.

Table 6. Criteria for describing the level of effects (EIANZ 2018). Where text is italicised it indicates 'significant effects' where mitigation is required.

Magnitude of Effect	Ecological Value				
	Very High	High	Moderate	Low	Negligible
Very High	<i>Very High</i>	<i>Very High</i>	<i>High</i>	<i>Moderate</i>	Low
High	<i>Very High</i>	<i>Very High</i>	<i>Moderate</i>	Low	Very Low
Moderate	<i>High</i>	<i>High</i>	<i>Moderate</i>	Low	Very Low
Low	<i>Moderate</i>	Low	Low	Very Low	Very Low
Negligible	Low	Very Low	Very Low	Very Low	Very Low
Positive	Net Gain	Net Gain	Net Gain	Net Gain	Net Gain



memorandum



TO	Michael Durand	FROM	Carl Steffens
	Planz Consultants	DATE	24 February 2025
RE	Technical Review – TRPL short-term consent application to discharge contaminants to ground – RFI Response		

West Coast Regional Council (WCRC) sent Taylorville Regional Park Limited (TRPL) a Request for Further Information (RFI) dated 19 December 2024 relating to the TRPL short-term consent application to discharge contaminants to ground at Taylorville Regional Park (TRP) landfill. On behalf of TRPL, ENGEO has provided a response to the RFI via a memo dated 30 January 2025. The response provided includes an additional separate memo prepared by Bioresarches that describes an ecological impact assessment (a request made in the WCRC RFI). Pattle Delamore Partners Limited (PDP) has been engaged by WCRC to carry out a technical review of the RFI response documents. Our review is detailed in this memo.

There are seven individual questions in the WCRC RFI. Each of these is set out below (in bold) followed by comment on whether we consider the additional information provided by ENGEO or Bioresarches sufficiently addresses the question posed.

- 1. Hydraulic conductivity estimates from onsite testing (pump testing and/or slug testing). If slug testing is used, then please provide testing data from at least 3 bore locations on the upper terrace surface.**

ENGEO report they have carried out slug testing at three onsite groundwater monitoring bores (MB13, MB14 and MB19d). Based on analysis of the slug testing using the methodology of Bouwer and Rice (1976), ENGEO report hydraulic conductivity (K) values for the Pleistocene gravels ranging from 1.3×10^{-6} to 1.11×10^{-5} m/s (0.11 to 0.96 m/d).

We note the use of bore MB19d for the testing is not considered ideal given the small screen interval and the location of much of the screen over the deep mudstone (although the filter pack around the screen extends into the gravelly strata). In addition, this bore is significantly upgradient of the proposed discharge location and we consider downgradient bores are more relevant to the proposal. We also note it is somewhat surprising that the analysis results from MB19d gave the highest values of K given the screen location in the mudstone.

Despite the comments above, overall it is considered that the information provided has sufficiently addressed the first request of the RFI.

- 2. Updated mounding and contaminant modelling simulations using the hydraulic conductivity estimates derived from the requested testing. Based on the testing results, please take into account that the results from testing (particularly slug testing) may provide local values for hydraulic conductivity but not necessarily be representative of the average hydraulic**

conductivity over the area of the landfill. Therefore, please provide multiple modelling simulations informed by the testing results using a realistic range of potential hydraulic conductivity values for the strata beneath the upper terrace surface (i.e. a sensitivity assessment based on varying the hydraulic conductivity that is informed by the testing results).

Mounding

ENGEO has carried out additional mounding calculations using a K value of 0.86 m/day. However, we note this value is at the higher range of values estimated from their slug testing analysis and is also significantly higher than the values derived from bore MB13 (values of 0.13 to 0.16 m/day), which is in close proximity of the proposed discharge location. Lower values of K will result in larger mounding and therefore the values estimated from testing in MB13 would result in higher mounding predictions. Given the location of the proposed discharge the K values from this bore are potentially more relevant. We also note the updated mounding assessment is based on a significantly larger infiltration area (20 m x 20 m instead of 10 m x 10 m in the original application) which will result in a very large excavation onsite.

ENGEO has not reduced the discharge volume for the new assessment (still 800 m³/day), although in their response they state the anticipated infiltration capacity is on the order of 200 to 300 m³/day. Therefore, the updated assessment appears to be inconsistent with the expected infiltration rate. The updated assessment predicts maximum mounding of over 8 m after a day of discharge, compared to the much lower mounding prediction of around 0.6 m in the original application. In addition, we note that our own calculations show the predicted mounding will continue to increase significantly for prolonged infiltration periods.

In late 2024, ENGEO indicated that an initial volume of 5,500 m³ of water needs to be removed from the onsite retention pond, although we assume that volume may have increased since that time. Assuming this volume is still valid, then 6.9 days of continuous discharge would be required at a rate of 800 m³/day to remove 5,500 m³ from storage. Our calculations indicate this situation would result in over 20 m of mounding at the end of the discharge period. If the discharge rate is reduced to 200 m³/day (in line with the infiltration capacity suggested by ENGEO) the mounding after 1 day reduces to around 4.5 m. However, to discharge 5,500 m³ would require 27.5 days and result in around 12 m of mounding. Furthermore, if K is reduced to 0.13 m/day (as per the slug testing results from MB13), mounding based on discharging 200 m³ in a day is over 9 m, while discharging that volume of water for 27.5 days results in predicted mounding of over 50 m.

Based on the low values of K estimated from the ENGEO slug testing, we have concerns that discharging the proposed quantities of water to ground will not be feasible, because the calculated mound heights are similar (some short-term discharge scenarios) or significantly greater (prolonged discharge scenarios) than the reported depth to water, which was measured at 7.9 m below ground level in monitoring bore MB13 (located closest to the proposed infiltration basin) at the time of the slug testing. Even for the situation where the proposed discharge rate is significantly reduced, it appears mounding could still be prohibitive, accentuated by the prolonged discharges that will be required. Further to the potential difficulties with disposing of the proposed quantities of groundwater to ground at the site, we assume that mounding of the scale predicted could potentially cause stability concerns (including within the landfill) on the terrace.

While ENGEO has addressed the request for additional mounding assessment, the results indicate that the discharge to ground may not be feasible.

Contaminant Modelling

ENGEO has provided updated RISC5 contaminant modelling using high and low K values estimated from their recent slug testing. Based on the information provided with the original short-term consent application, we understand the RISC5 modelling is based on a continuous discharge for 6 months. It isn't clear what discharge rate is used for the modelling, but we assume it is 800 m³/day, as per the mounding assessments that have been provided. A groundwater gradient of 0.0018 m/m was assumed for the original modelling along with a K value of 84 m/day.

The updated modelling uses much lower values of K (0.11 m/day and 0.86 m/day) based on the slug testing results. We assumed the most recent modelling is also based on a volume of 800 m³/day discharged continuously for 6 months. We note the outputs from the updated modelling show an assumed gradient of 0.018 m/m compared to 0.0018 m/m, but no reasoning is provided for this change.

The modelling results reported show that only aluminium and Nitrate-N have detectable concentrations at the inferred point of measurement, 310 m from the discharge location. However, for both parameters, the modelled concentrations have significantly reduced by around 3 to 4 orders of magnitude compared to the modelled source water at the point of discharge. The high hydraulic conductivity scenario indicates that peak concentrations at 310 m from the discharge occur around 13 years from the time of discharge, while the low hydraulic conductivity scenario takes around 100 years, or longer, for peak concentrations to occur. We note that the modelled peak concentrations are around an order of magnitude higher for the high hydraulic conductivity scenario compared to the low hydraulic conductivity scenario. This is generally at odds with what we would expect, as more dilution and dispersion would typically be expected to occur under a higher conductivity scenario.

We have concerns regarding the validity of the modelling results. It is stated by ENGEO that the modelling assumes no degradation along the flowpath and that contaminants will generally be attenuated through dilution and dispersion in the unconfined aquifer. However, we note that based on the range of K values estimated from the slug testing, the throughflow of groundwater beneath the terrace is very low. Assuming that groundwater flow occurs along a width of around 100 m (in reality we expect the width of aquifer impacting on the flow of contaminants is likely to be less) and the average thickness of groundwater in the gravel strata above the mudstone is 10 m, the cross-sectional area of groundwater flow beneath the terrace is only 1,000 m². Based on the highest modelled K value of 0.86 m/day and hydraulic gradient of 0.0018 m/m, the aquifer throughflow is only 1.55 m³/day. For the lowest derived slug testing K value of 0.11 m/day, the throughflow is only 0.2 m³/day. For a discharge of 800 m³/day (considered unlikely to be feasible based on the mounding results), the potential for dilution of contaminant concentrations is negligible which seems inconsistent with the modelling results.

We also note that the modelling results predict much higher concentrations of Nitrate N in the output compared to Ammonia N (no concentration provided). This is despite the source concentration of Ammonia N (98 mg/L) being much greater than the source concentration of Nitrate N (6.5 mg/L) and the statement by ENGEO that the modelling assumes no degradation of contaminants.

Based on the low potential for dilution (but noting we have not carried out our own detailed modelling), we consider the modelling provided may not be reliable. To address these concerns, we would need either further explanation from ENGEO, or to carry out our own modelling to justify/confirm the large decrease in contaminant concentrations shown by the ENGEO modelling.

In addition to our concerns around the modelling, we consider it is possible these results may not represent the actual discharge that could occur, as the results of the mounding assessments indicate discharging large quantities of water to ground at the site may not be achievable.

3. Updated contaminant modelling simulations using input parameters more consistent with high parameter concentrations that have been measured in leachate and wastewater.

As per the previous section, we have concerns around the validity of the modelling results and further information from ENGEO or additional modelling by PDP would be required to address our concerns.

4. An ecological survey of the surface water features and aquatic habitat surrounding the site, including Noname Creek, the drain along Taylorville Road, including that part of the drain that begins at spring seep TCD1 and the streams on the southern side of Taylorville Road (excluding the Grey River) and an impact assessment on surface water quality and ecological values resulting from the proposed discharge.

An ecological impact assessment has been carried out by Bioresarches and summarised in a memo dated 7 February 2025. Bioresarches consider the ecological value of NoName Creek, the roadside drain originating at TCD1 and the Grey Tributary on the lower terrace surface below Taylorville Road is low based on the Ecological Impact Assessment (EclA) Guidelines for use in New Zealand (Roper-Linkday et al., 2017).

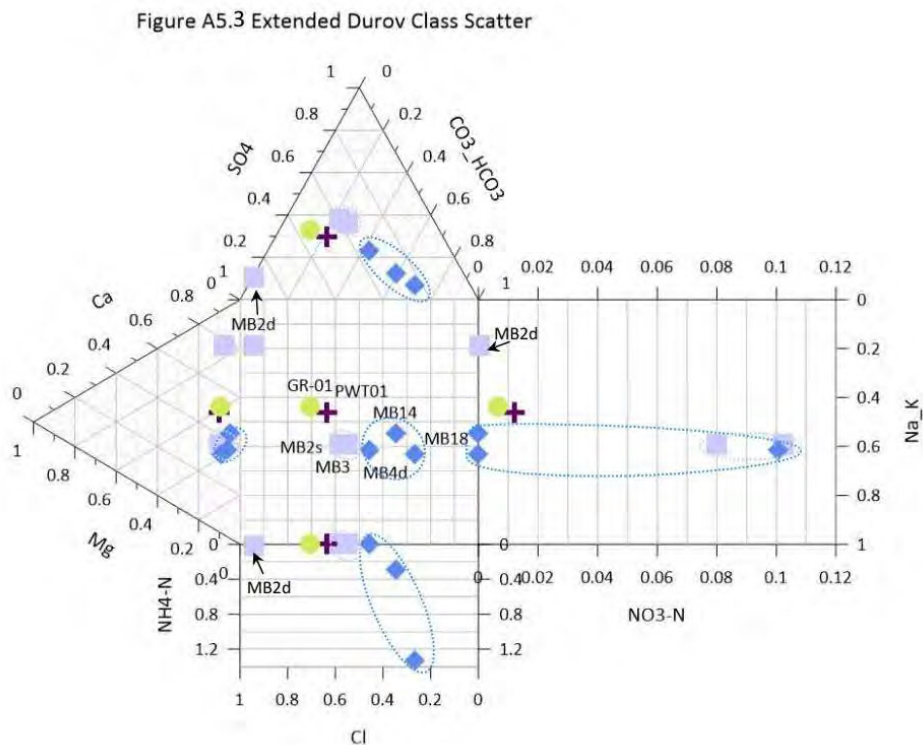
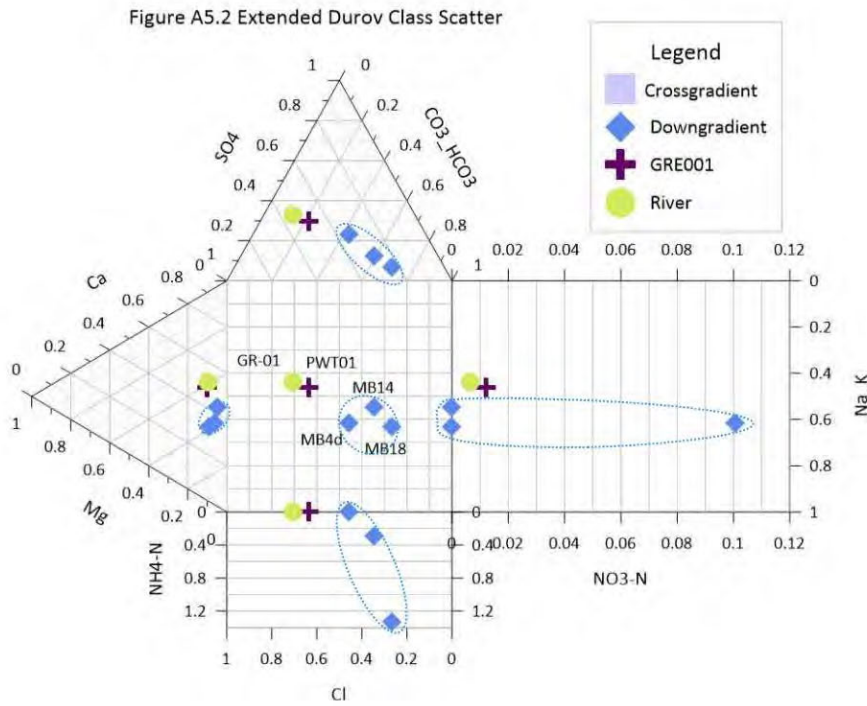
A review by PDP's ecology team would be required to confirm the results of the ecological assessment, however at this stage we have not carried out that review because we consider other effects assessments have not been adequately addressed. Our current view is that discharge to ground is unlikely to be feasible due to the low K soils underlying the terrace and the associated large mounding effects that are predicted to occur. In addition, we don't consider we have sufficient information to confirm the potential contaminant concentrations that may reach the surface water bodies in the area. We consider these two issues should be resolved before we review the ecological assessment, however our view is that the currently available information, presented in response to question 1, indicates that the proposed discharge to ground is unlikely to be feasible.

Regarding contaminant concentrations, if it could be confirmed that ecological receiving water standards wouldn't be breached during discharge of the contaminants to ground, then the results of the ecological assessment are less critical. We do not think that the current information is sufficient to reach that conclusion, but if further work shows that to be the case, then the ecological work that has been carried out may be sufficient.

5. A water quality comparison between water samples collected from the GDC water supply, the Grey River, groundwater upgradient of the landfill north of Taylorville Road and surface water samples collected at TRD2 and TRD4. The samples should include analysis for anions and cations and nitrogen species, in addition to other parameters sampled to date by TRPL. The purpose of the sampling is to inform whether the GDC water supply source is dominantly Grey River water and there is minimal input to the supply bores from upgradient groundwater and surface drainage.

ENGEO collected a water quality sample from the raw (untreated) water at the GDC treatment plant (sample PWT01), in addition to a sample from the Grey River upstream of the treatment plant and GDC supply bores (sample GR-01). The samples were analysed for a number of chemical parameters, including general water quality parameters, nitrogen species, dissolved metals and ions in solution.

The results for ions were compared to October 2024 results from samples collected from monitoring bores MB2s, MB2d, MB3, MB4d, MB14 and MB18 and plotted on an extended Durov diagram (Figures A5.2 and A5.3 of the RFI response). These figures are shown below. ENGEO notes that bores MB2s, MB2d and MB3 are cross gradient bores, while MB4d, MB14 and MB18 are downgradient bores with respect to the waste storage areas.



As reported by ENGEO, samples PWT01 and GR-01 plot relatively closely together. This indicates similar water composition between Grey River water and the water in the GDC treatment plant, although they are not exactly the same. The Durov diagrams indicate some deviation in ion chemistry between downgradient groundwater samples and those collected from the Grey River and GDC treatment plant, however the treatment plant sample plots very slightly closer to the downgradient bores compared to the Grey River sample. There is more similarity between samples PWT01 and GR-01 and the cross-gradient bore samples compared with those from down-gradient locations, with the exception of sample MB2d, although the results from this bore may be affected by the screen being in the underlying mudstone beneath the terrace. Overall, based on the ion information provided by ENGEO, there is a relatively good match between raw GDC treatment plant water and Grey River water, although the GDC raw water does suggest some influence from the water chemistry in the groundwater monitoring bores on the terrace above Taylorville Road.

With regard to the results for other parameters, we note the sampling did not include analysis for all parameters previously sampled by TRPL as requested in the RFI. For example, samples PWT01 and GR-01 did not include analysis for dissolved aluminium, iron, manganese or DRP. Ideally, the samples at PWT01 and GR-01 would have been compared with new groundwater samples collected at the same time (rather than comparison with older samples) and we also note that comparison has not been provided with TRD2 and TRD4 as requested.

Furthermore, we note that concentrations of copper, lead and zinc in treatment plant sample PWT01 are unexpectedly high and exceed ecological guideline criteria, whereas Grey River sample GR-01 recorded all dissolved metals at concentrations below the laboratory detection level. Together with the slight deviation of sample PWT01 towards the groundwater bore samples, this potentially suggests an influence on water chemistry in the supply bores from the groundwater north of the river.

With regard to copper (0.193 mg/L) and lead (0.006 mg/L), the concentrations measured in PWT01 are particularly high compared to all other water samples presented in the applicant's information. Both of these parameters were measured in PWT01 at concentrations well above those measured by TRPL in upgradient groundwater and surface water. Furthermore, the dissolved copper and lead concentrations in PWT01 are also higher than those measured in any of the samples collected from wastewater or leachate to date at the landfill (highest concentration in leachate or wastewater of 0.058 mg/L for copper and 0.0024 mg/L for lead), although presumably these concentrations will vary depending on the different wastes that are being received at any particular time. While nickel and zinc were also detected/elevated in PWT01, higher (nickel) or very similar (zinc) concentrations have been measured previously by TRPL in upgradient groundwater bores.

The groundwater quality results from PWT01 and GR-01 potentially indicate that groundwater north of the Grey River may influence the chemistry of water taken for the GDC community water supply. This is despite the supply bores being situated immediately adjacent to the Grey River. We consider this is a significant finding and assuming these results are not due to some other factor (such as potential contamination within the treatment plant), this finding may potentially alter our previous assessment that there is a low risk to the water supply from the operation of the landfill. It would be helpful (and important) for the applicant to provide comment on the potential cause of the elevated concentrations of copper, lead, nickel and zinc in the water treatment plant sample.

6. Yield versus drawdown data from the GDC supply bores to confirm the gravel deposits immediately adjacent to the Grey River are of high permeability with a high hydraulic connection to the Grey River.

ENGEO has provided driller's logs for GDC bores BH01 and BH02 (two of the three GDC community drinking-water supply bores) and data from step-drawdown and constant rate pump testing in BH02. We note the McMillan Drilling report provided indicates that BH01 has been decommissioned and therefore we assume that the drilling and testing in BH02 was carried out to replace BH01.

The driller's log for BH02 describes 0.2 m of fill underlain by sand to 1.3 m below ground level (bgl) and gravel dominated strata (sandy gravel and gravel) thereafter to 8.8 m bgl. Silty clay hard mudstone was observed from 8.8 m to the base of the borehole at 10.5 m bgl. BH02 is screened between approximately 6 and 8.5 m bgl and the static water level in the bore following drilling was measured at 4.23 m bgl. In contrast, the driller's log for BH01 describes fill to 0.4 m bgl, clay to 0.8 m bgl, underlain by sandy gravel to 2.9 m bgl. A thick deposit of coarse sand (with gravels and cobbles) was encountered between 2.9 m and 8 m bgl, underlain by sandy gravel to 8.6 m and clayey gravel to 8.7 m bgl. Silty clay mudstone was encountered thereafter to the base of the borehole at around 10.8 m bgl. No bore construction details are provided for BH01, however based on the logs provided it is clear that large thicknesses of relatively fine-grained strata (sand and clay) was observed during drilling of this bore, which is likely to have negatively impacted on the potential yield that could be achieved.

Raw water level observations are also provided from BH02 during step-drawdown testing and constant rate testing in the bore, as well as pumping rate and sand ingress measurements. In addition, plots of these measurements are also provided, along with analyses of the step-drawdown testing data with associated estimates of K and transmissivity (T) (provided by McMillan Drilling). Based on those analyses, ENGEO report that values for K of 3.42×10^{-2} m/s (2,955 m/day) and 1.58×10^{-2} m/s (1,365 m/day) were obtained. We have reviewed the analyses and consider these are not accurate estimates. However, it is clear that pumping from BH02 induces only low levels of drawdown in the bore. Measured drawdown of up to 1 m was measured in the bore at a flow rate of up to approximately 70 L/s.

It would be possible to carry out a more accurate analysis of the pumping test data provided to better estimate K and T, however we don't consider this is necessary as it is clear that operation of BH02 induces very little drawdown in the bore. Drawdown outside the bore casing can be expected to be less than that measured during the testing as additional well losses from turbulent head losses will occur within the bore casing. This information would typically suggest that operation of the GDC bores draws very large volumes of water directly from the Grey River. However, the elevated metals concentrations in sample PWT01 suggests there is still some other significant influence on the chemistry of water abstracted for the GDC water supply. As indicated in the previous section, we consider it will be important for the applicant to provide comment on the potential cause of the elevated concentrations of copper, lead, nickel and zinc in the water treatment plant sample.

7. XRD sample analysis from the Kaiata mudstone underlying the landfill to confirm clay mineralogy in the mudstone and confirm ENGEO comments that swelling within any fractures that may be present in the mudstone will block the transmission of groundwater via a fracture network.

ENGEO has quoted previous XRD analysis (Leaver, 1999) from two Kaiata Mudstone outcrops on the west coast. Based on that data, they state the mineralogy of the Kaiata mudstone comprises

a mixture of kaolinite, illite and vermiculite. They consider this mineralogy comprises a mixture of swelling and non-swelling clay minerals and therefore it is inferred there is negligible flow through defects due to clay swelling. ENGEO do consider there is the potential for preferential pathways through interbedded layers of non-clay materials (silts and sands), but that groundwater that intersects these layers will most likely be transported west-southwest, along the dip of the mudstone bedding planes and away from the GDC water supply.

A review of clay swelling behaviour indicates that kaolinite and illite don't appear to have high swelling ability. The swelling ability of vermiculite may be higher (but still not high), although is also highly dependent on temperature. In addition, there are other properties that affect the swelling ability of these clay minerals (for example pH, salinity and presence of cations) which haven't been discussed by ENGEO. We also note the sample results are from sites distant from the landfill and therefore may not be relevant to local conditions. Furthermore, ENGEO indicate there are layers within the mudstone that may not contain clay minerals, but in those cases groundwater flow will occur along the direction of bedding away from the GDC water supply bores. However, ENGEO has not provided any information to confirm there is no potential for fracture networks at other angles to the dip of the bedding. Based on the information provided, we are currently not satisfied there is no potential for groundwater flow in the direction of the GDC supply bores within fractures within the mudstone.

Limitations

This memorandum has been prepared by Pattle Delamore Partners Limited (PDP) on the basis of information provided by West Coast Regional Council and others (not directly contracted by PDP for the work), including Planz, ENGEO and Biorearches. PDP has not independently verified the provided information and has relied upon it being accurate and sufficient for use by PDP in preparing the memorandum. PDP accepts no responsibility for errors or omissions in, or the currency or sufficiency of, the provided information.

This memorandum has been prepared by PDP on the specific instructions of West Coast Regional Council for the limited purposes described in the memorandum. PDP accepts no liability if the memorandum is used for a different purpose or if it is used or relied on by any other person. Any such use or reliance will be solely at their own risk.

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Prepared by



Carl Steffens

Technical Director – Water Resources

Site Memorandum

Project Name	Taylorville Resource Park		
Memo Topic	RC-2024-0122 and RC-2021-0097-V10 - RFI Response		
Memo Date	5 March 2025	Date of Site Visit	N/A
ENGEO Memo No.	09	ENGEO Project No.	26200.000.001
To	Taylorville Resource Park Ltd	Copy	Barker and Associates
Author	Lyn Nugent	Reviewers	Dave Robotham, James Thurber, Richard Justice

ENGEO has reviewed the memo prepared by PDP (Steffens, C [24 February 2025] to Michael Durand “*RE Technical Review – TRPL short-term consent to discharge contaminants to ground – RFI Response*”). The original queries (in grey) and ENGEO responses (in blue), reflecting comments by PDP, are provided below.

1. Hydraulic conductivity estimates from onsite testing (pump testing and/or slug testing). If slug testing is used, then please provide testing data from at least 3 bore locations on the upper terrace surface

PDP considers that the information provided by ENGEO sufficiently addresses this item of the RFI, with commentary regarding the results. We do note that the range of hydraulic conductivity values (1.05×10^{-5} m/s – 1.3×10^{-6} m/s) generally fall within one standard deviation of the mean, which suggests a small variance (good reproducibility) in the dataset. We further note the hydraulic conductivity values should not be interpreted as a precise measurement; rather, they represent a likely range of conductivity values, of which the most conservative should be applied (refer next item).

2. Updated mounding and contaminant modelling simulations using the hydraulic conductivity estimates derived from the requested testing. Based on the testing results, please take into account that the results from testing (particularly slug testing) may provide local values for hydraulic conductivity but not necessarily be representative of the average hydraulic conductivity over the area of the landfill. Therefore, please provide multiple modelling simulations informed by the testing results using a realistic range of potential hydraulic conductivity values for the strata beneath the upper terrace surface (i.e. a sensitivity assessment based on varying the hydraulic conductivity that is informed by the testing results).

PDP provided a lengthy response to the ENGEO assessment for this item. For brevity, we have consolidated their queries in our response.

Groundwater

PDP indicates that ENGEO applied a discharge rate of 800 m³/d in the RISC model. To be conservative, we did not consider vadose (unsaturated zone) transport in our RISC model; as such, the discharge rate is not an input parameter (we refer to the model input summary in Appendix 3 of the ENGEO memo dated 30 January 2025).

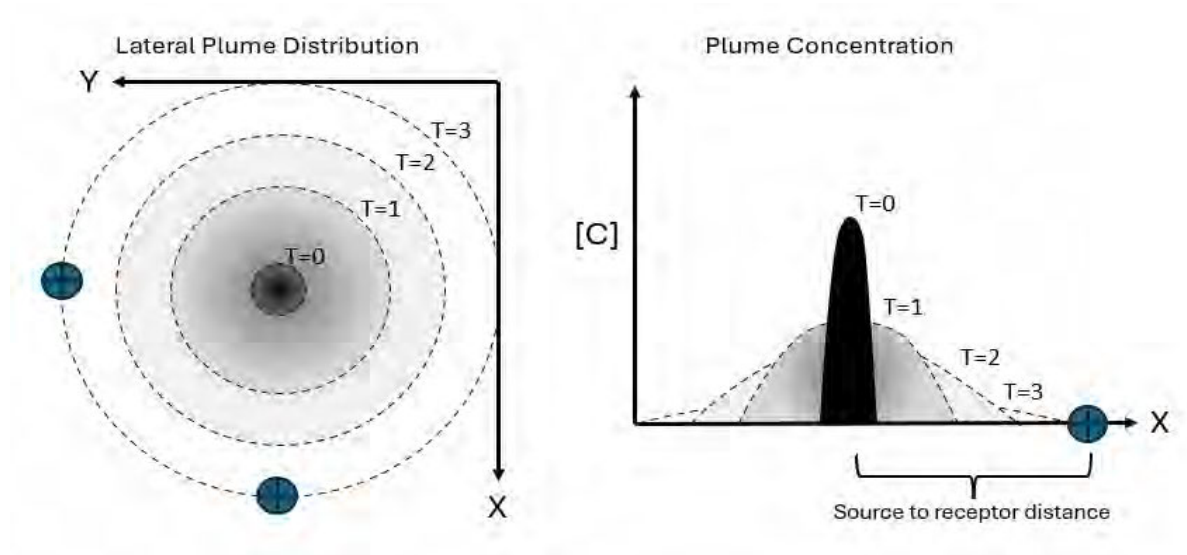
In our model we considered a saturated source with cross-sectional area measuring 80 m² (8 m thickness and 10 m width) and a length of 10 m.

PDP queried the change in inferred hydraulic gradient between the original modelling works (0.0018 m/m; ENGEO, 2024) and the updated modelling (0.018 m/m; January 2025). The hydraulic gradient was revised following the installation of additional monitoring wells in October 2024. This update was provided to WCRC in a memorandum prepared by ENGEO dated 6 November 2024.

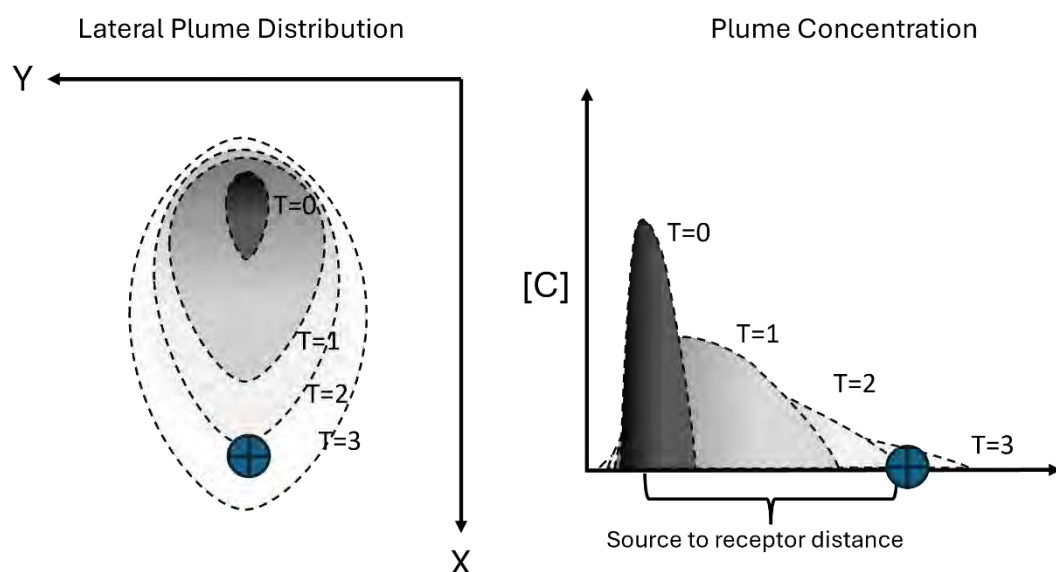
PDP notes that they have not undertaken their own modelling but questions the validity of the ENGEO modelling results based on their expectation that the modelled peak concentrations should be inversely proportional to hydraulic conductivity (i.e., lower peak concentrations with higher hydraulic conductivity).

The movement of a contaminant plume is governed by groundwater flow (i.e., advection and dispersion) and short-range interactions between water molecules (e.g., concentration-dependent diffusion, Brownian motion) (Fetter, 2021). In a scenario with no water flow ($K=0$), a contaminant plume would uniformly migrate away from the source outward as the concentrations equilibrate due to short-range interactions. As such, the concentration (C) will be the same in all directions at a given radius. Figure 1 shown the migration of a plume with $K=0$ at multiple time steps ($T=0$, $T=1$, $T=2$, $T=3$).

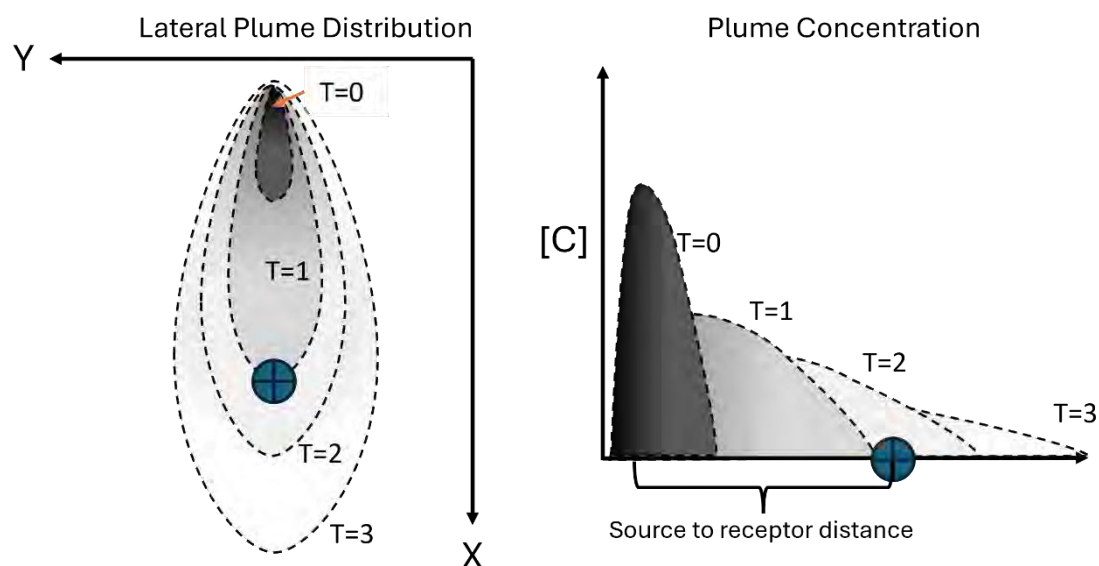
Figure 1: Plume Migration at $K=0$



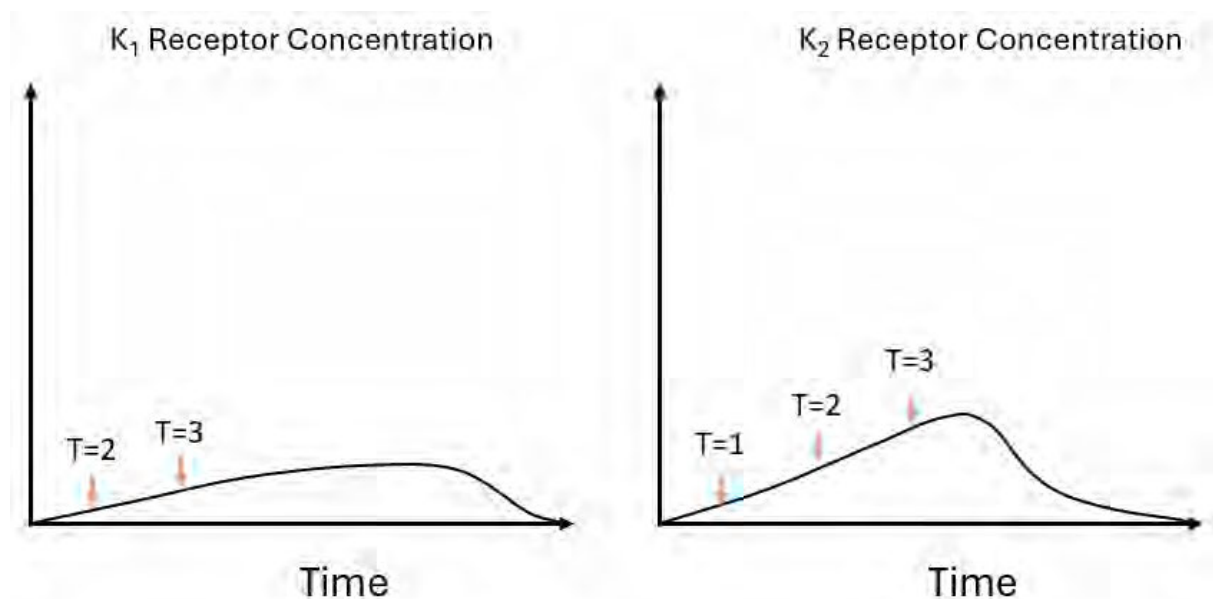
If groundwater flow is introduced and hydraulic conductivity is greater than zero (K_1), the plume will still diffuse due to short-range interactions; however, it will also elongate in the direction of flow (X) due to advection and spread laterally and vertically due to dispersion (refer Figure 2).

Figure 2: Plume Migration at K_1 

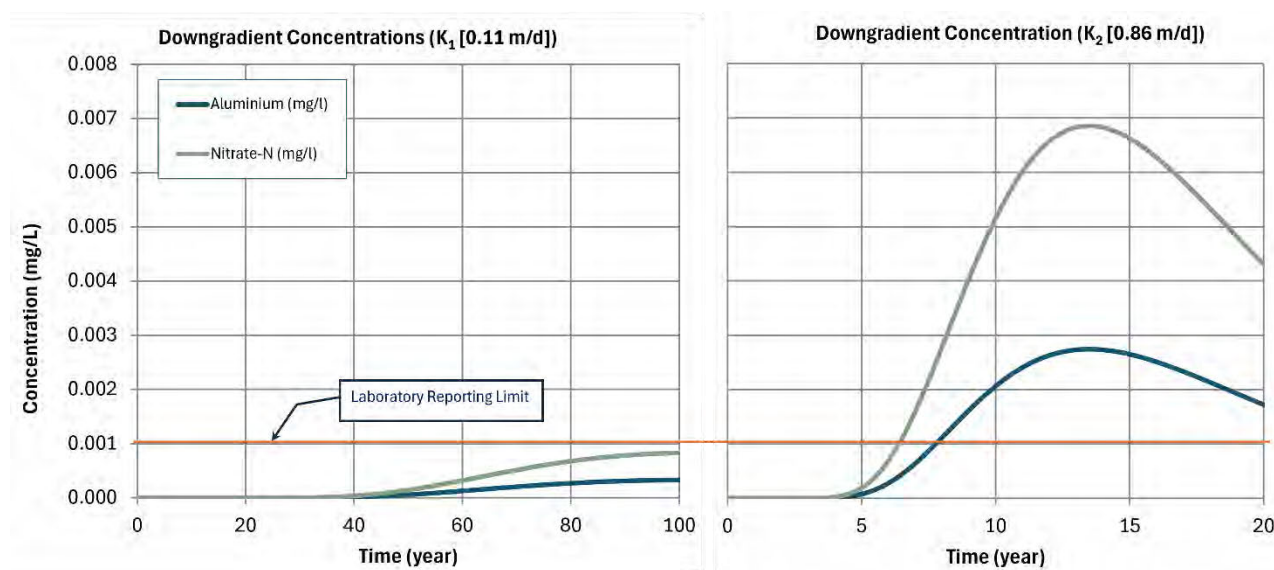
If the hydraulic conductivity is increased ($K_2 > K_1$), the plume will further elongate in the direction of flow (X) and short-range interactions will become less significant (Figure 3).

Figure 3: Plume Migration at k_2 

Note that in all scenarios, the mass of the contaminant in the plume at any time interval is constant. As such, with greater elongation the mass of contaminants in the plume will be more constrained in the lateral (Y) direction and concentrations will be higher along the flow axis. As shown in Figure 4, the downgradient effect will be that, as the hydraulic conductivity increases, the plume will reach the receptor sooner, the maximum concentration will be higher, and the plume will pass more quickly (noting that the source is a 6-month pulse so concentrations will not reach steady state).

Figure 4: Receptor Concentration with Time, where $K_1 < K_2$ 

This corresponds with the results provided in the ENGEO memo dated 30 January 2025 (Figure 5), which showed the concentrations at a point located directly 310 m downgradient of the soakage area.

Figure 5: RISC Modelled Downgradient Concentration at K₁ and K₂

Note: Concentration graphs have been excerpted and modified for relevance.

As demonstrated, when all other inputs (e.g., porosity) are held constant and there is no retardation, modelled peak concentrations should be proportional to hydraulic conductivity (i.e., lower peak concentrations with lower hydraulic conductivity).

Furthermore, we reiterate that the modelled concentrations in both scenarios (0.11 m/d and 0.86 m/d) are at least two orders of magnitude below the most conservative human or ecological health-based criteria (0.055 mg/L, aluminium [NZWQS 95%]; 11.3 mg/L, nitrate-N [NZDWS]).

The effect of dilution by upgradient groundwater will not be significant (i.e., relative to advection / dispersion) where there is no retardation. There will likely be some dilution from surface infiltration along the flow path, which will *reduce* concentrations; however, this was not considered in the model. The cross-sectional area of the aquifer is not germane.

In summary, the findings reported in the ENGEO memo dated 30 January 2025 are valid and consistent with anticipated results related to a change in hydraulic conductivity. Furthermore, the hydraulic conductivity applied in the ENGEO model (0.86 m/d) is a conservative estimate from the slug testing and, even under this scenario, the maximum modelled contractions at the boundary are either not detectable or orders of magnitude below the human or ecological-health based criteria.

Mounding

The predicted maximum mounding is based on a conservative scenario whereby the mound extends from the water table up to the base of the soakage area (saturated conditions below the soakage area). As stated in the ENGEO memo dated 21 November 2024, the anticipated infiltration of the soakage is on the order of 200 m³/d – 300 m³/d. We acknowledge that the length and width of the basin bottom (10 m x 10 m) were not halved when the dimensions were input into the model. The corrected Hantush model (Hantush, 1967) is included in Appendix 1. As shown, correcting the dimensions of the soakage area resulted in a 58% *reduction* to the mounding height while still allowing for a daily discharge of 200 m³/day.

As noted in our letter dated 30 October 2024, the actual discharge volumes will vary and depend on the infiltration capacity in the soakage area, rainfall, and evaporation rates. Uncertainties related to the infiltration rate do not present a potential risk to off-site receptors, as they can be managed through conditions of consent. WCRC has been provided with proposed conditions of consent (Cook, K [25/11/24] email to Michael Durand “RE: RC-2024-0122 – s92 response”) to address uncertainties relating to the infiltration rate, specifically:

12	Discharging shall be undertaken at a rate that is no greater than the infiltration capacity of the soakage area, with a <i>maximum</i> discharge rate of 800 m ³ /day.
14	The soakage area shall maintain a freeboard of <i>at least</i> 0.5 m. If the rate of discharge is such that a 0.5 m freeboard cannot be maintained, the flowrate shall be reduced to a rate that maintains a <i>minimum</i> 0.5 m freeboard.

In summary, we do not agree with the PDP statement that the results of the mound modelling indicate that discharge to ground may not be feasible. While ENGEO concurs that the soakage area may not be able to accommodate the maximum discharge of 800 m³/d, it will still be able to receive discharges to achieve the application objectives without causing a *greater than minor* offsite effect.

- Updated contaminant modelling simulations using input parameters more consistent with high parameter concentrations that have been measured in leachate and wastewater.

PDP considers that additional information is required to assess the validity of the modelling results. The above section addresses PDP queries.

- An ecological survey of the surface water features and aquatic habitat surrounding the site, including Noname Creek, the drain along Taylorville Road, including that part of the drain that begins at spring seep TCD1 and the streams on the southern side of Taylorville Road (excluding the Grey River) and an impact assessment on surface water quality and ecological values resulting from the proposed discharge.

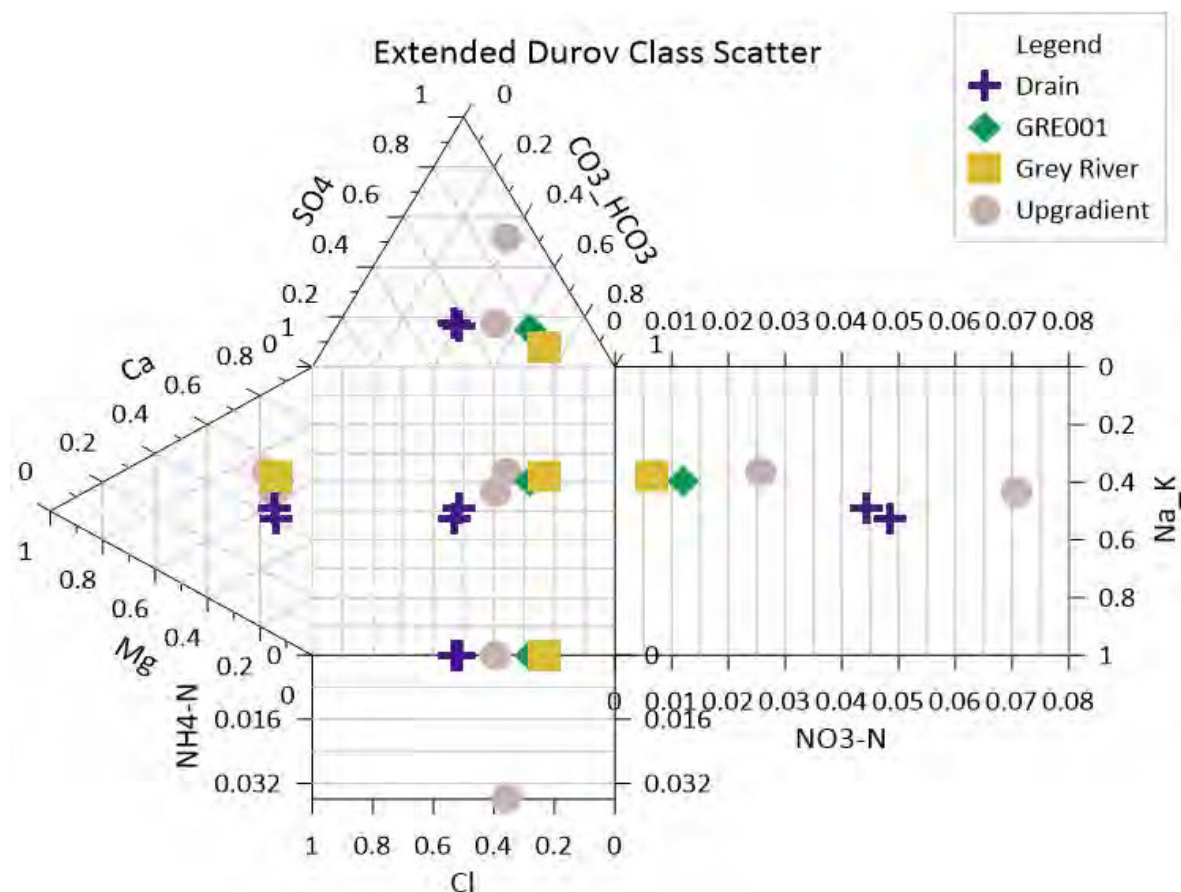
PDP has not reviewed the ecological impact assessment prepared by Bio-researchers.

5. A water quality comparison between water samples collected from the GDC water supply, the Grey River, groundwater upgradient of the landfill north of Taylorville Road and surface water samples collected at TRD2 and TRD4. The samples should include analysis for anions and cations and nitrogen species, in addition to other parameters sampled to date by TRPL. The purpose of the sampling is to inform whether the GDC water supply source is dominantly Grey River water and that there is minimal input to the supply bores from upgradient groundwater and surface drainage.

PDP notes that samples collected from TRD2 and TRD4 were not included in the comparison. We have revised our plot (Figure 6) to include only upgradient bores (MB1, MB19s), the Grey River (GR-01), untreated water for the GRE001 supply (PWT01), and the requested drain locations (TRD2 and TRD4). The drain locations were excluded from the original assessment due to the influence of other inputs along Taylorville Road and because water quality in samples collected from monitoring bores MB4, MB14, and MB18 are more relevant to assess inputs from the operations.

We refer to Figure A5.1 of the 30 January 2025 memo for sample locations.

Figure 6: Durov Plot of GRE001 Water and Upgradient Locations



As shown, the water sample from untreated water from the GRE100 supply (◆) plots most closely with the sample collected from the Grey River (■) with some similarity to upgradient groundwater (●).

We note that there is another deposit of the Pleistocene gravels (on the south side of Taylorville Road) that is not hydraulically connected to the subject site (ENGEO, 2024). Considering that ambient groundwater quality is likely similar in both deposits and the abstraction area is located approximately

100 m due south of the deposits on the south side of Taylorville Road, it is inferred that groundwater discharges from these deposits may be influencing water quality in the abstraction gallery, rather than waters from upgradient of the site.

The sample collected from the untreated water at the GRE001 water supply was most dissimilar to the samples collected from the drain adjacent to the site (+).

PDP notes that, compared to water samples collected from downgradient of the operations, there is more similarity between the water sample collected from the raw (untreated) water at GRE001 (PWT01), the Grey River approximately 300 – 400 m upstream of GRE001 (GR-01), and water samples collected upgradient of the operations. As stated in the ENGEO memo dated 30 January 2025, this is to be expected as all samples were collected from the same catchment.

In the ENGEO memo dated 30 January 25, it was noted that concentrations of dissolved metals (copper, lead, nickel, zinc) in the sample collected from the GRE100 supply (PWT01) differed significantly from the sample collected from the Grey River (GR-01) and samples collected along the southern site boundary. PDP states that, unless the applicant can identify another potential cause of the elevated metals, it may be interpreted that these concentrations are due to landfill operations.

We note that the concentrations detected in PWT01 are higher—by several orders of magnitude—than *any concentrations recorded* in water samples collected from Taylorville Road Drain, locations in the water treatment system, or monitoring bores located downgradient of the operations (ENGEO, 2024) (ENGEO, 2025). As such, there is no basis to suggest that the concentrations of dissolved metals in PWT01 are related to the operations.

It is beyond the scope of the proposed works to require the applicant to undertake a detailed investigation of all potential inputs to the GDC abstraction gallery which, as has been noted (ibid), is upgradient of the operations.

As stated in the memo dated 30 January 2025, it is inferred that concentrations in PWT-01 are naturally occurring (e.g., concentrations in the Holocene gravels) or due to the influence of waters originating at another (offsite) source area.

6. Yield versus drawdown data from the GDC supply bores to confirm the gravel deposits immediately adjacent to the Grey River are of high permeability with a high hydraulic connection to the Grey River.

ENGEO analysed pumping test data provided by GDC using published and industry accepted methods (Boulton [1963], Neuman [1975]) and hydrogeological software (Aquifer Test 11.0) to estimate hydraulic conductivity in the Holocene gravels.

PDP states, without details, they do not consider that these are accurate estimates; however, they do not consider further analysis necessary as there was very little drawdown in the bore.

ENGEO notes only that these estimates are consistent with a variety of published estimates for unconsolidated gravels (Fetter, 2021).

7. XRD sample analysis from the Kaiata mudstone underlying the landfill to confirm clay mineralogy in the mudstone and confirm ENGEO comments that swelling within any fractures that may be present in the mudstone will block the transmission of groundwater via a fracture network.

As stated in the memo dated 30 January 2025, XRD analysis of two mudstone outcrops show that it is comprised of a mixture of swelling and non-swelling clay minerals. PDP maintains that there is still potential from migration from the site to the GRE100 supply bores through fracture flow in the mudstone.

There is no confining pressure at the site or natural force that will cause significant quantities of groundwater to migrate vertically from the Pleistocene gravels into the mudstone—which has a hydraulic conductivity that is lower by orders of magnitude—much less induce it to migrate hydraulically upgradient via fracture flow.

Regardless of the extent of investigations, due to natural heterogeneity and uneven stress / strain placed on rock formations, it is not feasible to disprove the *possibility* of fractures that trend in the direction from the site to the GRE001 abstraction bores and that fractures (if any) could *possibly* transmit minute quantities of water along a 300–400 m flowpath. However, based on the presence of swelling clays and absence of observable fracturing, we consider this a *highly unlikely* migration pathway.

We note that the threshold for the purpose of processing the application is not to prove a null hypothesis; it is to assess the likelihood of a *greater than minor* deleterious effect, and that the applicant is being asked to provide evidence that far exceeds the scale and significance of the proposed activity.

To consider potential effects, we used Darcy's Equation ($Q=KiA$) to calculate an estimated daily flow (Q) of groundwater from upgradient onto the site that is on the order of $50 \text{ m}^3/\text{day}^1$. In correspondences from GDC to the West Coast Regional Council (Lane, L [19 January 2024] to Chris Barnes, "*RE: Flow Rate Compliance Report*"), GDC reports an average flow rate from the GRE001 abstraction bores ($8,513 \text{ m}^3/\text{day}$) that is 170 times higher than the *total flow* of groundwater across the site.

In summary, it is considered *highly unlikely* that groundwater from the site will migrate through the mudstone from the site to the abstraction bores and, noting the relative flow rates, it is *highly unlikely* that any incidental migration would be detectable or have a *greater than minor* effect.

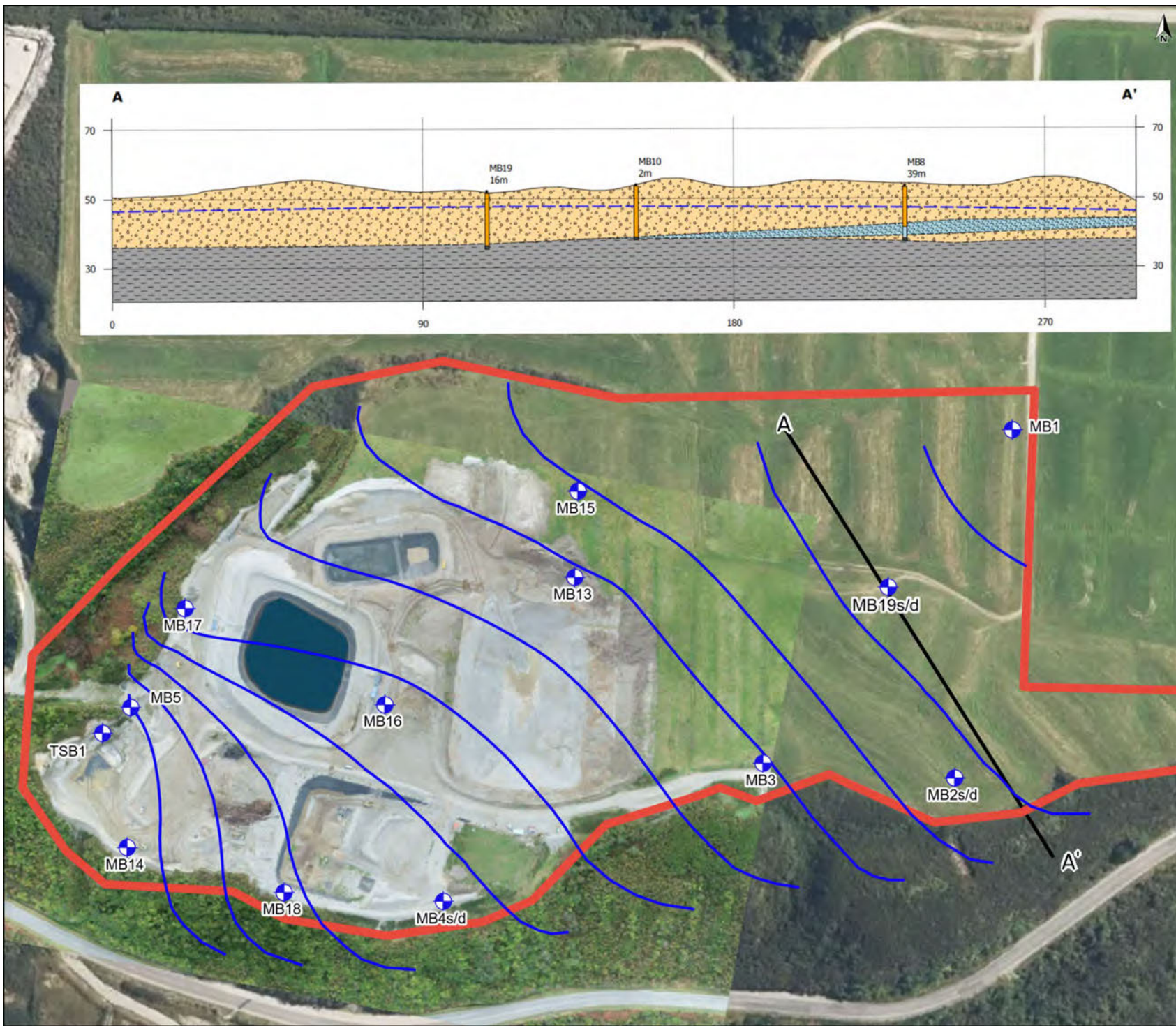
Based on our assessment works to date and in conjunction with the proposed conditions of consent, it is deemed *highly unlikely* that the proposed discharges will have a deleterious effect on human or ecological health.

This memorandum has been prepared for the use of our client, their professional advisers and the relevant Territorial Authorities in relation to the specified project brief described in this memo. No liability is accepted for the use of any part of the report for any other purposes or by any other person or entity. This Limitation should be read in conjunction with the Engineering NZ / ACENZ Standard Terms of Engagement.

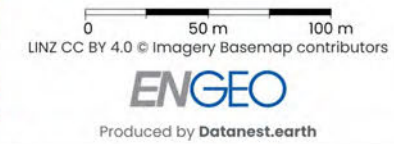
¹ Estimated using the hydraulic conductivity ($K=1 \times 10^{-5} \text{ m/s}$), the hydraulic gradient ($i=0.018$), and the approximate cross-sectional area of the saturated aquifer perpendicular to flow between the head of Noname Creek and the extent of the Pleistocene gravels ($A=3,000 \text{ m}^2$ – $3,600 \text{ m}^2$; refer Figure 6.)

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- ENGEO. (2024). *Preliminary Hydrogeological Assessment*. Job No. 26200.000.001: Prepared for TRPL.
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- Fetter, C. (2021). *Applied Hydrogeology*. Upper Saddle River, NJ: Prentice-Hall, Inc.
- Hantush, M. (1967). Growth and Decay of Groundwater Mounds in Response to Uniform Percolation. *Water Resources Journal* (3), 227-234.



- Legend**
- A-A' Cross Section (2x scale)
 - Monitoring Well Locations
 - Inferred Groundwater Contour Lines
 - Approximate Site Boundary
 - Line of Cross Section



Title: Pleistocene Aquifer Cross Section		
Client: Taylorville Resource Park Ltd		Size: A4
Project: Taylorville Resource Park	Drawn: LN	Figure No: 6
Date: 21-01-2025	Checked: DR	
Proj No: 26200.000.001	Scale: 1:3068	Version: 1.0

APPENDIX 1

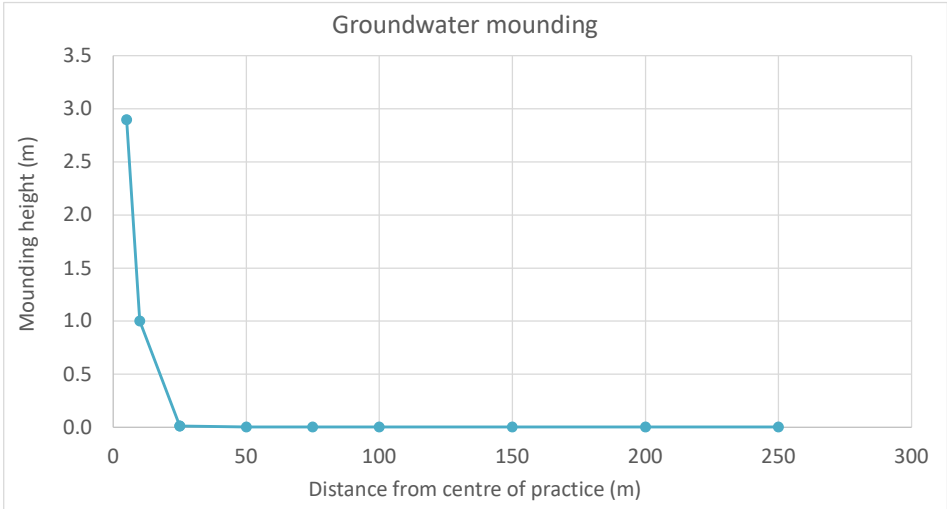
Corrected Hantush Model

This spreadsheet will calculate the height of a groundwater mound beneath an stormwater infiltration BMP.
More information can be found in the U.S. Geological Survey Scientific Investigations Report 2010-5102
"Simulation of groundwater mounding beneath hypothetical stormwater infiltration basins".

[Go to the USGS report](#)

Input Values		
Recharge (infiltration) rate (m/day)	2.000	R
Specific yield, S_y (dimensionless)	0.190	S_y
Horizontal hydraulic conductivity, K_h (m/day)	0.860000	K_h
1/2 length of basin (x direction, in m)	5.0	x
1/2 width of basin (y direction, in m)	5.0	y
Duration of infiltration period (days)	1	t
Initial thickness of saturated zone (m)	8.0	$h_i(0)$

CALCULATE



Distance from center of infiltration BMP (m)	Ground-water mounding (m)
5	2.90
10	1.00
25	0.01
25	0.01
50	0.00
75	0.00
100	0.00
150	0.00
200	0.00
250	0.00

IF YOU CHANGE THE SCALE OF THE
DISTANCE MEASUREMENTS
(OR ANY OTHER PARAMETER),
YOU MUST RECALCULATE!!

Site Memorandum

Project Name	Taylorville Resource Park		
Memo Topic	RC-2024-0122 and RC-2021-0097-V10 - RFI Response		
Memo Date	17 March 2025	Date of Site Visit	N/A
ENGEO Memo No.	10	ENGEO Project No.	26200.000.001
To	Taylorville Resource Park Ltd	Copy	Barker and Associates
Author	Lyn Nugent	Reviewers	Richard Griffiths Neil Charters

ENGEO attended discussions with the West Coast Regional Council (WCRC) on 7 March 2025, to discuss further information requested by WCRC in relation to resource consent application RC-2024-0012. WCRC has requested additional information related to the further information request issued by WCRC (dated 21-Oct-24) under Section 92(1) of the Resource Management Act 1991. WCRC queries are shown in grey, with ENGEO responses are shown in blue.

1. Will mounding beneath the soakage area affect slope stability?

Infiltrating water beneath the soakage area may mound to an elevation that intersects with the base of the soakage area, which will be approximately 8 m above the current water table. The rate of infiltration will be limited by the vertical hydraulic conductivity in the Pleistocene gravels. As the gravels are not bedded, we assume a relatively isotropic formation, meaning that the horizontal conductivity is likely very similar to the vertical and the extent of mounding will be proportionally limited (i.e., the formation will not accept more water than it can disperse).

Our conservative mounding estimate (Appendix 1) shows the lateral extent of any mounding will be limited to a radius of ~20 m beyond the infiltration basin. Beyond that distance, the mounding effect will be significantly less than the inferred seasonal variability of the water table (< 4 cm).

The proposed soakage area is located approximately 100 m away from the head of Noname Creek and approximately 150 m away from the extents of the Pleistocene gravels, which demarcate the nearest sloping ground. This negligible groundwater response from mounding will not increase the risk of slope instability.

As noted in our letter dated 30 October 2024, the actual discharge volumes will vary and depend on the infiltration capacity in the soakage area, rainfall, and evaporation rates. Uncertainties related to the infiltration rate do not present a potential risk to offsite receptors, as they can be managed through conditions of consent. WCRC has been provided with proposed conditions of consent (Cook, K [25/11/24] email to Michael Durand “RE: RC-2024-0122 – s92 response”) to address uncertainties relating to the infiltration rate, specifically:

12	Discharging shall be undertaken at a rate that is no greater than the infiltration capacity of the soakage area, with a maximum discharge rate of 800 m3/day.
14	The soakage area shall maintain a freeboard of at least 0.5 m. If the rate of discharge is such that a 0.5 m freeboard cannot be maintained, the flowrate shall be reduced to a rate that maintains a minimum 0.5 m freeboard.

2. Provide updated contaminant modelling showing a scenario with no retardation.

Updated modelling is attached in Appendix 2. As shown, even in a high hydraulic conductivity ($K=0.86$ m/d) scenario with no attenuation / degradation and no retardation ($K_d/K_{oc}=0$), and using the maximum concentrations detected at the site, concentrations at the downgradient boundary will not exceed the Drinking Water Standards for New Zealand (DWSNZ) or the Australian and New Zealand Guidelines for Fresh and Marine Water Quality default guideline values (DGVs) for 90% protection of freshwater species. This modelling is based on several conservative assumptions:

- The concentrations of the discharges will be equivalent to the highest concentrations ever recorded at the site (e.g., undiluted leachate from Cell C in July 2023);
- There will be no interactions between contaminants and aquifer materials (e.g., inert aquifer media);
- There will be no interactions with organic matter in the Pleistocene gravels; and
- There will be no nitrification / oxidation along the flow path.

These assumptions overestimate the actual anticipated, detectable concentrations at the boundary for several reasons:

- The modelled concentrations were undiluted and have not been detected in over a year. They were likely related to the discharge of greenwaste (i.e., tomato plants) in Cell C (January 2023). Waters in the retention pond (waters proposed for discharge) are significantly lower due to treatment (flocculation), aeration, and dilution with stormwater.
- Site-specific K_d values have been developed for heavy metals (Table 1); as shown, the mobility is generally classified as slightly mobile to immobile.
- We have not quantified partition coefficients for nitrogen compounds; however, there may be some adsorption to the substrate media. Based on field measurements (e.g., redox, dissolved oxygen) and laboratory results (Appendix 3), the subsurface conditions are suboxic near the proposed soakage area and become oxic as water moves downgradient towards the Pleistocene gravels, indicating that conditions are favourable to nitrification, which would reduce the concentration of ammoniacal-N.

This demonstrates it is *highly unlikely* that discharges will have a *greater than minor effect* at the downgradient property boundary.

WCRC has been provided with proposed conditions of consent (Cook, K [25/11/24] email to Michael Durand “RE: RC-2024-0122 – s92 response”) to address uncertainties relating to concentrations in discharges and the fate and transport of determinands in discharge waters, specifically:

17	Prior to the commencement of discharges, baseline groundwater samples shall be collected from monitoring bores MB13 and MB15 and analysed for the following determinands: • Sum of Anions • Sum of Cations • An/Cat balance (%) • pH • Total Alkalinity • Bicarbonate. CO₃, CO₂ • Total Hardness • Electrical conductivity (mS/m) • Dissolved aluminium • Total aluminium • Dissolved calcium • Dissolved iron • Total iron • Dissolved magnesium • Total magnesium • Dissolved manganese • Total manganese • Dissolved potassium • Dissolved sodium • Chloride • Total Ammoniacal-N • Nitrite-N • Nitrate-N • TKN • Total-N • Total-P • DRP • Sulphate • cBOD₅ • Total and dissolved heavy metals (arsenic, cadmium, chromium, copper, lead, nickel, zinc)
18	For six weeks following the commencement of discharges, groundwater samples shall be collected at 2-week intervals from monitoring bores MB13 and MB15 and analysed for the determinands in Condition 17.
19	Following the two-weekly groundwater monitoring described in Condition 16 groundwater samples shall be collected at monthly intervals for the duration of discharges from monitoring bores MB13 and MB15 and analysed for the determinands in Condition 17.
20	Following each monitoring round described in Conditions 18–19, groundwater analytical results shall be compared to the baseline concentrations (refer Condition 17). If the <i>change</i> in groundwater quality exceeds 50% of the New Zealand Drinking Water Standard (NZDWS), discharges shall cease and the actions in Condition 21 shall be implemented.
21	If the <i>change</i> in groundwater quality at monitoring bores MB13 or MB15 exceeds 50% of the NZDWS, the Applicant shall: Cease discharges to the soakage area; Notify the WCRC Group Manager – Regulatory & Policy; Engage a Suitably Qualified and Experienced Practitioner (SQEP) to Assess if the change in groundwater quality at MB13 and MB15 is due to discharges or other (e.g., upgradient water quality) factors; Characterise the potential offsite effect of changes due to changes in water quality on likely human and ecological receptors; If necessary, develop a revised monitoring programme at downgradient monitoring bores and develop mitigation measures; and Prepare a Response Plan for submission to the WCRC.

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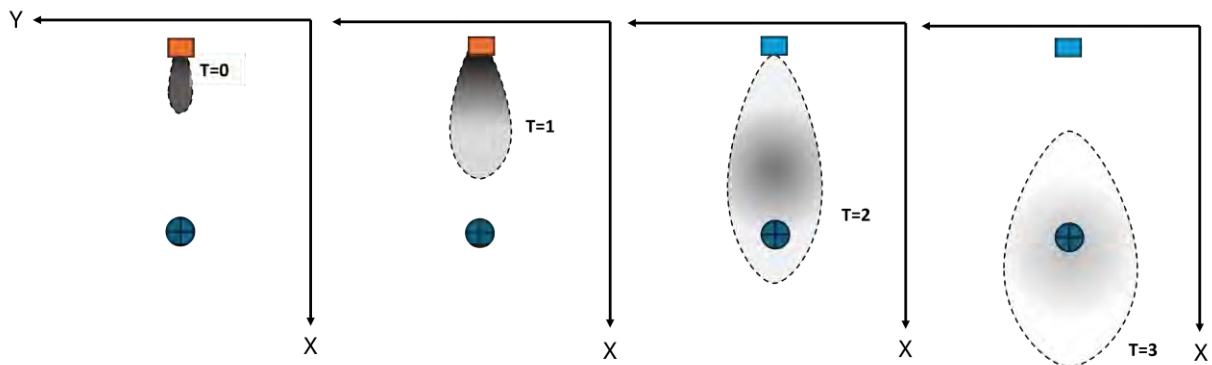
All groundwater quality monitoring results shall be provided to the WCRC Group Manager – Regulatory & Policy within one week of them being received in an electronic format, suitable for automatic upload to a water quality database. At the conclusion of discharges, a summary email will be prepared for the WCRC that includes:

- Estimated total volume of discharges
- Tabulated analytical results and a trend analysis
- Exceedances of Condition 20 criteria and any mitigation actions

3. How will a pulse source (six-month discharge) affect the mass flux of contaminants or concentrations at the receptor?

The plume maximum will extend downgradient from the source area, with a stable source area for the duration of discharges (Figure 1). However, the concentrations will not achieve steady-state conditions and, once the discharges cease, the maximum concentration will be transported downgradient and reduced due to minute differences in flow velocity (Hemond, 2000). The total mass of contaminants in the plume will be the same as the mass of contaminants discharged (excluding degradation / attenuation / decay). Concentrations at the receptor are provided in Appendix 2.

Figure 1: Plume Migration with a Pulse Source at Different Time Intervals



Note: Source area discharging at T=0 and T=1 (orange rectangle), no discharges from the source area at T=2 and T=3 (blue rectangle). Downgradient receptor shown as a blue circle.

4. Provide additional comparison between water collected at the GRE001 supply (PWT01), the Grey River (GR-01), Taylorville Road drain locations (TRD1—TRD5), upgradient groundwater monitoring bores (MB1, MB19s/d) and locations downgradient of the operations.

As discussed in our memos dated 30 January and 5 March 2025, concentrations of heavy metals in the sample collected from PWT-01 are higher—by several orders of magnitude—than any concentrations recorded in water samples collected from:

- Taylorville Road drain
- The water treatment system, or
- Monitoring bores located downgradient of the operations.

To further assess for any potential contributions from the site to the sample collected at the GRE001 supply, we compared the analytical results for copper, nickel, lead, and zinc in all samples collected during the January 2025 monitoring round (Table 2). While dissolved metals concentrations are more applicable for this migration pathway, we have considered concentrations of both total and dissolved metals.

The ternary diagrams in Appendix 3 show that the proportions of metals in the sample collected at PWT-01 are significantly different from any groundwater samples collected at the site (upgradient and downgradient), Taylorville Road drain, the Grey River and the river tributaries.

The solubility of heavy metals is dependent on the pH and presence of anionic species (e.g., hydroxide, sulphate); however, in general, at a neutral pH (7), the copper will be relatively insoluble followed by zinc, lead, and nickel (most soluble) (Karimi, 2017) (Lewis, 2010) (MacBerthouex, 2019) (Brezonik, 2011). If the concentrations at PWT-01 originated from any of the other sample locations, the proportion of soluble metals (e.g., nickel) would be higher than insoluble (e.g., copper) relative to the source; however, the inverse condition exists.

We also compared the concentrations of the target heavy metal determinands (Appendix 3). As shown, the absolute and relative concentrations of heavy metals in the sample collected from PWT-01 do not resemble those in *any other sample collected*. This is additional to the fact that there is no mechanism that would cause metals concentrations from the site to concentrate by orders of magnitude on the flow path from the site to the abstraction bores.

This demonstrates that there is no basis to suggest that the concentrations of total and dissolved metals in PWT-01 are related to the operations and it is beyond the scale and significance of the proposed works to require the applicant to undertake additional sampling or investigation works.

WCRC requested a review of the Durov scatter plots to explain differences in the plot provided on 30 January and the plot provided on 5 March 2025. Both plots were based on the sample dataset; however, the plotted samples differed between the two memoranda. The original plot did not include samples from the upgradient bores and Taylorville Road drain locations, but did include downgradient / cross gradient bores. In accordance with WCRC's request of 25 February 2025, the plot was updated to include only upgradient bores (MB1, MB19s), the Grey River (GR-01), untreated water for the GRE001 supply (PWT01), and the requested drain locations (TRD2 and TRD4). We do note that the axis titles for the anions were incorrectly placed in the figures provided on 30 January; this was corrected in the 5 March memorandum.

5. WCRC/PDP is not satisfied that swelling within any fractures that may be present in the mudstone will block the transmission of groundwater via a fracture network.

We understand this question is essentially in relation to a concern that contaminated groundwater may be able to travel below / downgradient of the soakage area to the GDC supply bores through fractures in the mudstone.

We have drawn up a cross section from the supply bores to the landfill (Figure 2). We consider it *highly unlikely* that fracture flow of potentially contaminated groundwater (if any) will have a *greater than minor effect* on users of the GRE001 water supply based on the following rationale:

- Previous investigations (Golder and Associates, 2005) (Bishop, 2004) estimated the horizontal hydraulic conductivity of the Kaiata Mudstone (1×10^{-8} m/s), with an inferred ratio of horizontal to vertical conductivity of at least 10:1 (Domenico, 1990). Based on field observations made by ENGEO along exposures of the Kaiata Mudstone, the formation is massive with little evidence of fracturing.
- The measured hydraulic conductivity in the overlying Pleistocene gravels is higher by 2–3 orders of magnitude (1×10^{-5} m/s– 1×10^{-6} m/s). There is no confining pressure at the site or natural force that would cause significant quantities of groundwater to migrate vertically from the much higher permeability Pleistocene gravels into the lower permeability mudstone.

- Flow from the discharge area to the abstraction bores would require a *highly unlikely* scenario which would need to involve:
 - o Discharges from the soakage area migrating vertically (perpendicular to the flow of groundwater) to the mudstone, **and**
 - o The discharges encounter a fracture(s) that runs perpendicular to the regional orientation of mapped faults (northeast–southwest) and are connected along the 300–400 m flow path to the abstraction wells, **and**
 - o The fractures do not intersect with any preferential pathways in the bedding planes in the mudstone and travel west-southwest along the dip direction, **and**
 - o The fractures have an angle(s) that will cause them to daylight near the abstraction bores (i.e., and not in the road cut or the banks of the Grey River channel), **and**
 - o The mudstone does not contain swelling clays at an adequate proportion to hinder the flow of groundwater, **and**
 - o The discharges are not diluted by any other groundwater flowing through the mudstone; **and**
 - o The discharges are not captured or diluted by groundwater flow in the Holocene gravels, **and**
 - o The discharges intersect with the radius of influence of the abstraction wells (noting that there is very little measured drawdown).

Even assuming a *highly conservative* scenario whereby 1% of the maximum daily discharges (8 m³/d) are transported via the *highly unlikely* scenario outlined above, the discharges would still comprise *less than 0.1%* of the average flow rate from the abstraction bores (8,513 m³/d) and the resulting concentrations would be orders of magnitude below any of the human health-based criteria.

PDP has recommended a Lugeon test to assess for the *possibility* of fracture flow. Noting again that the threshold for notification is based on the likelihood and magnitude of potential effect(s), rather than possibility, a Lugeon test has several limitations that make it unsuitable to address this query:

- It will not differentiate different flow pathways (e.g., bedding vs. fracture),
- It will not determine the orientation and angle of any fractures, and
- It will not determine the connectedness of any flow paths between the site and the abstraction wells.

As such, this issue is most appropriately addressed through a logic test and, as described above, it is considered *highly unlikely* that groundwater from the site will migrate through the mudstone from the site to the abstraction bores and, noting the relative flow rates, it is *highly unlikely* that any incidental migration would be detectable or have a *greater than minor* effect.

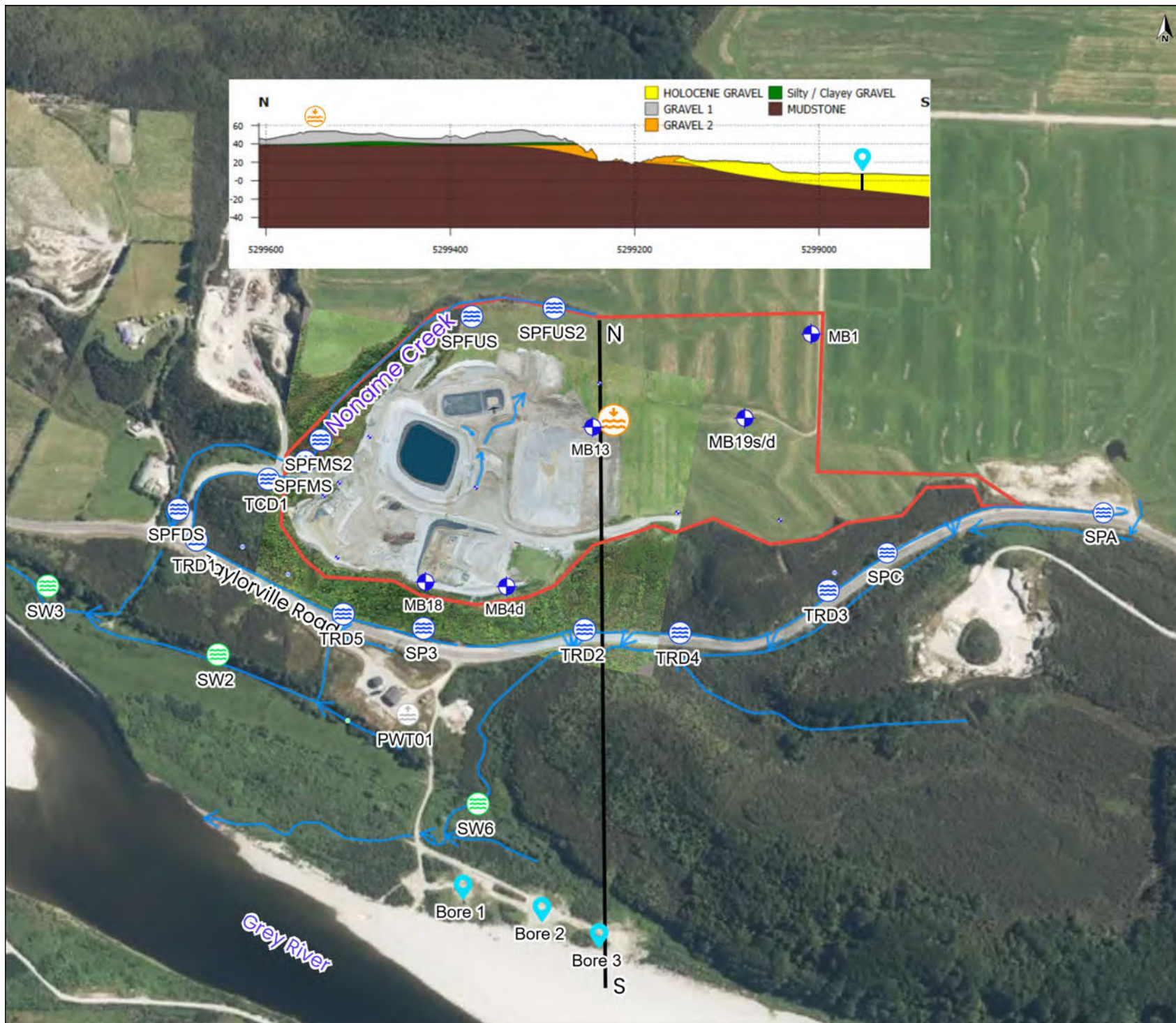
If there are additional queries regarding fracture flow transport through the mudstone, we recommend a meeting between one of the Principal Engineering Geologists at ENGEO and their PDP counterpart(s).

Based on our assessment works to date and in conjunction with the proposed conditions of consent, it is deemed **highly unlikely** that the proposed discharges will have an adverse effect on human or ecological health.

This memorandum has been prepared for the use of our client, their professional advisers and the relevant Territorial Authorities in relation to the specified project brief described in this memo. No liability is accepted for the use of any part of the report for any other purposes or by any other person or entity. This Limitation should be read in conjunction with the Engineering NZ / ACENZ Standard Terms of Engagement.

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- Legend**
- Surface Water Sampling Locations
 - Coal Creek WTP
 - Grey River Sample Location
 - Monitoring Well Locations
 - Tributary Surface Water Sample Locations
 - GRE001 Abstraction Wells
 - Proposed Discharge Location
 - Approximate Site Boundary
 - Line of Cross Section

0 100 m 200 m
LINZ CC BY 4.0 © Imagery Basemap contributors

ENGEO

Produced by **Datanest.earth**

Title:
Sampling Locations and Cross Section
(soakage area to abstraction wells)

Client: Taylorville Resource Park Ltd		Size: A4
Project: Taylorville Resource Park	Drawn: PS	Figure No: 2
Date: 12-03-2025	Checked: DW	
Proj No: 26200.000.001	Scale: 1:5730	Version: 1.0

Table 1: Gravel Analytical Results and Calculated Partition Coefficients

Analyte	BH13	BH13	BH13	BH15	BH15	BH15	BH19	BH19	BH19
	0m - 1.5m	9m - 10m	17m - 18m	0m - 1.5m	13m - 15m	16m - 18m	0m - 1.5m	7m - 8m	15.6m - 16.5m
	18-10-2024	18-10-2024	18-10-2024	18-10-2024	18-10-2024	18-10-2024	18-10-2024	18-10-2024	18-10-2024
Total Metals (mg/kg)									
Arsenic	3	5	13	3	7	13	3	3	15
Chromium (III+VI)	20	13	34	14	23	31	10	13	29
Copper	9	11	17	8	12	16	7	9	15
Lead	11	18.7	11.6	9.3	19.5	11.7	8.5	16	12.8
Nickel	14	13	36	11	15	36	9	13	40
Zinc	38	50	81	31	52	76	26	42	73
SPLP Metals (mg/L)									
Arsenic	<i>0.0011</i>	0.0013	0.0022	<i>0.0011</i>	0.0028	<i>0.0011</i>	<i>0.0011</i>	0.002	<i>0.0011</i>
Chromium (III+VI)	0.003	0.0023	0.0057	<i>0.00053</i>	0.0086	0.0007	<i>0.00053</i>	0.0049	<i>0.00053</i>
Copper	0.00191	0.0032	<i>0.00053</i>	<i>0.00053</i>	0.0061	0.00054	<i>0.00053</i>	0.0047	<i>0.00053</i>
Lead	0.002	0.0044	0.00021	<i>0.00011</i>	0.0086	0.00012	<i>0.00011</i>	0.0082	<i>0.00011</i>
Nickel	0.00172	0.0027	0.0049	<i>0.00053</i>	0.006	0.00132	<i>0.00053</i>	0.0039	0.0013
Zinc	0.0055	0.0122	<i>0.0011</i>	<i>0.0011</i>	0.023	<i>0.0011</i>	<i>0.0011</i>	0.0158	<i>0.0011</i>
Soil-Water Partition Coefficient (L/kg)									
Arsenic	2707	3826	5889	2707	2480	11798	2707	1480	13616
Chromium (III+VI)	6647	5632	5945	26395	2654	44266	18848	2633	54697
Copper	4692	3418	32055	15074	1947	29610	13188	1895	28282
Lead	5480	4230	55218	84525	2247	97480	77253	1931	116344
Nickel	8120	4795	7327	20735	2480	27253	16961	3313	30749
Zinc	6889	4078	73616	28162	2241	69071	23616	2638	66344

Notes:

All values provided in units of milligrams per kilogram (mg/kg).

Values in *italics* were detected below the laboratory reporting limit (RL shown)

McCall's Soil Mobility Classification Scheme Colour Key:

Mobility class	Range of Kd
Very high	0 - 50
High	50 - 150
Medium	150 - 500
Low	500 - 2000
Slightly	2000 - 5000
Immobile	> 5000

Table 2: January 2025 TRP Sample Results

Analyte	NZDWS ¹	NZWQS (95%) ²	NZWQS (90%) ²	NZWQS (80%) ²	Grey River Tributaries			Grey Rvr	GRE001	Noname Creek				
					SW2	SW3	SW6	GR-01	PWT01	SPFUS	SPFMS	SPFDS	SPFUS2	SPFMS2
					20/01/25	20/01/25	20/01/25	20/01/25	20/01/25	20/01/25	20/01/25	20/01/25	20/01/25	20/01/25
Sum of Anions					5.2	2.9	0.78	0.58	0.68	0.87	1.95	1.86	1.12	0.84
Sum of Cations					4.8	2.9	0.81	0.56	0.65	0.86	2	1.89	1.19	0.87
An/Cat balance (%)														
pH					7.2	7.4	7.3	7.5	7	6.8	7.6	7.9	6.6	6.7
Total Alkalinity					27	70	12.9	21	22	12.3	63	57	11	11.9
Bicarbonate					33	85	15.7	25	27	15	77	69	13.4	14.5
Total Hardness					188	118	25	19.3	21	29	79	73	32	30
Electrical conductivity (mS/m)					51.6	30.3	8.8	5.7	6.9	10.2	20.5	19.7	13.8	9.9
Dissolved aluminium	1	0.055	0.08	0.15	0.061	0.006	0.125	0.035	0.012	0.07	0.015	0.023	0.21	0.065
Total aluminium		0.055	0.08	0.15	0.19	0.039	0.22	0.159	0.0183	0.084	0.074	0.044	0.28	0.117
Dissolved calcium					61	42	7.7	6.2	6.8	9.7	28	26	10.4	9.9
Dissolved iron					0.03	< 0.02	0.16	0.05	< 0.02	0.07	0.08	0.06	0.23	0.11
Total iron					0.171	0.29	0.28	0.26	< 0.021	0.132	0.55	0.128	0.43	0.5
Dissolved magnesium					8.5	3.2	1.3	0.9	1.0	1.2	2.2	2.1	1.34	1.19
Total magnesium					8.9	3.2	1.2	1.0	1.0	1.2	2.1	2.1	1.36	1.22
Dissolved manganese	0.4	1.9	2.5	3.6	0.0065	0.0034	0.0062	0.0009	< 0.0005	0.0151	0.096	0.0048	0.059	0.022
Total manganese		1.9	2.5	3.6	0.084	0.147	0.0125	0.0141	0.00059	0.021	0.162	0.0127	0.067	0.051
Dissolved potassium					3.4	3.8	0.8	0.71	0.94	1.48	3.3	3	6.8	1.26
Dissolved sodium					21	10.7	6.4	3.5	4.6	5.3	8.1	8	7.9	5.3
Chloride					17.6	10.5	8.7	3.7	5.2	13.2	10.2	9.9	27	13
Total-N					0.86	0.42	0.29	-	-	0.67	0.67	0.47	1.58	0.57
Total Ammoniacal-N		0.9	1.43	2.3	0.055	0.01	< 0.010	-	-	< 0.010	0.02	< 0.010	0.08	< 0.010
Nitrite-N	0.912				0.013	< 0.002	< 0.002	< 0.002	< 0.002	< 0.002	0.008	< 0.002	0.013	< 0.002
Nitrate-N	11.3				0.23	< 0.002	0.125	0.079	0.26	0.35	0.3	0.25	0.34	0.3
Nitrite+Nitrate-N					0.24	< 0.002	0.126	0.08	0.26	0.35	0.31	0.25	0.35	0.3
TKN					0.62	0.42	0.16	-	-	0.32	0.36	0.21	1.23	0.26
DRP					< 0.004	< 0.004	< 0.004	-	-	0.005	< 0.004	< 0.004	0.035	< 0.004
Total-P					0.025	0.021	0.021	-	-	0.017	0.027	0.007	0.172	0.015
Sulphate					197	58	12.7	2.3	3.3	10.9	18.1	20	4.6	10.5
cBOD5					< 2	< 2	< 2	-	-	< 2	< 2	< 2	< 2	< 2
COD					32	28	12	-	-	9	11	11	63	7
Dissolved arsenic	0.01	0.013	0.42	1.4	0.0013	0.0011	< 0.0010	< 0.0010	< 0.0010	< 0.0010	< 0.0010	< 0.0010	0.0015	< 0.0010
Dissolved cadmium	0.004	0.0002	0.0004	0.0008	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005	< 0.00005
Dissolved chromium	0.05	0.0033	0.006	0.04	0.0006	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005
Dissolved copper	2	0.0014	0.0018	0.0025	< 0.0005	< 0.0005	< 0.0005	< 0.0005	0.0115	0.0006	0.0006	0.0008	0.0024	0.0005
Dissolved lead	0.01	0.0034	0.0056	0.0094	< 0.00010	< 0.00010	< 0.00010	< 0.00010	0.00061	< 0.00010	< 0.00010	< 0.00010	0.00012	< 0.00010
Dissolved nickel	0.08	0.011	0.013	0.017	0.0008	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005	< 0.0005
Dissolved zinc		0.008	0.015	0.031	0.001	< 0.0010	< 0.0010	< 0.0010	0.0113	< 0.0010	0.0017	< 0.0010	0.0057	0.0014
Total arsenic		0.013	0.42	1.4	0.0016	0.0014	< 0.0011	< 0.0011	< 0.0011	< 0.0011	< 0.0011	< 0.0011	0.0018	< 0.0011
Total cadmium		0.0002	0.0004	0.0008	< 0.000053	< 0.000053	< 0.000053	< 0.000053	< 0.000053	< 0.000053	< 0.000053	< 0.000053	< 0.000053	< 0.000053
Total chromium (IV)		0.0033	0.006	0.04	0.00113	0.00084	< 0.00053	< 0.00053	< 0.00053	< 0.00053	0.00059	0.0006	0.0006	< 0.00053
Total copper		0.0014	0.0018	0.0025	0.00102	< 0.00053	0.00054	< 0.00053	0.053	0.00061	0.00074	0.00088	0.0027	0.00058
Total lead		0.0034	0.0056	0.0094	0.00025	0.0002	0.00014	0.00025	0.0023	< 0.00011	0.00071	0.00012	0.00031	< 0.00011
Total nickel		0.011	0.013	0.017	0.00083	< 0.00053	< 0.00053	< 0.00053	< 0.00053	< 0.00053	< 0.00053	< 0.00053	< 0.00053	< 0.00053
Total zinc		0.008	0.015	0.031	0.0047	0.0013	< 0.0011	< 0.0011	0.043	0.0014	0.003	0.0024	0.0059	0.0016
Benzo(a)Pyrene					-	-	-	-	-	-	-	-	-	-

Notes:

All values provided in units of miligrams per liter (mg/L) unless otherwise specified.

¹ Drinking-water Standards for New Zealand 2022. Table 2 (maximum acceptable values

² Australia and New Zealand Guidelines from Fresh & Marine Water Quality. Default Gui

Bordered values exceed the NZDWS MAV

Values in *italics* exceed the NZWQS, 95% level of protection

Values in **bold** exceed the NZWQS, 90% level of protection

Gray highlighted values exceed the NZWQS, 80% level of protection

APPENDIX 1

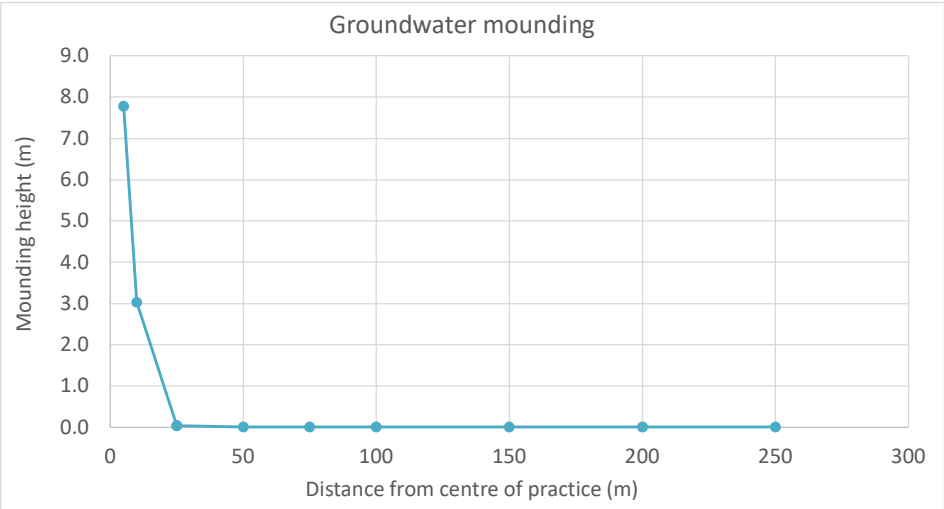
Hantush Model

This spreadsheet will calculate the height of a groundwater mound beneath an stormwater infiltration BMP.
More information can be found in the U.S. Geological Survey Scientific Investigations Report 2010-5102
"Simulation of groundwater mounding beneath hypothetical stormwater infiltration basins".

[Go to the USGS report](#)

Input Values		
Recharge (infiltration) rate (m/day)	6.000	R
Specific yield, S_y (dimensionless)	0.190	S_y
Horizontal hydraulic conductivity, K_h (m/day)	0.860000	K_h
1/2 length of basin (x direction, in m)	5.0	x
1/2 width of basin (y direction, in m)	5.0	y
Duration of infiltration period (days)	1	t
Initial thickness of saturated zone (m)	8.0	$h_i(0)$

CALCULATE



Distance from center of infiltration BMP (m)	Ground-water mounding (m)
5	7.78
10	3.03
25	0.04
25	0.04
50	0.01
75	0.01
100	0.01
150	0.01
200	0.01
250	0.01

IF YOU CHANGE THE SCALE OF THE
DISTANCE MEASUREMENTS
(OR ANY OTHER PARAMETER),
YOU MUST RECALCULATE!!

APPENDIX 2

RISC Model

Time (year)	Aluminium (mg/l)	Ammonium-N (mg/l)	Arsenic (mg/l)	Chromium (VI) (mg/l)	Copper (mg/l)	Lead (mg/l)	Manganese (mg/l)	Nickel (soluble salts) (mg/l)	Nitrate-N (mg/l)
0	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
1	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00	0.0E+00
2	6.5E-11	2.5E-09	2.5E-11	5.5E-12	4.8E-12	3.3E-13	1.7E-10	3.9E-12	1.6E-10
3	1.8E-07	6.7E-06	6.7E-08	1.5E-08	1.3E-08	8.8E-10	4.6E-07	1.1E-08	4.4E-07
4	8.1E-06	3.0E-04	3.0E-06	6.8E-07	6.0E-07	4.0E-08	2.1E-05	4.8E-07	2.0E-05
5	6.9E-05	2.6E-03	2.6E-05	5.8E-06	5.1E-06	3.4E-07	1.8E-04	4.1E-06	1.7E-04
6	2.5E-04	9.6E-03	9.6E-05	2.2E-05	1.9E-05	1.3E-06	6.6E-04	1.5E-05	6.4E-04
7	5.9E-04	2.2E-02	2.2E-04	5.0E-05	4.4E-05	3.0E-06	1.5E-03	3.6E-05	1.5E-03
8	1.0E-03	3.9E-02	3.9E-04	8.8E-05	7.7E-05	5.2E-06	2.7E-03	6.2E-05	2.6E-03
9	1.5E-03	5.7E-02	5.7E-04	1.3E-04	1.1E-04	7.5E-06	3.9E-03	9.0E-05	3.8E-03
10	1.9E-03	7.3E-02	7.3E-04	1.6E-04	1.4E-04	9.6E-06	5.0E-03	1.2E-04	4.8E-03
11	2.3E-03	8.5E-02	8.5E-04	1.9E-04	1.7E-04	1.1E-05	5.8E-03	1.4E-04	5.6E-03
12	2.5E-03	9.3E-02	9.3E-04	2.1E-04	1.8E-04	1.2E-05	6.4E-03	1.5E-04	6.2E-03
13	2.6E-03	9.7E-02	9.7E-04	2.2E-04	1.9E-04	1.3E-05	6.6E-03	1.5E-04	6.4E-03
14	2.6E-03	9.7E-02	9.7E-04	2.2E-04	1.9E-04	1.3E-05	6.6E-03	1.5E-04	6.4E-03
15	2.5E-03	9.4E-02	9.4E-04	2.1E-04	1.9E-04	1.3E-05	6.5E-03	1.5E-04	6.3E-03
16	2.4E-03	9.0E-02	9.0E-04	2.0E-04	1.8E-04	1.2E-05	6.1E-03	1.4E-04	5.9E-03
17	2.2E-03	8.3E-02	8.3E-04	1.9E-04	1.6E-04	1.1E-05	5.7E-03	1.3E-04	5.5E-03
18	2.0E-03	7.6E-02	7.6E-04	1.7E-04	1.5E-04	1.0E-05	5.2E-03	1.2E-04	5.1E-03
19	1.8E-03	6.9E-02	6.9E-04	1.5E-04	1.4E-04	9.2E-06	4.7E-03	1.1E-04	4.6E-03
20	1.6E-03	6.2E-02	6.2E-04	1.4E-04	1.2E-04	8.2E-06	4.2E-03	9.8E-05	4.1E-03
21	1.5E-03	5.5E-02	5.5E-04	1.2E-04	1.1E-04	7.3E-06	3.7E-03	8.7E-05	3.6E-03
22	1.3E-03	4.8E-02	4.8E-04	1.1E-04	9.5E-05	6.4E-06	3.3E-03	7.7E-05	3.2E-03
23	1.1E-03	4.2E-02	4.2E-04	9.4E-05	8.3E-05	5.6E-06	2.9E-03	6.7E-05	2.8E-03
24	9.7E-04	3.7E-02	3.7E-04	8.2E-05	7.2E-05	4.8E-06	2.5E-03	5.8E-05	2.4E-03
25	8.4E-04	3.2E-02	3.2E-04	7.1E-05	6.2E-05	4.2E-06	2.2E-03	5.0E-05	2.1E-03
26	7.2E-04	2.7E-02	2.7E-04	6.1E-05	5.4E-05	3.6E-06	1.9E-03	4.3E-05	1.8E-03
27	6.2E-04	2.3E-02	2.3E-04	5.2E-05	4.6E-05	3.1E-06	1.6E-03	3.7E-05	1.5E-03
28	5.3E-04	2.0E-02	2.0E-04	4.5E-05	3.9E-05	2.6E-06	1.4E-03	3.2E-05	1.3E-03
29	4.5E-04	1.7E-02	1.7E-04	3.8E-05	3.4E-05	2.3E-06	1.2E-03	2.7E-05	1.1E-03
30	3.8E-04	1.4E-02	1.4E-04	3.3E-05	2.9E-05	1.9E-06	9.9E-04	2.3E-05	9.6E-04
31	3.3E-04	1.2E-02	1.2E-04	2.8E-05	2.4E-05	1.6E-06	8.4E-04	2.0E-05	8.2E-04
32	2.8E-04	1.0E-02	1.0E-04	2.3E-05	2.1E-05	1.4E-06	7.1E-04	1.7E-05	6.9E-04
33	2.3E-04	8.8E-03	8.8E-05	2.0E-05	1.7E-05	1.2E-06	6.0E-04	1.4E-05	5.8E-04
34	2.0E-04	7.5E-03	7.5E-05	1.7E-05	1.5E-05	9.9E-07	5.1E-04	1.2E-05	4.9E-04
35	1.7E-04	6.3E-03	6.3E-05	1.4E-05	1.2E-05	8.3E-07	4.3E-04	1.0E-05	4.2E-04
36	1.4E-04	5.3E-03	5.3E-05	1.2E-05	1.0E-05	7.0E-07	3.6E-04	8.4E-06	3.5E-04
37	1.2E-04	4.5E-03	4.5E-05	1.0E-05	8.8E-06	5.9E-07	3.0E-04	7.1E-06	3.0E-04

38	1.0E-04	3.8E-03	3.8E-05	8.4E-06	7.4E-06	5.0E-07	2.6E-04	6.0E-06	2.5E-04
39	8.4E-05	3.2E-03	3.2E-05	7.1E-06	6.2E-06	4.2E-07	2.2E-04	5.0E-06	2.1E-04
40	7.0E-05	2.6E-03	2.6E-05	5.9E-06	5.2E-06	3.5E-07	1.8E-04	4.2E-06	1.8E-04
41	5.9E-05	2.2E-03	2.2E-05	5.0E-06	4.4E-06	2.9E-07	1.5E-04	3.5E-06	1.5E-04
42	4.9E-05	1.9E-03	1.9E-05	4.2E-06	3.7E-06	2.5E-07	1.3E-04	3.0E-06	1.2E-04
43	4.1E-05	1.6E-03	1.6E-05	3.5E-06	3.1E-06	2.1E-07	1.1E-04	2.5E-06	1.0E-04
44	3.5E-05	1.3E-03	1.3E-05	2.9E-06	2.6E-06	1.7E-07	9.0E-05	2.1E-06	8.7E-05
45	2.9E-05	1.1E-03	1.1E-05	2.5E-06	2.2E-06	1.5E-07	7.5E-05	1.7E-06	7.3E-05
46	2.4E-05	9.2E-04	9.2E-06	2.1E-06	1.8E-06	1.2E-07	6.3E-05	1.5E-06	6.1E-05
47	2.0E-05	7.7E-04	7.7E-06	1.7E-06	1.5E-06	1.0E-07	5.3E-05	1.2E-06	5.1E-05
48	1.7E-05	6.4E-04	6.4E-06	1.4E-06	1.3E-06	8.5E-08	4.4E-05	1.0E-06	4.3E-05
49	1.4E-05	5.4E-04	5.4E-06	1.2E-06	1.1E-06	7.1E-08	3.7E-05	8.6E-07	3.6E-05
50	1.2E-05	4.5E-04	4.5E-06	1.0E-06	8.9E-07	6.0E-08	3.1E-05	7.2E-07	3.0E-05
51	1.0E-05	3.8E-04	3.8E-06	8.4E-07	7.4E-07	5.0E-08	2.6E-05	6.0E-07	2.5E-05
52	8.3E-06	3.1E-04	3.1E-06	7.1E-07	6.2E-07	4.2E-08	2.2E-05	5.0E-07	2.1E-05
53	7.0E-06	2.6E-04	2.6E-06	5.9E-07	5.2E-07	3.5E-08	1.8E-05	4.2E-07	1.7E-05
54	5.8E-06	2.2E-04	2.2E-06	4.9E-07	4.3E-07	2.9E-08	1.5E-05	3.5E-07	1.5E-05
55	4.9E-06	1.8E-04	1.8E-06	4.1E-07	3.6E-07	2.4E-08	1.3E-05	2.9E-07	1.2E-05
56	4.1E-06	1.5E-04	1.5E-06	3.4E-07	3.0E-07	2.0E-08	1.0E-05	2.4E-07	1.0E-05
57	3.4E-06	1.3E-04	1.3E-06	2.9E-07	2.5E-07	1.7E-08	8.8E-06	2.0E-07	8.5E-06
58	2.8E-06	1.1E-04	1.1E-06	2.4E-07	2.1E-07	1.4E-08	7.3E-06	1.7E-07	7.1E-06
59	2.4E-06	8.9E-05	8.9E-07	2.0E-07	1.8E-07	1.2E-08	6.1E-06	1.4E-07	5.9E-06
60	2.0E-06	7.5E-05	7.5E-07	1.7E-07	1.5E-07	9.9E-09	5.1E-06	1.2E-07	5.0E-06
61	1.7E-06	6.2E-05	6.2E-07	1.4E-07	1.2E-07	8.3E-09	4.3E-06	9.9E-08	4.1E-06
62	1.4E-06	5.2E-05	5.2E-07	1.2E-07	1.0E-07	6.9E-09	3.6E-06	8.3E-08	3.5E-06
63	1.2E-06	4.3E-05	4.3E-07	9.8E-08	8.6E-08	5.8E-09	3.0E-06	6.9E-08	2.9E-06
64	9.6E-07	3.6E-05	3.6E-07	8.2E-08	7.2E-08	4.8E-09	2.5E-06	5.8E-08	2.4E-06
65	8.0E-07	3.0E-05	3.0E-07	6.8E-08	6.0E-08	4.0E-09	2.1E-06	4.8E-08	2.0E-06
66	6.7E-07	2.5E-05	2.5E-07	5.7E-08	5.0E-08	3.4E-09	1.7E-06	4.0E-08	1.7E-06
67	5.6E-07	2.1E-05	2.1E-07	4.7E-08	4.2E-08	2.8E-09	1.4E-06	3.4E-08	1.4E-06
68	4.7E-07	1.8E-05	1.8E-07	4.0E-08	3.5E-08	2.3E-09	1.2E-06	2.8E-08	1.2E-06
69	3.9E-07	1.5E-05	1.5E-07	3.3E-08	2.9E-08	2.0E-09	1.0E-06	2.3E-08	9.8E-07
70	3.3E-07	1.2E-05	1.2E-07	2.8E-08	2.4E-08	1.6E-09	8.4E-07	2.0E-08	8.2E-07
71	2.7E-07	1.0E-05	1.0E-07	2.3E-08	2.0E-08	1.4E-09	7.0E-07	1.6E-08	6.8E-07
72	2.3E-07	8.6E-06	8.6E-08	1.9E-08	1.7E-08	1.1E-09	5.9E-07	1.4E-08	5.7E-07
73	1.9E-07	7.2E-06	7.2E-08	1.6E-08	1.4E-08	9.5E-10	4.9E-07	1.1E-08	4.8E-07
74	1.6E-07	6.0E-06	6.0E-08	1.3E-08	1.2E-08	7.9E-10	4.1E-07	9.5E-09	4.0E-07
75	1.3E-07	5.0E-06	5.0E-08	1.1E-08	9.9E-09	6.6E-10	3.4E-07	8.0E-09	3.3E-07
76	1.1E-07	4.2E-06	4.2E-08	9.4E-09	8.2E-09	5.5E-10	2.9E-07	6.6E-09	2.8E-07
77	9.3E-08	3.5E-06	3.5E-08	7.8E-09	6.9E-09	4.6E-10	2.4E-07	5.6E-09	2.3E-07
78	7.7E-08	2.9E-06	2.9E-08	6.5E-09	5.7E-09	3.9E-10	2.0E-07	4.6E-09	1.9E-07
79	6.5E-08	2.4E-06	2.4E-08	5.5E-09	4.8E-09	3.2E-10	1.7E-07	3.9E-09	1.6E-07

80	5.4E-08	2.0E-06	2.0E-08	4.6E-09	4.0E-09	2.7E-10	1.4E-07	3.2E-09	1.3E-07
81	4.5E-08	1.7E-06	1.7E-08	3.8E-09	3.3E-09	2.3E-10	1.2E-07	2.7E-09	1.1E-07
82	3.8E-08	1.4E-06	1.4E-08	3.2E-09	2.8E-09	1.9E-10	9.7E-08	2.3E-09	9.4E-08
83	3.1E-08	1.2E-06	1.2E-08	2.7E-09	2.3E-09	1.6E-10	8.1E-08	1.9E-09	7.9E-08
84	2.6E-08	9.9E-07	9.9E-09	2.2E-09	2.0E-09	1.3E-10	6.8E-08	1.6E-09	6.6E-08
85	2.2E-08	8.3E-07	8.3E-09	1.9E-09	1.6E-09	1.1E-10	5.6E-08	1.3E-09	5.5E-08
86	1.8E-08	6.9E-07	6.9E-09	1.5E-09	1.4E-09	9.2E-11	4.7E-08	1.1E-09	4.6E-08
87	1.5E-08	5.8E-07	5.8E-09	1.3E-09	1.1E-09	7.7E-11	3.9E-08	9.2E-10	3.8E-08
88	1.3E-08	4.8E-07	4.8E-09	1.1E-09	9.5E-10	6.4E-11	3.3E-08	7.7E-10	3.2E-08
89	1.1E-08	4.0E-07	4.0E-09	9.0E-10	7.9E-10	5.3E-11	2.8E-08	6.4E-10	2.7E-08
90	8.9E-09	3.4E-07	3.4E-09	7.6E-10	6.6E-10	4.5E-11	2.3E-08	5.4E-10	2.2E-08
91	7.5E-09	2.8E-07	2.8E-09	6.3E-10	5.5E-10	3.7E-11	1.9E-08	4.5E-10	1.9E-08
92	6.2E-09	2.3E-07	2.3E-09	5.3E-10	4.6E-10	3.1E-11	1.6E-08	3.7E-10	1.6E-08
93	5.2E-09	2.0E-07	2.0E-09	4.4E-10	3.9E-10	2.6E-11	1.3E-08	3.1E-10	1.3E-08
94	4.4E-09	1.6E-07	1.6E-09	3.7E-10	3.2E-10	2.2E-11	1.1E-08	2.6E-10	1.1E-08
95	3.6E-09	1.4E-07	1.4E-09	3.1E-10	2.7E-10	1.8E-11	9.4E-09	2.2E-10	9.1E-09
96	3.0E-09	1.1E-07	1.1E-09	2.6E-10	2.3E-10	1.5E-11	7.8E-09	1.8E-10	7.6E-09
97	2.5E-09	9.6E-08	9.6E-10	2.2E-10	1.9E-10	1.3E-11	6.5E-09	1.5E-10	6.4E-09
98	2.1E-09	8.0E-08	8.0E-10	1.8E-10	1.6E-10	1.1E-11	5.5E-09	1.3E-10	5.3E-09
99	1.8E-09	6.7E-08	6.7E-10	1.5E-10	1.3E-10	8.9E-12	4.6E-09	1.1E-10	4.4E-09
100	1.5E-09	5.6E-08	5.6E-10	1.3E-10	1.1E-10	7.4E-12	3.8E-09	8.9E-11	3.7E-09

APPENDIX 3

Redox Model

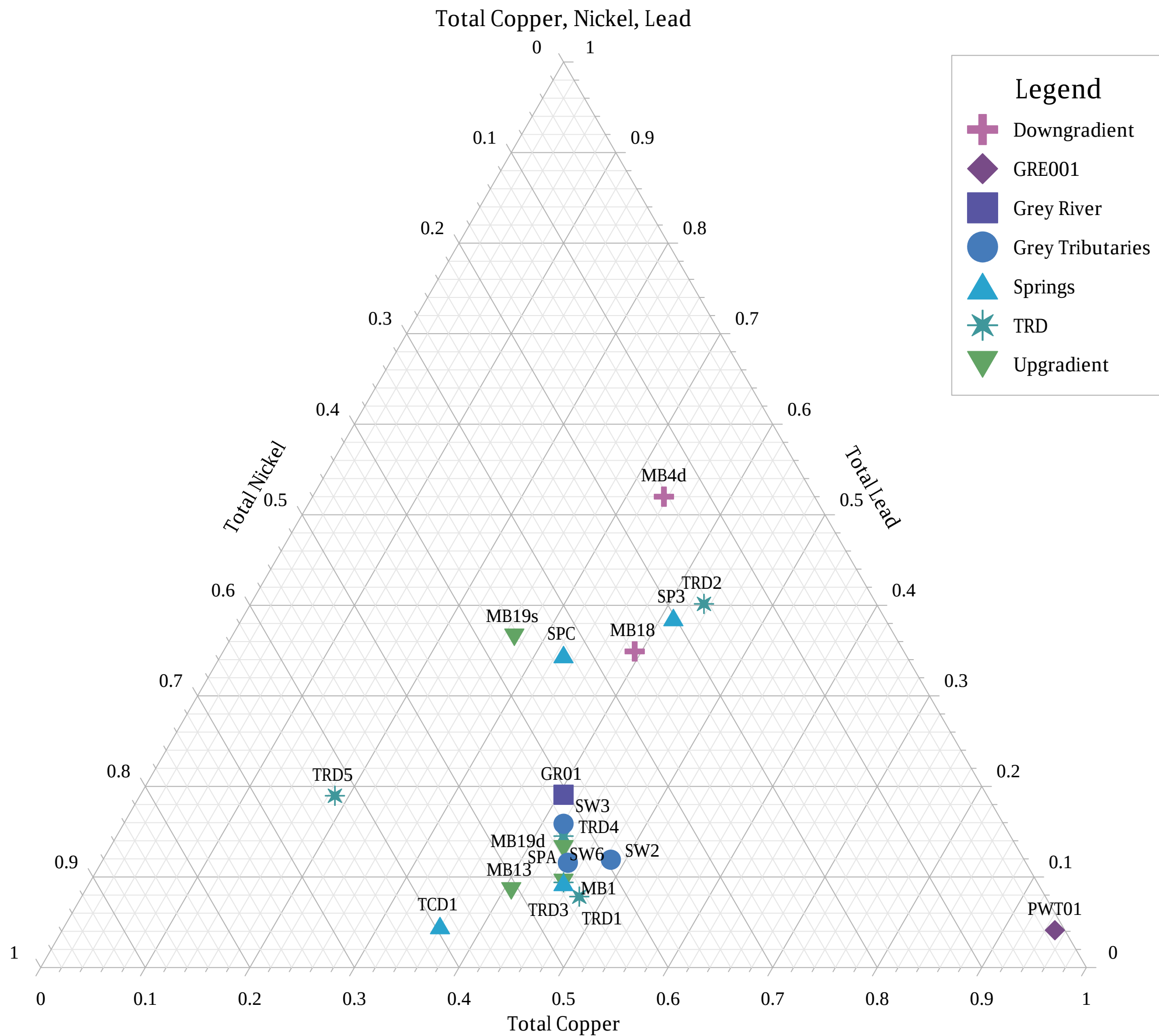
Redox Assignment

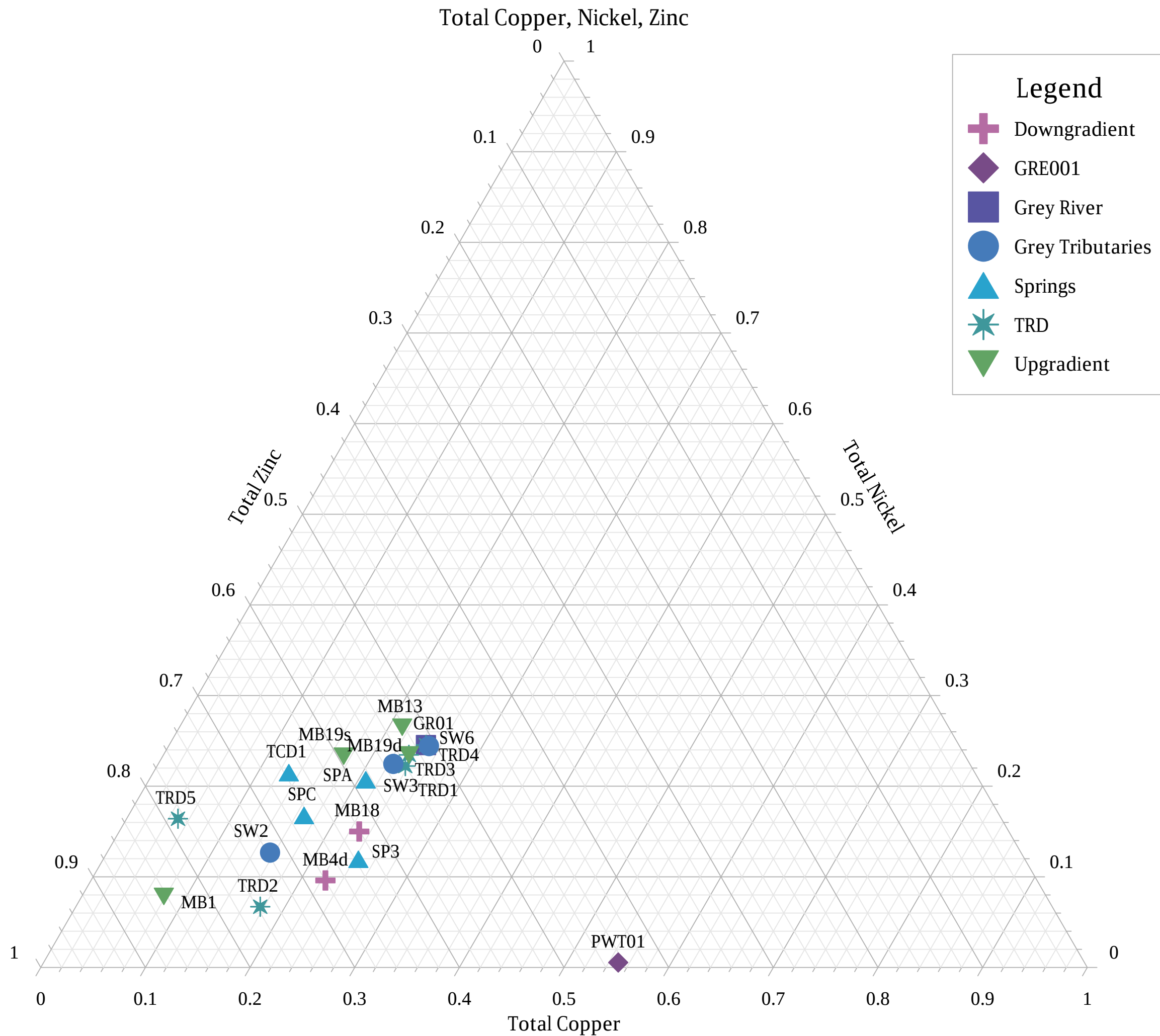
Sample ID	Redox	Redox Variables	Dissolved O ₂	NO ₃ ⁻ (as Nitrogen)	Mn ²⁺	Fe ²⁺	SO ₄ ²⁻	Num of Params	General Redox Category	Redox Process
	mV	Units Threshold values	millig/L ▼	millig/L ▼	microg/L ▼	microg/L ▼	millig/L ▼			
			0.5	0.5	50	100	0.5			
MB4d	144		2.96	1.05	0.028	0.04	12.3	5	Oxic	O2
MB13	-17		0.23	0.064	0.89	2.7	2.8	5	Suboxic	Suboxic
MB18	6		0.76	0.002	2.3	0.85	155	5	Oxic	O2

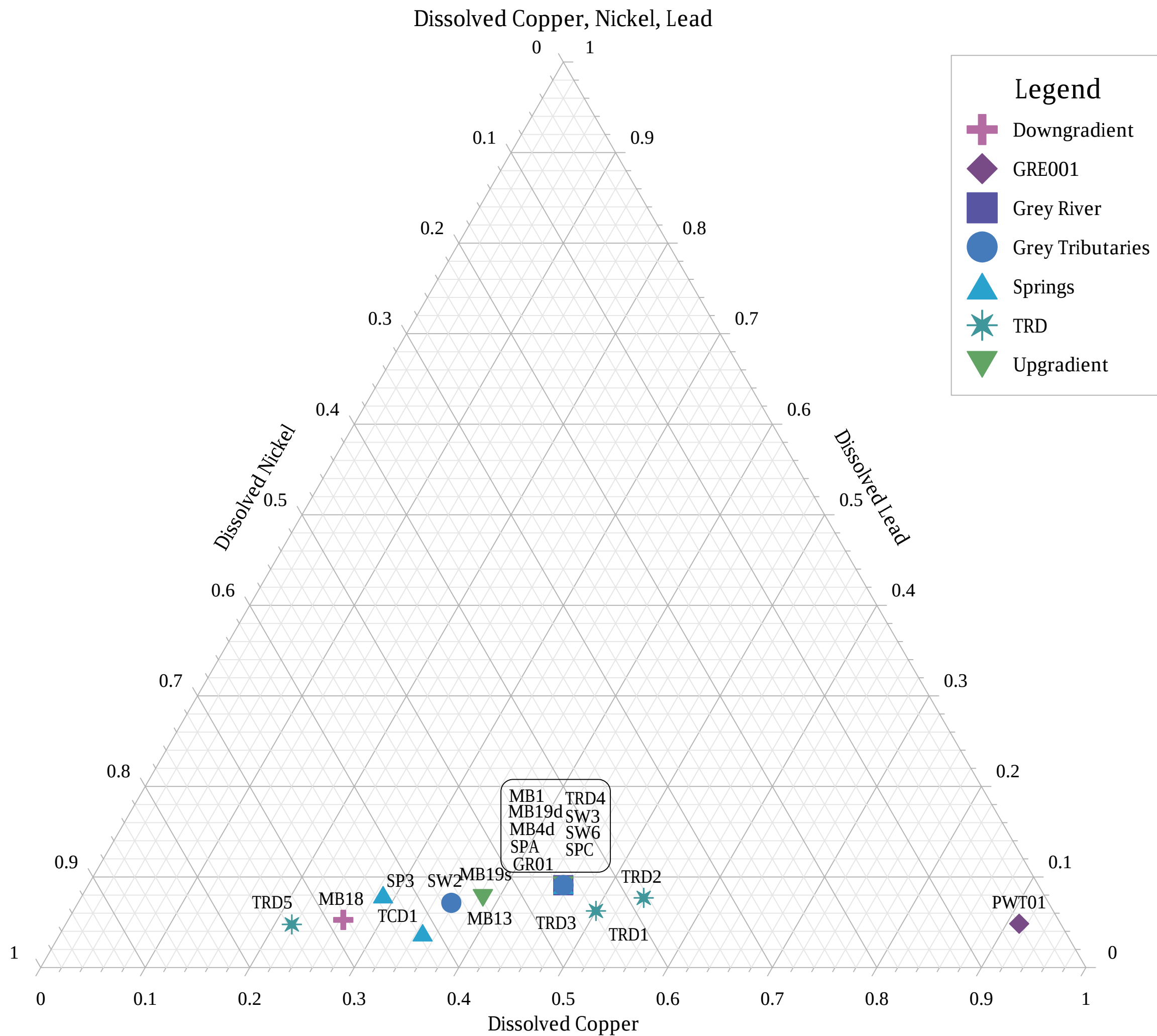
Modified from Jurgens, B.C., McMahon, P.B., Chapelle, F.H., and Eberts, S.M., 2009, An Excel® workbook for identifying redox processes in ground water: U.S. Geological Survey Open-File Report 2009–1004 8 p.

APPENDIX 4

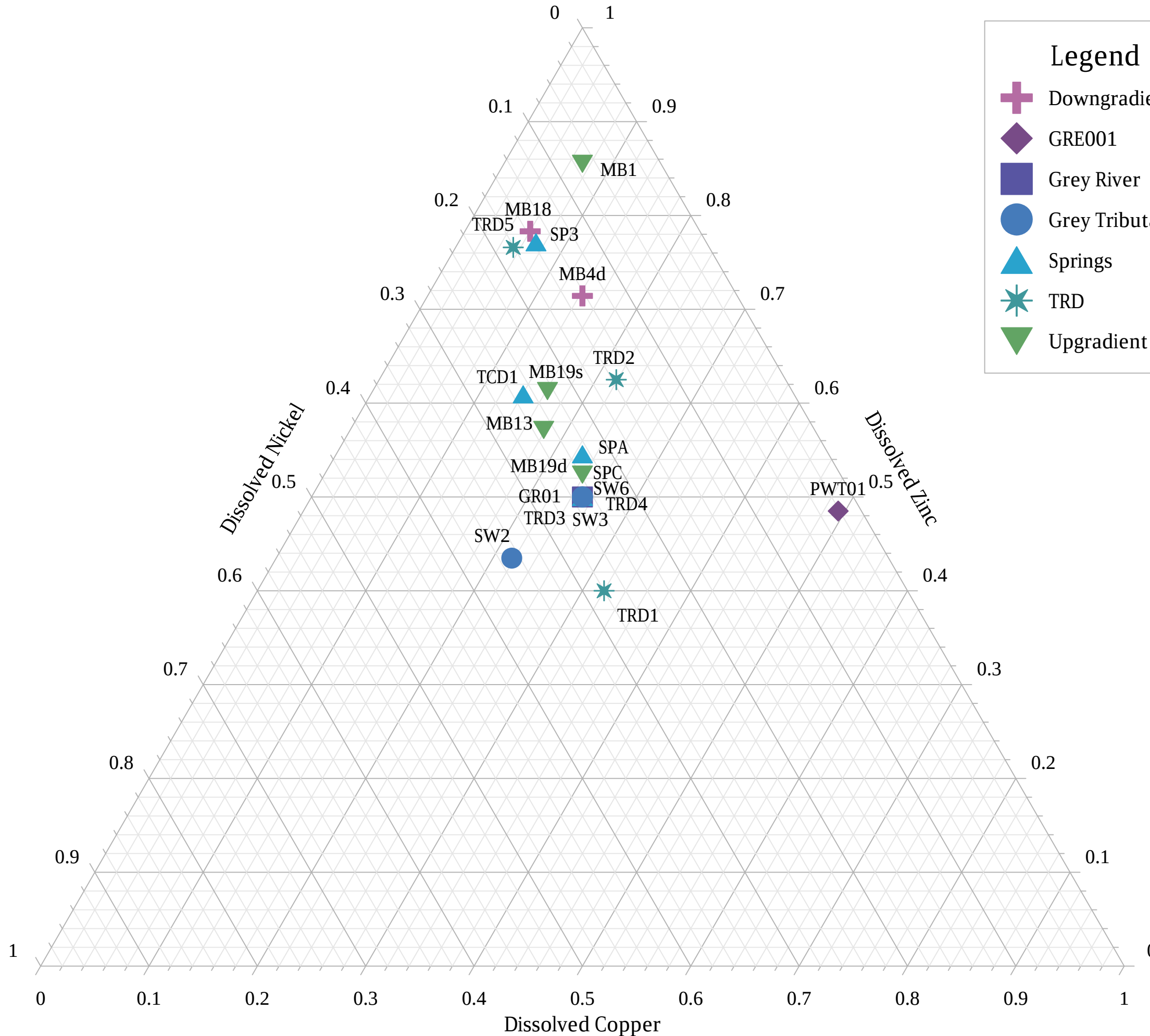
Heavy Metal Concentration Graphs



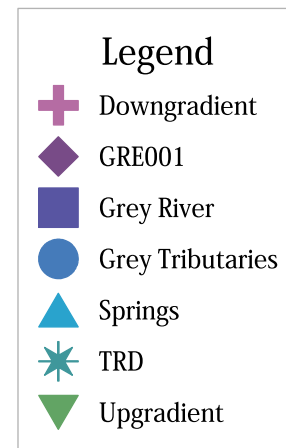
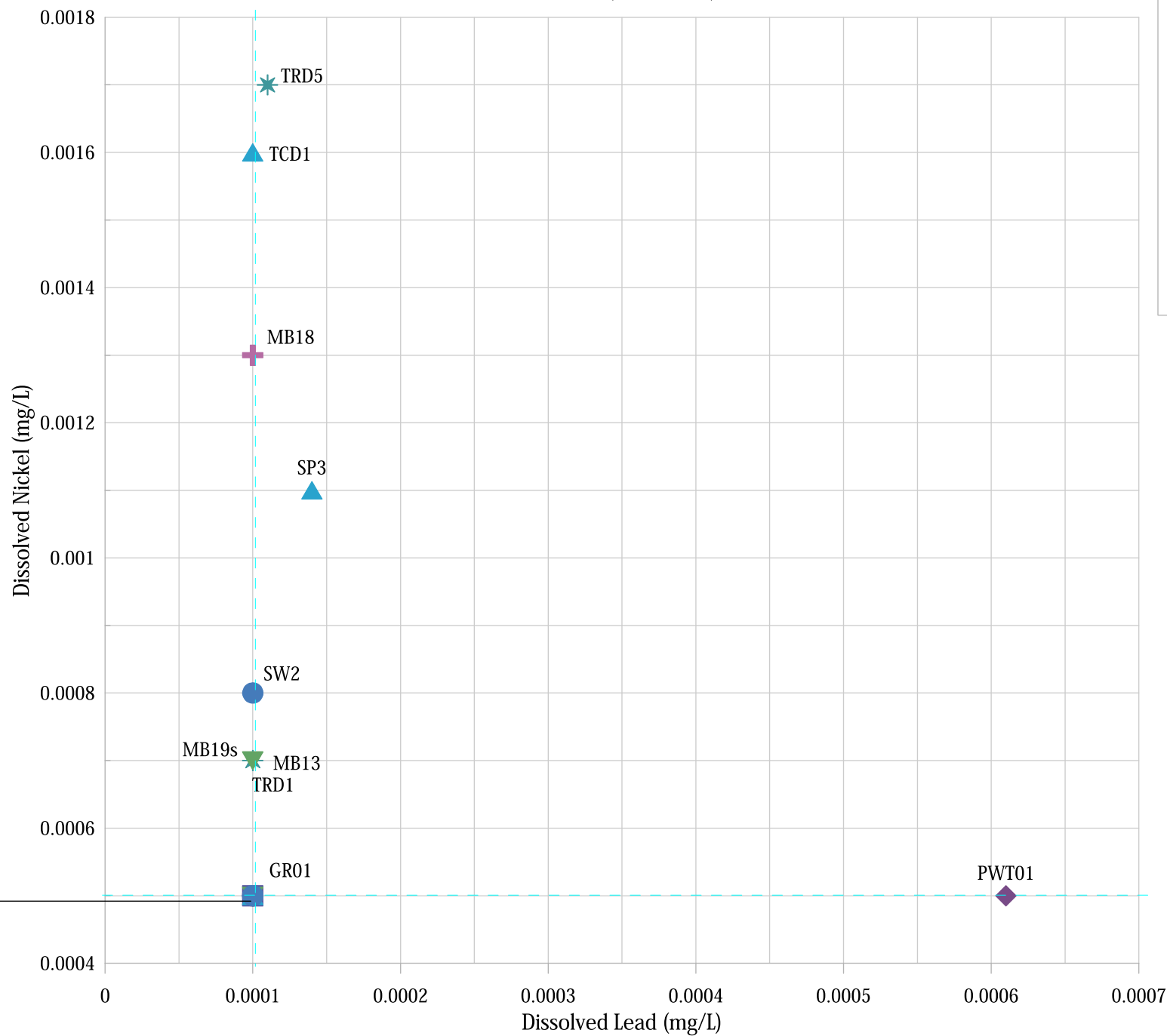




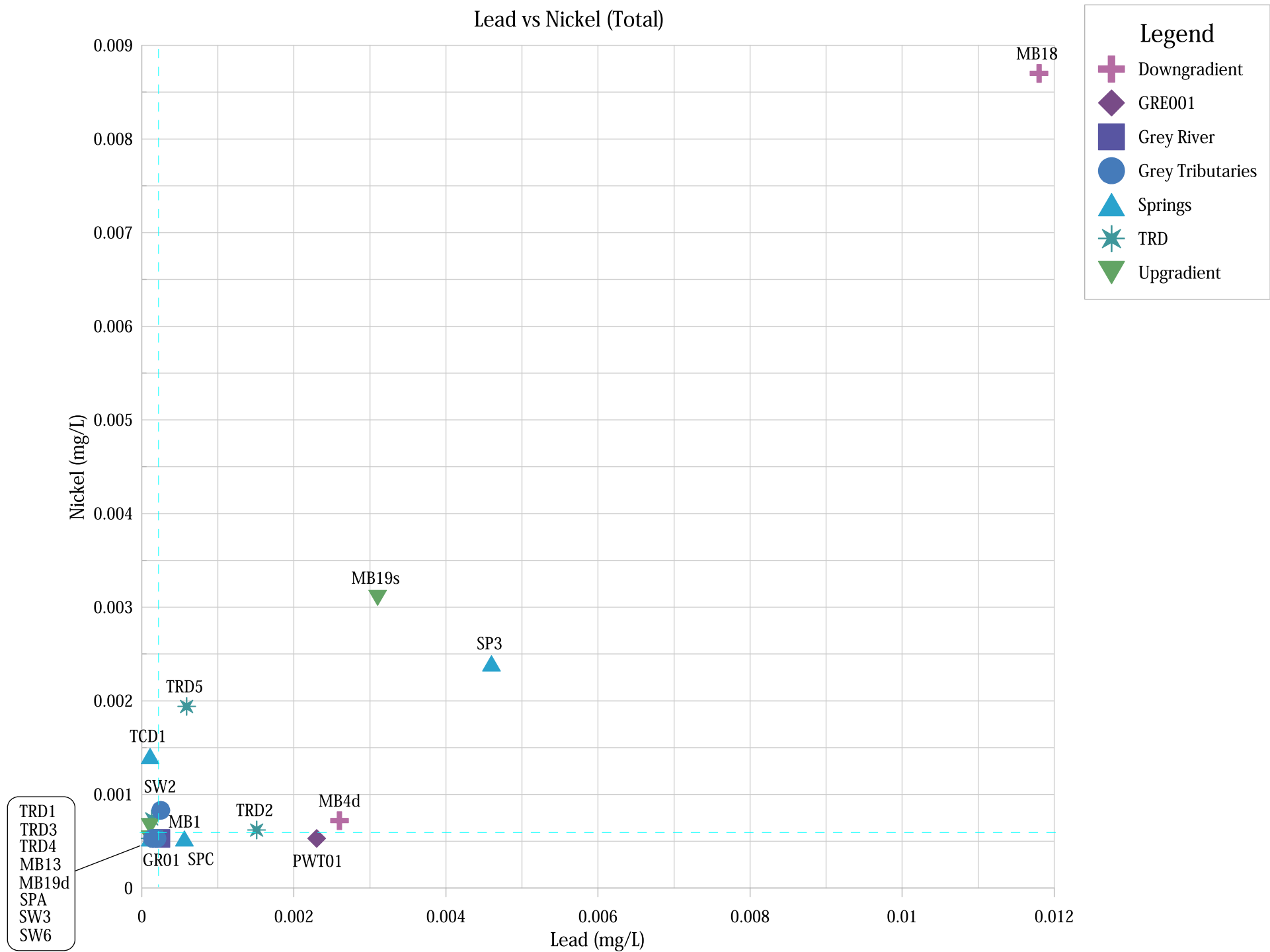
Dissolved Copper, Nickel, Zinc

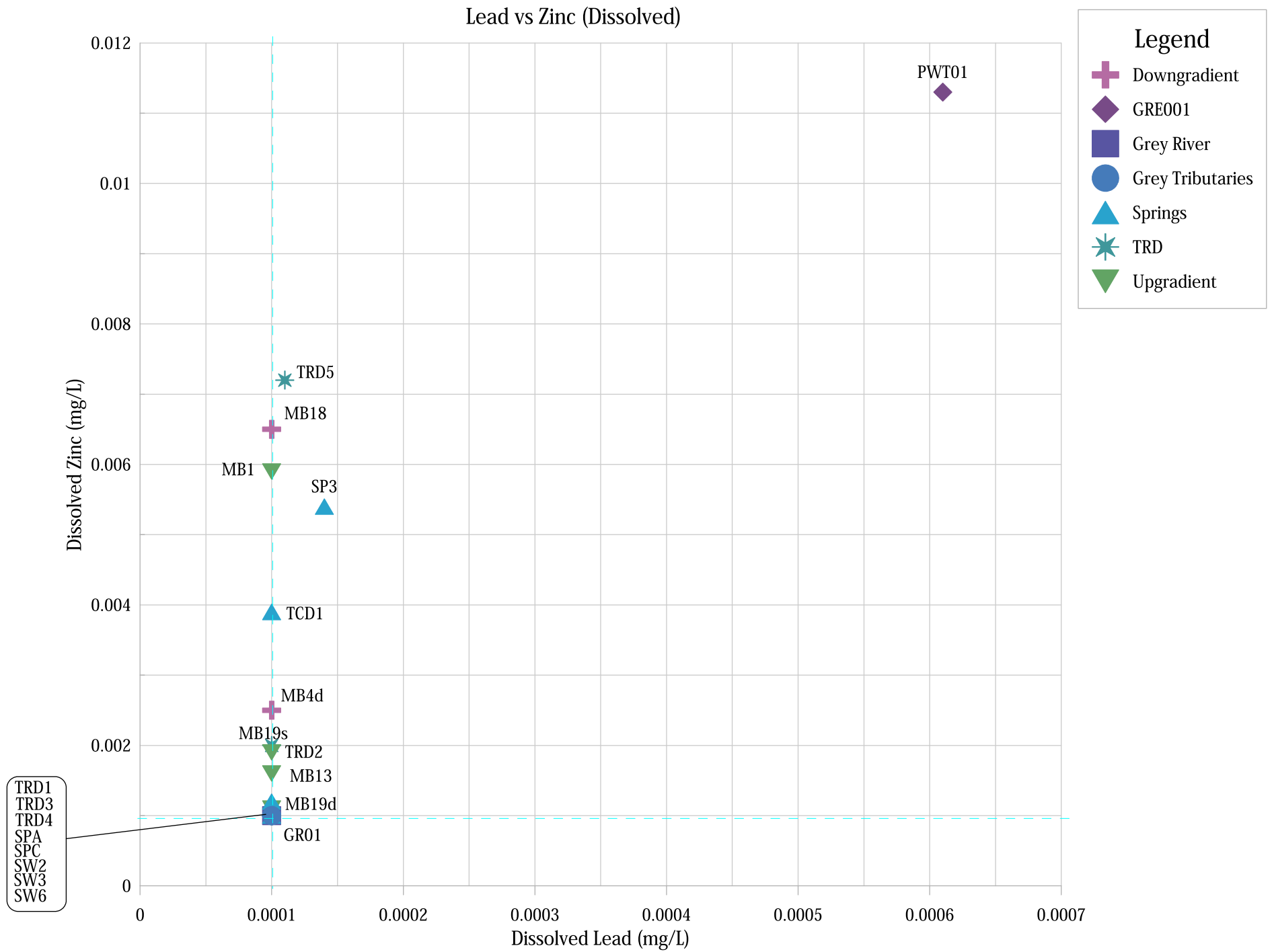


Lead vs Nickel (Dissolved)

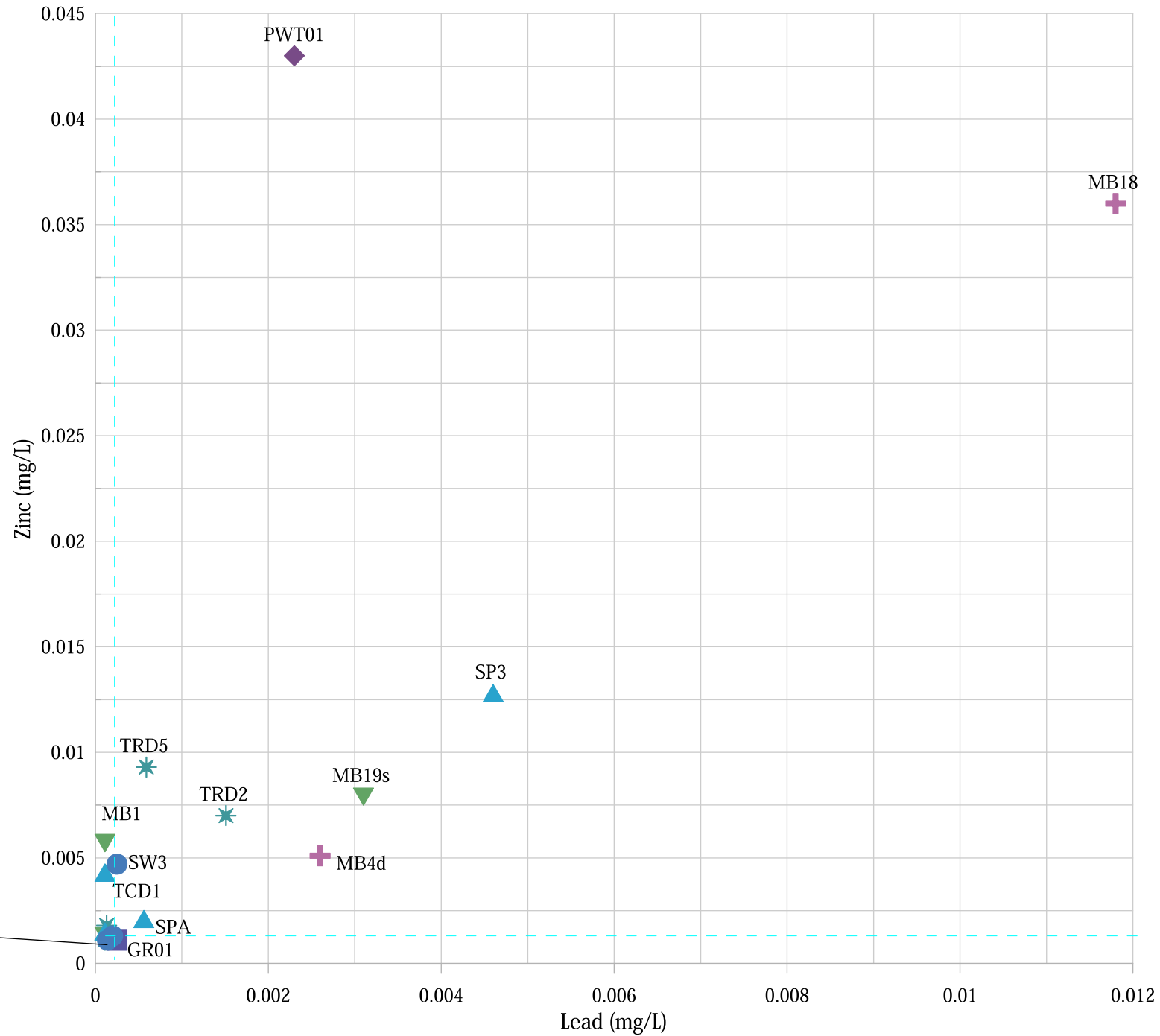


TRD2
TRD3
TRD4
MB1
MB4d
MB19d
SPA
SPC
SW3
SW6



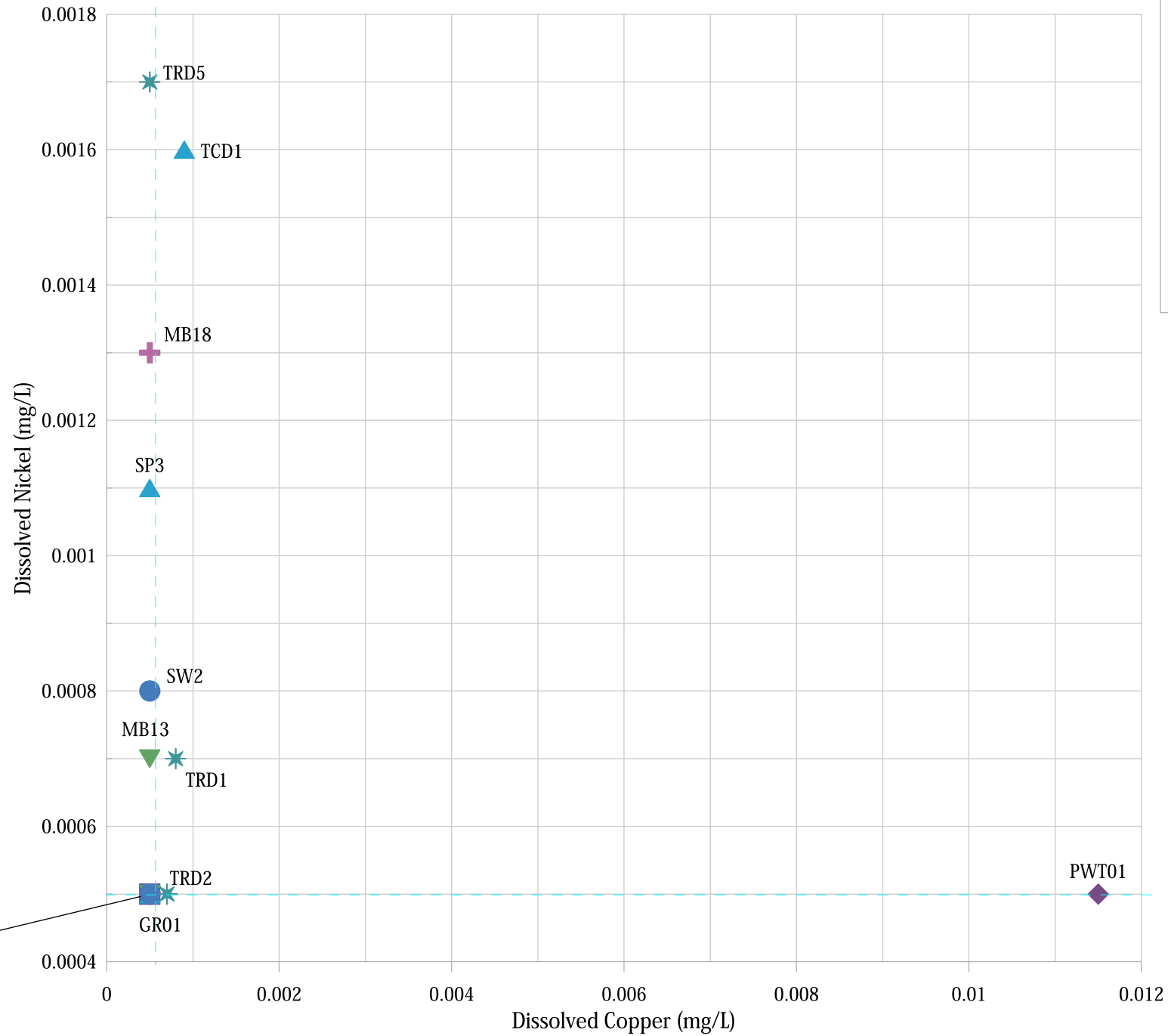


Lead vs Zinc (Total)



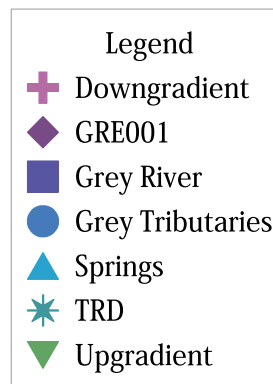
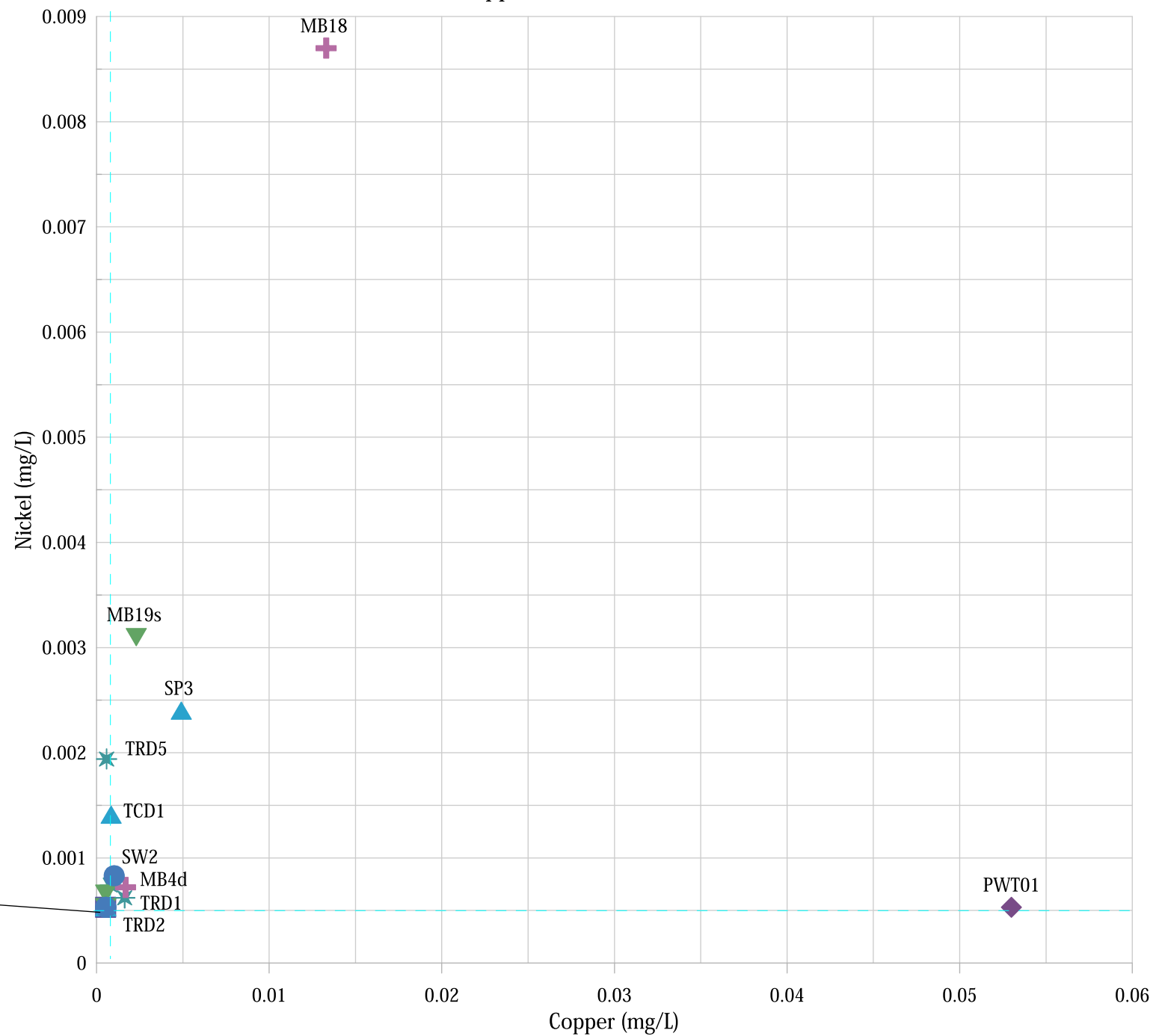
TRD1
TRD3
TRD4
MB13
MB19d
SPC
SW2
SW6

Copper vs Nickel (Dissolved)

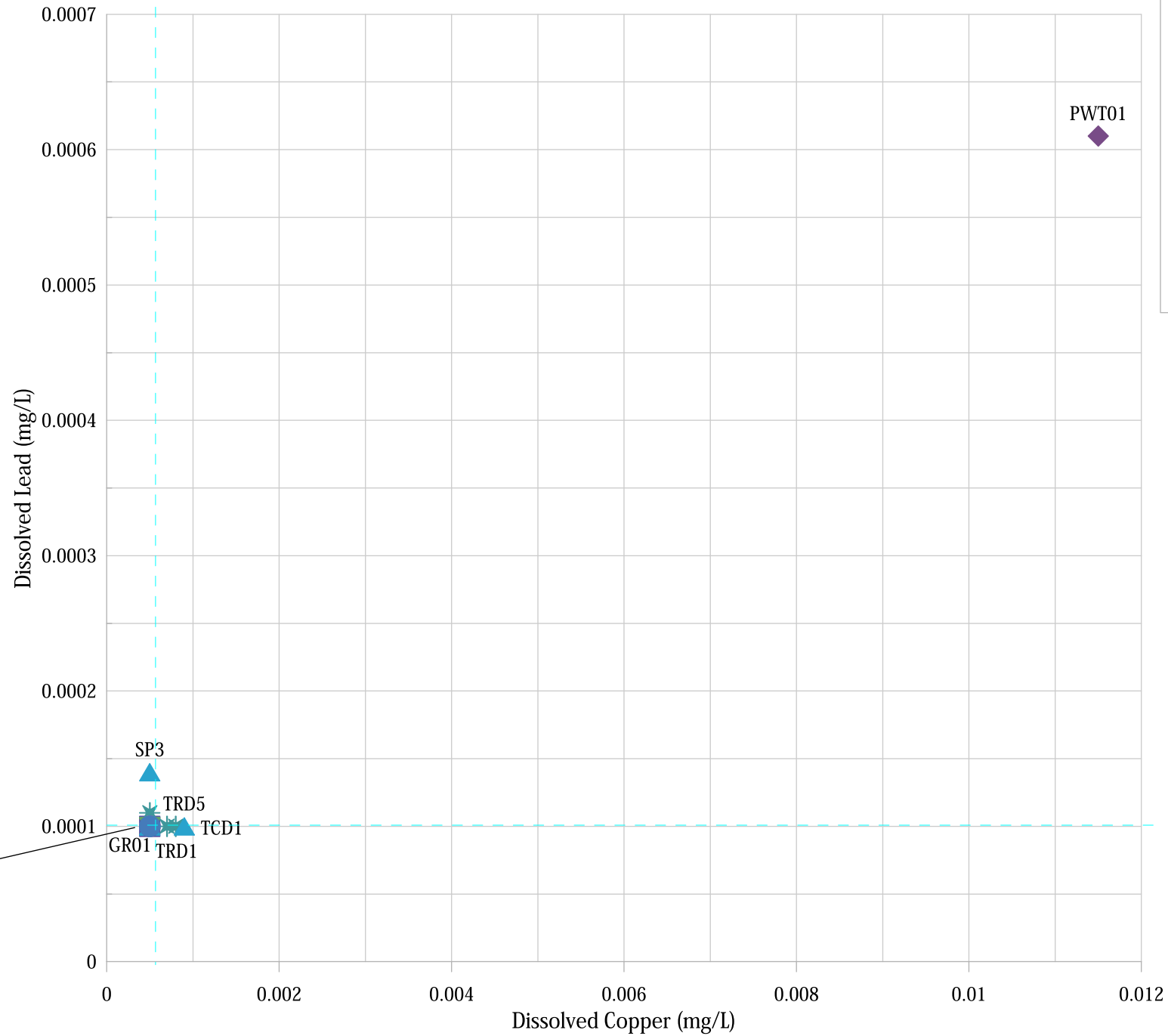


TRD3
TRD4
MB1
MB4d
MB19s
MB19d
SPA
SPC
SW3
SW6

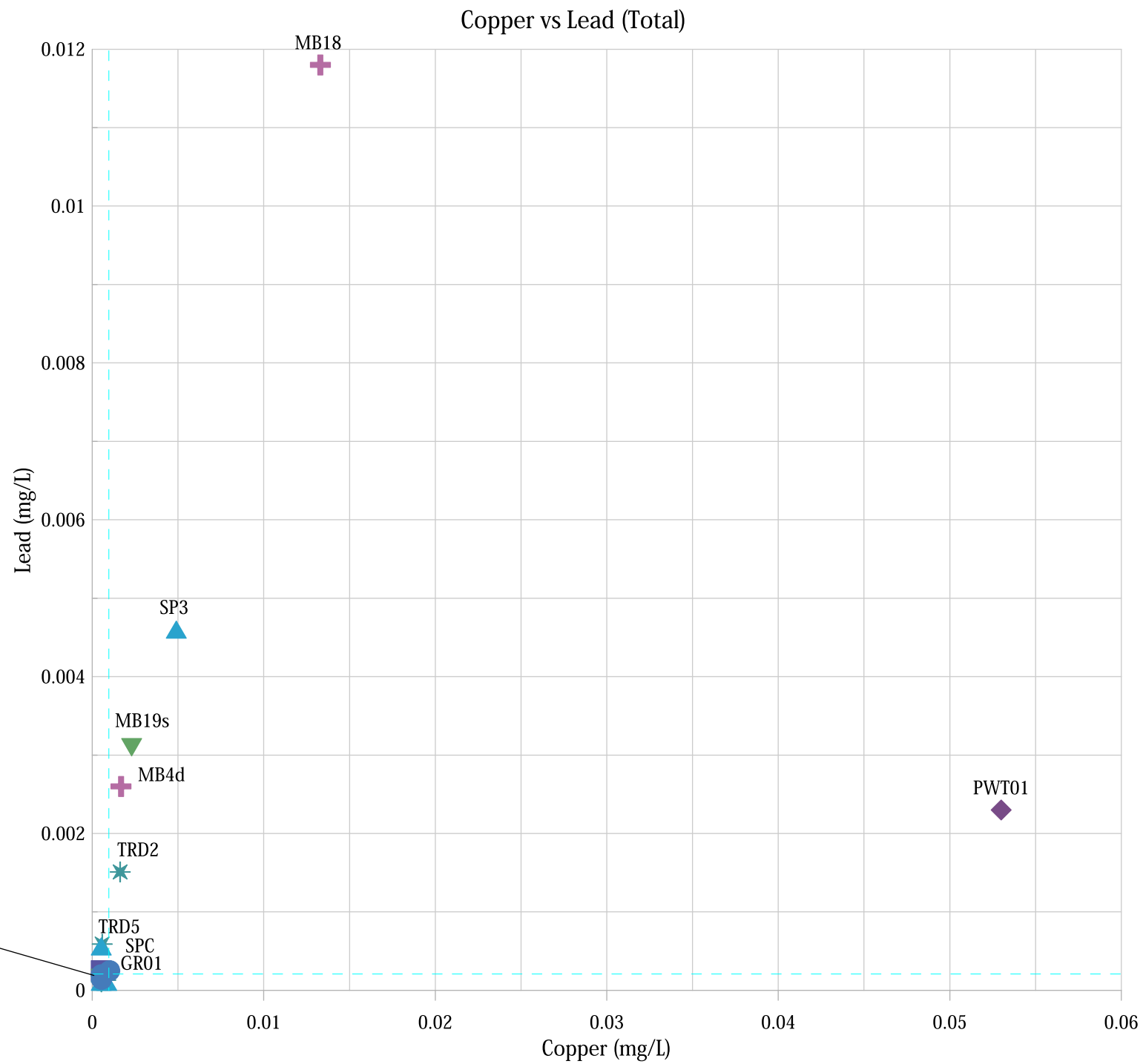
Copper vs Nickel (Total)



Copper vs Lead (Dissolved)

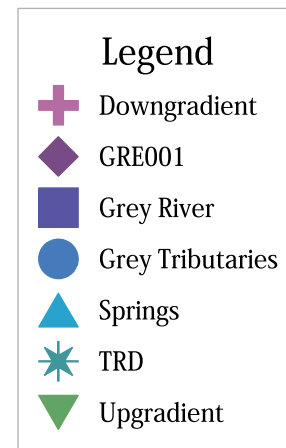
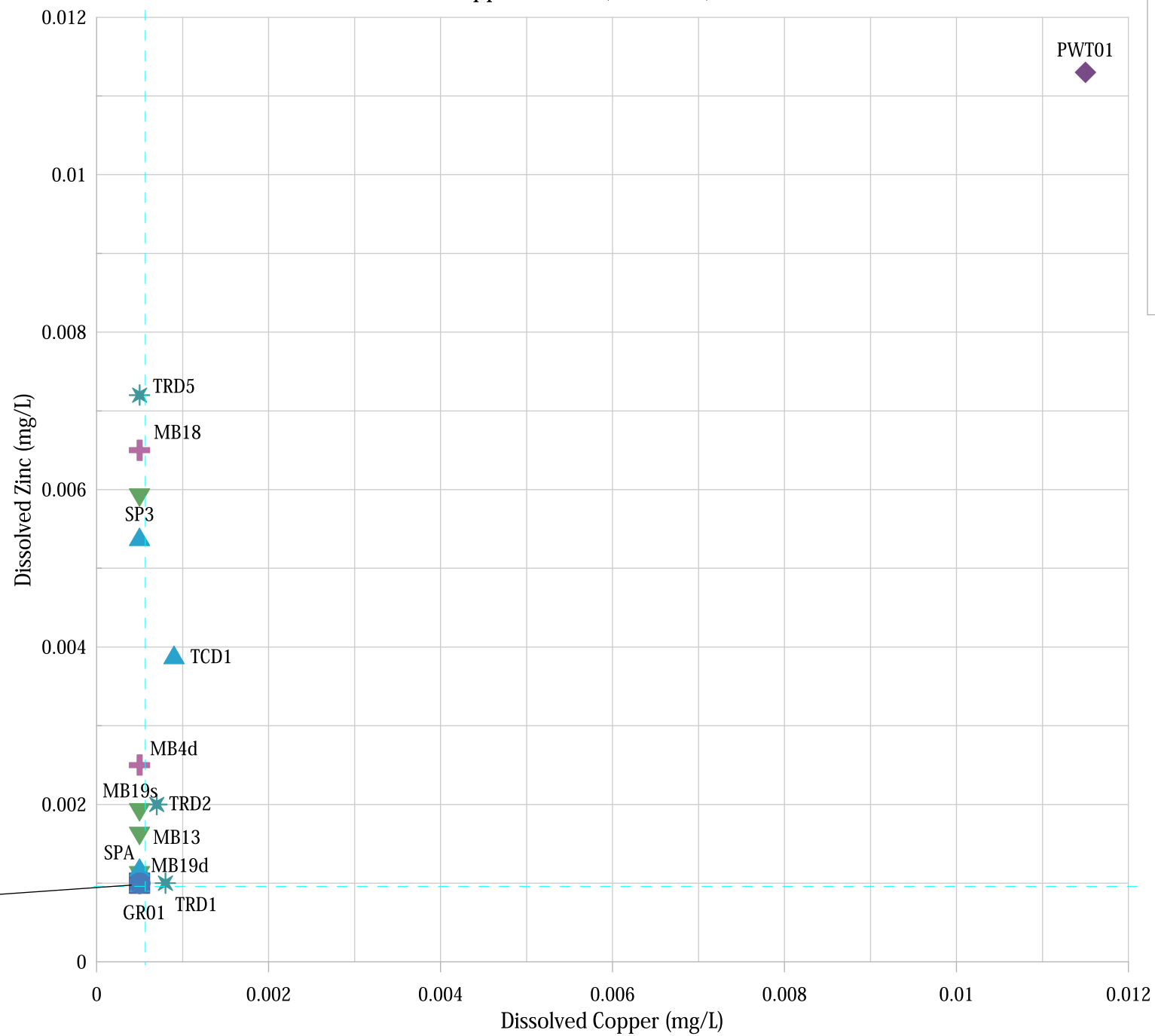


TRD2
TRD3
TRD4
MB1
MB4d
MB13
MB18
MB19s
MB19d
SPA
SPC
SW2
SW3
SW6

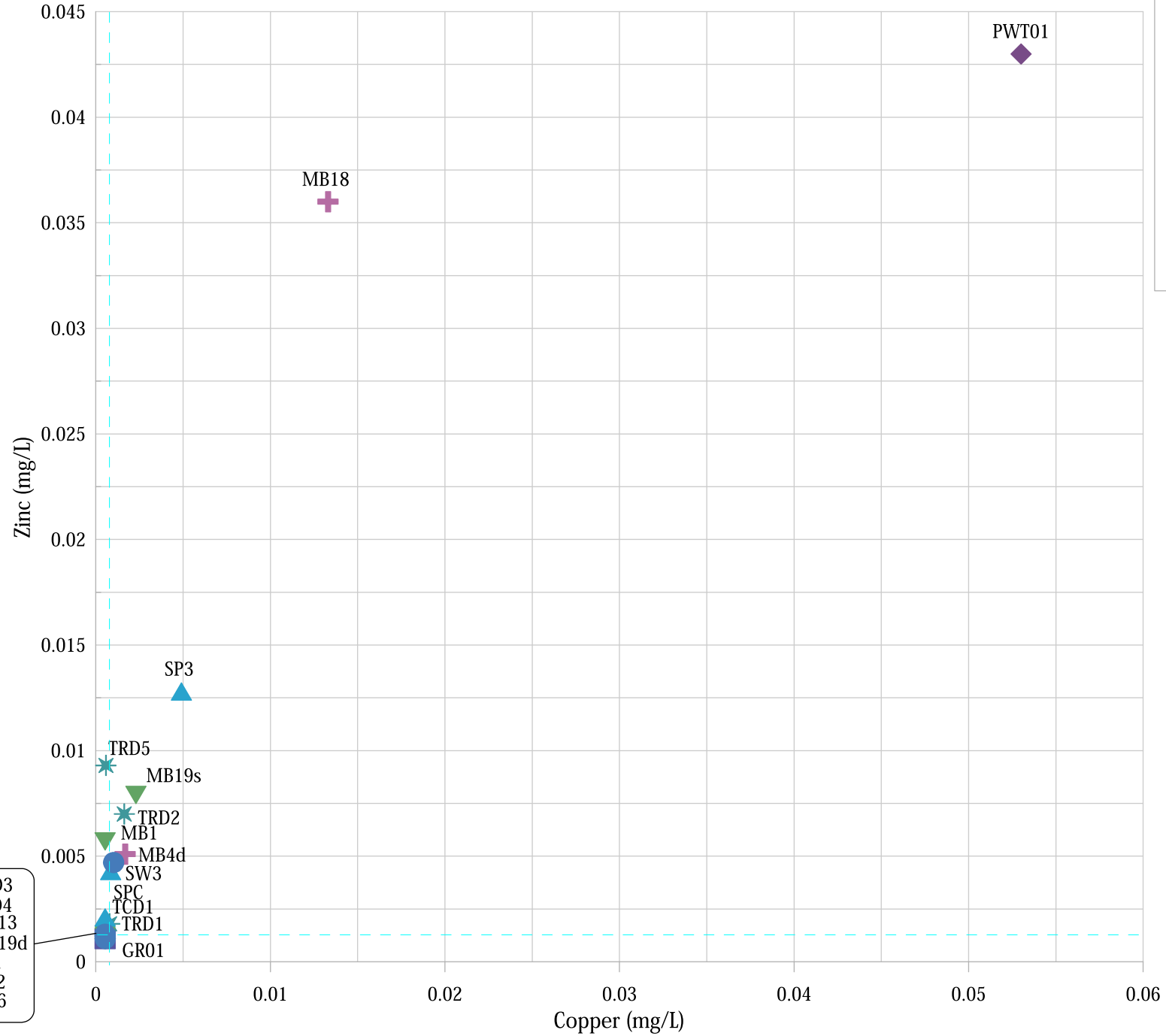


TRD1
TRD3
TRD4
MB1
MB13
MB19d
TCD1
SPA
SW2
SW3
SW6

Copper vs Zinc (Dissolved)



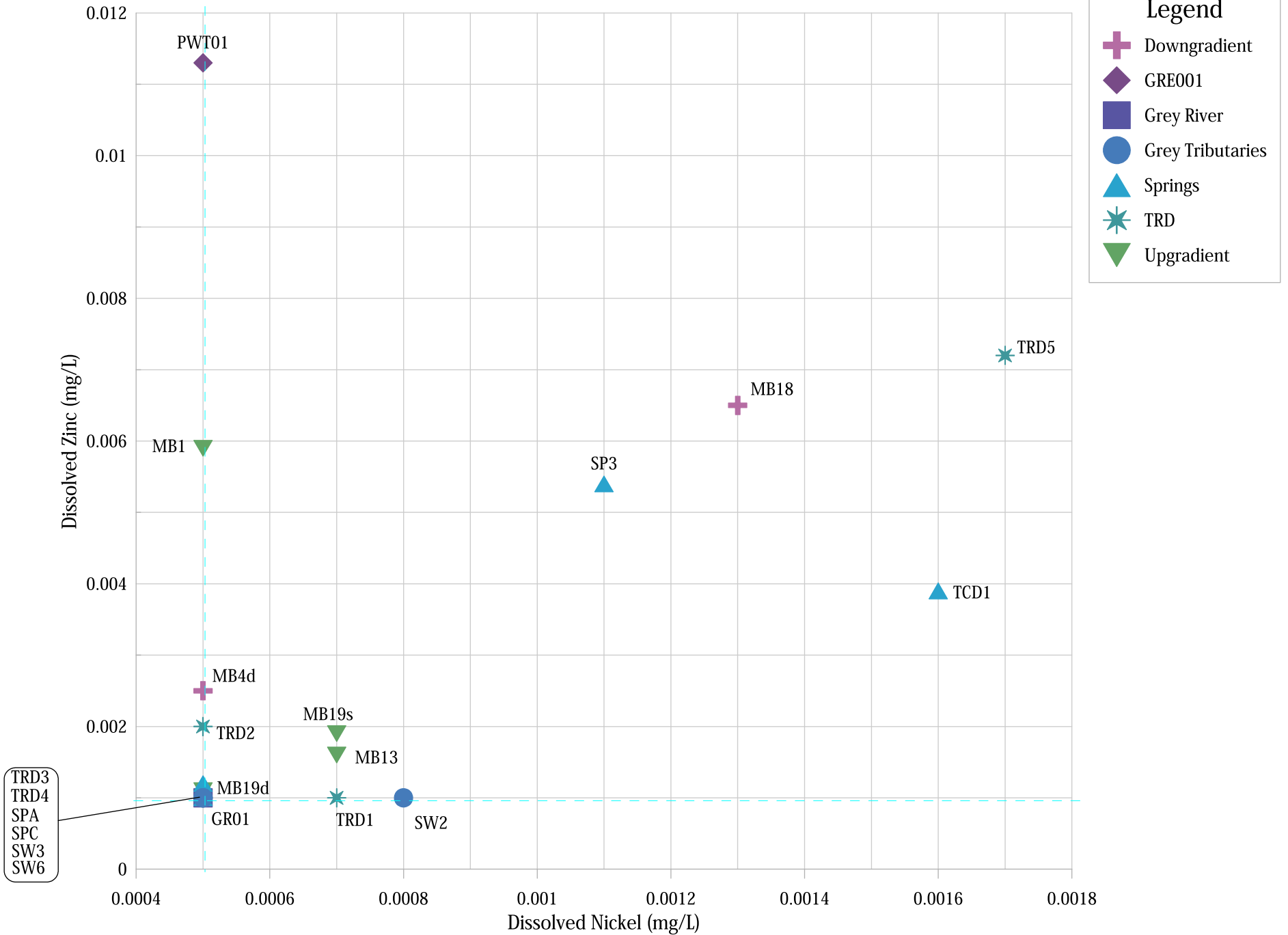
Copper vs Zinc (Total)



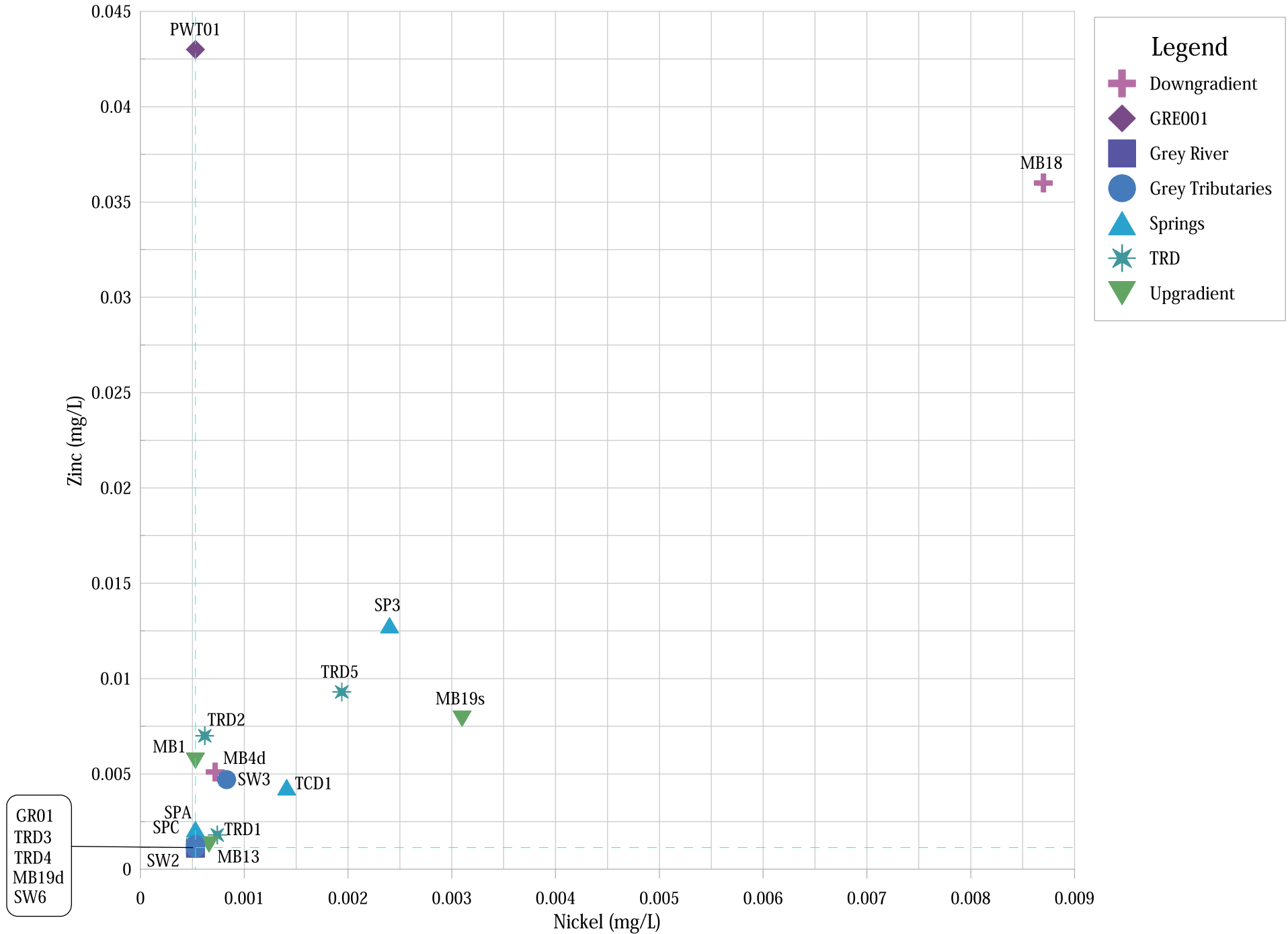
Legend

- + Downgradient
- ◆ GRE001
- Grey River
- Grey Tributaries
- ▲ Springs
- ✱ TRD
- ▼ Upgradient

Nickel vs Zinc (Dissolved)



Nickel vs Zinc (Total)





memorandum



TO	Mike Durand	FROM	Duncan Gray
	Planz	DATE	20 May 2025
RE	Taylorville Resource Park Limited Short-term Consent Application – Ecological Review		

1.0 Introduction

1.1 Context

Bioresearches was engaged by ENGEO, on behalf of their client Taylorville Resource Park Limited (TRPL), to undertake an Ecological Impact Assessment (EcIA) for the temporary discharge of leachate contaminated water to land (RC-2024-0122). The Bioresearches assessment is presented in a memo dated 7 February 2025. That ecological assessment was in response to a s92 RMA request sent by the West Coast Regional Council (WCRC) (WCRC letter dated 19 December 2024) to address questions raised during consent processing.

This memorandum is a review of the Bioresearches EcIA and consideration of the potential ecological effects of the TRPL proposal. To inform my review I have considered the information contained within the ENGEO Preliminary Hydrogeological Assessment (the original effects assessment provided with the resource consent application dated 19 September 2024), a later ENGEO memo detailing water quality results from samples collected in January 2025, images created by EHS Support (relevant to previous consent applications for the site) and conversations with Carl Steffens (PDP), whom has undertaken a formal review of hydrogeological issues relevant to the application on behalf of WCRC.

1.2 Assessment Overview

The Bioresearches EcIA is a brief document based on a walk-over or images provided by a third party. It is not apparent that Bioresearches visited the site or what assessments were undertaken. Bioresearches Figure 1 (provided below) indicates that macroinvertebrates and fish were surveyed, but that information is not presented in the document. Throughout there are statements made about values that have little to no basis or supporting material and are counter to my own interpretations of the information provided.

1.3 Site Layout

I note that the alignment of streams shown by Bioresearches (Figure 1) is incorrect and I have included what appears to be a correct interpretation of waterways presented in an EHS Support report from 2024 (this document Figure 2).

Noname Creek drains a gully to the north and north-west of the landfill site, crosses under Taylorville Road and drops down to the Grey River flats. There it is joined by Grey River tributary which appears to arise on the flats at the base of a bank below Taylorville Road. The Taylorville Road roadside drain is the only obvious headwater to the Grey River tributary (via a culvert passing under the road). The stream is assumed to be otherwise groundwater fed along the flats.

The Roadside drain collects water from the western edge of the landfill and along the adjacent access road to the neighbour's property, in addition to receiving groundwater seepage via a spring at the head of the drain (western boundary of the landfill). At the intersection of the neighbour's driveway and Taylorville Road (west of the landfill) the roadside drain also collects water from along the Taylorville Road and flows to the east. However, as shown in Figure 2 below, further east along the road the direction of flow in the roadside drain reverses on two occasions. Drain water flowing in opposite directions converges at culverts shown at the locations of TRD5 and TRD2 in Figure 2. The culvert at TRD5 drains water under the road into the Grey River tributary as described above, while the culvert at TRD2 transfers water to another stream situated on the lower terrace surface below the road and which drains land to the east and south of the Grey District Council (GDC) Coal Creek water treatment plant. For the purpose of this review the stream to the east of the GDC treatment plant is referred to as the Noname Creek 2.

The Noname Creek 2 and associated drains collect water from an eastern catchment but also receive water from the roadside drain directly below the southern flanks of the landfill terrace via the culvert at TRD2. This section of the roadside drain, in addition to roadside drainage further east (that crosses into the Noname Creek 2 via an additional culvert at TRD4 as seen in Figure 2) is also close to the proposed TRPL discharge to land site (Figure 3). The Noname Creek 2 ultimately meanders across the flats on the floodplain of the Grey River in the vicinity of the GDC Coal Creek water supply bores before discharging to the Grey River. I note the ecological value of the Noname Creek 2 has not been assessed despite being in relatively close proximity to potential groundwater and surface water flow paths that could be affected by the proposed discharge to land.

Based on the descriptions provided above, we note that Bioresearches Figure 1 indicates that the Noname Creek 2 and the Grey River tributary are connected and the lower section of the Noname Creek 2 on the Grey River floodplain are also not shown in the Bioresearches figure. In addition, the Bioresearches figure doesn't show the culverts at the locations of TRD2 and TRD4 that provide a pathway for drain water on the northern side of Taylorville Road to enter into the Noname Creek 2.

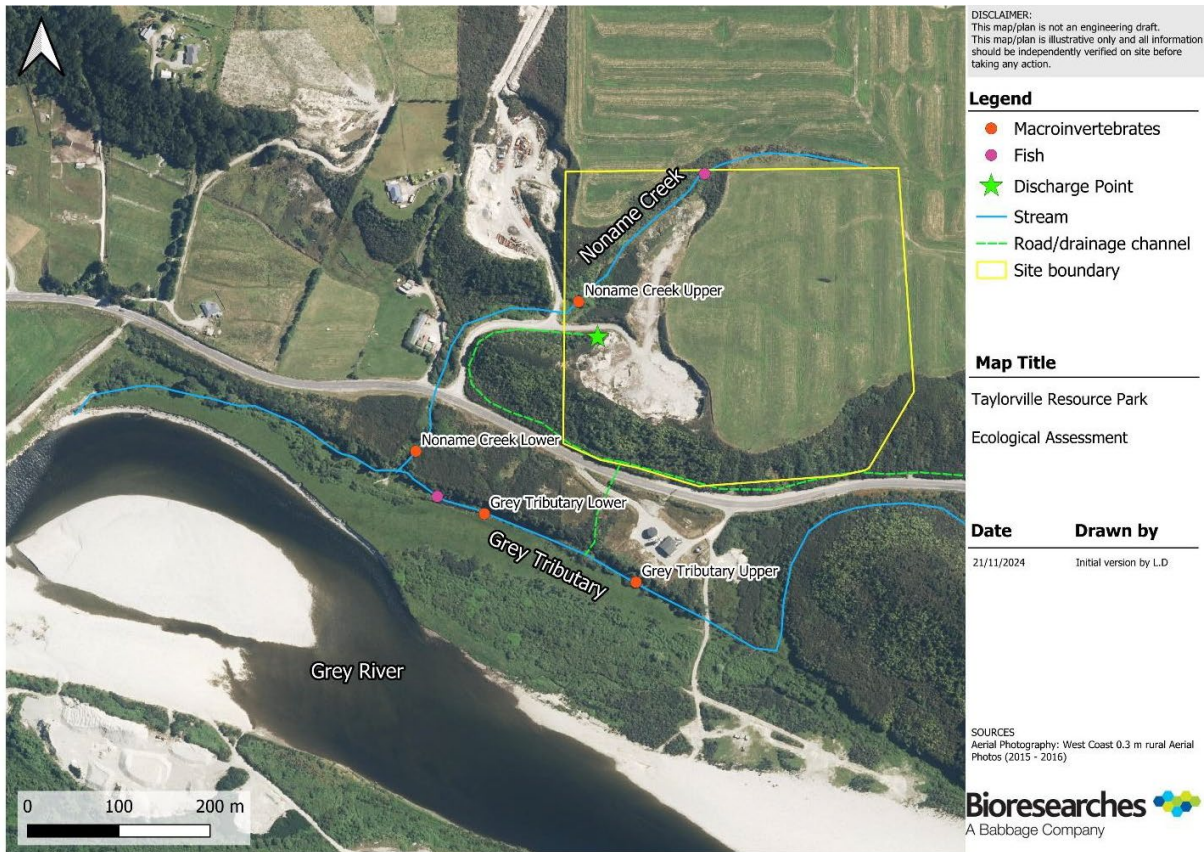


Figure 1. Bioresearchers Figure 1 showing an erroneous interpretation of the stream networks and locations of sampling sites that were not subsequently reported.



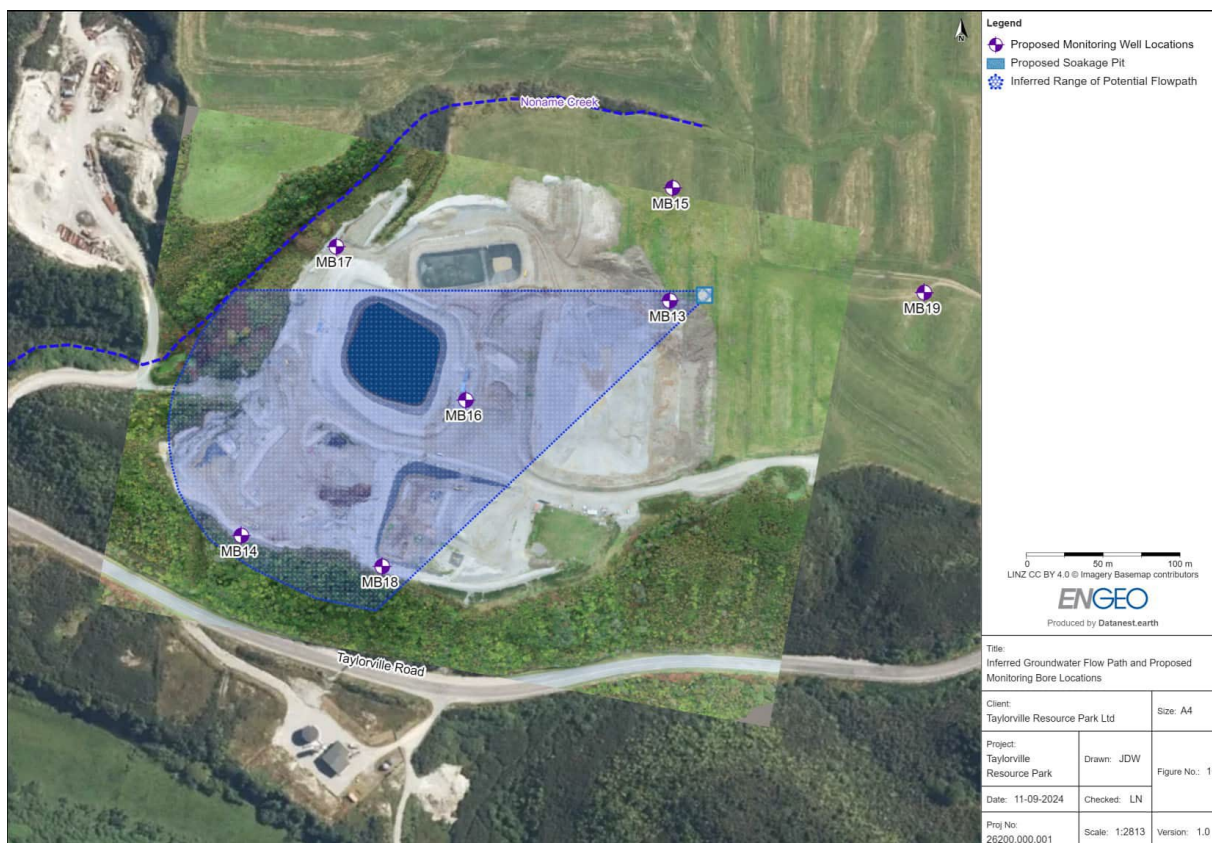


Figure 3. Location of proposed discharge of excess leachate and stormwater to land and inferred range of potential groundwater flowpath (ENGEO preliminary hydrogeological investigation).

2.0 Ecological Values

2.1 Noname Creek

Noname Creek is described by Bioresearches as a small waterway with a generally cobble / gravel bed and dense riparian vegetation that is a mix of native and exotics. Bioresearches describe some fine sediment being present on the bed, but no formal assessment of deposited sediment has been made to determine if that sediment would be likely to impact aquatic ecosystems. Based on the description I consider this stream would offer good habitat to a diversity of fish and macroinvertebrates.

Water quality results taken from Noname Creek are reviewed by ENGEO in their preliminary hydrogeological assessment and presented in Appendix 5 (Table A6.4 and 6.5) of that report. Those results indicated that in general the water in Noname Creek was clean (although the number of parameters sampled from early 2023 to mid-2024 were limited), albeit with consistently elevated Nitrate-N and occasional spikes in aluminium and zinc. The concentrations of Nitrate-N peaked at 1.44 mg/l at the upstream site (sample SPFUS) and such concentrations might contribute to the growth of undesirable algae in the stream where it is less shaded. However, in terms of toxicity the recorded concentrations are suitable for a moderately disturbed system with effects on 5 % of sensitive species (Hickey, 2013; MfE, 2023).

Subsequent information provided in an ENGEO memo dated 17 March 2025 provides updated water quality results for the Noname Creek. January 2025 results for Nitrate-N peaked at 0.25 mg/l in the downstream Noname Creek site. The ENGEO memo (March 2025) describes how elevated parameter concentrations in wastewater peaked during January 2023 and were likely due to the inappropriate deposition of a large volume of organic matter (tomato plants) in Cell C. We note this was the same time

that the concentrations of Nitrate-N peaked in Noname Creek. It is possible that the effects of that organic matter may have passed (or have temporarily reduced due to seasonal changes in groundwater and rainfall interaction within Cell C). Based on the above, the elevated nitrate concentrations (1.44 mg/l in January 2023) may not be representative of Noname Creek over the longer term. This indicates that Noname Creek water quality and ecological health, in terms of nitrate, may be better than indicated by the initial data derived from ENGEO in their preliminary hydrogeological assessment and that was used by Bioresarches.

Given the descriptions of water quality and habitat for the Noname Creek I am unable to agree with the description of the stream as having low ecological values or likely a depauperate fish community. The Noname Creek discharges to the Grey River within 5 km of the Tasman Sea. A large proportion of New Zealand native fish are diadromous meaning they must migrate between the ocean and freshwater to complete their lifecycle (McDowall, 1998). Consequently, the number of fish species likely to be found in a stream is strongly regulated by distance to the ocean.

In the absence of barriers to fish passage I would expect Noname Creek to provide habitat to a broader array of fish than described by Bioresarches, including some 'At Risk', 'Threatened' or culturally valuable species. In addition to eels (two species) and any of several bully species, I suggest that migratory and non-migratory galaxiids, lamprey and torrent fish could be present in the stream. The most likely additional inhabitants of the Noname Creek are the galaxiids. The Koaro (*Galaxias beivipinnis*) classified as 'At Risk – declining' (Dunn et al., 2018), banded kokopu (*Galaxias fasciatus*) classified as 'Not Threatened', and the shortjaw kokopu (*Galaxias postvectis*) classified as 'Threatened – nationally vulnerable' are all capable of ascending relatively steep obstacles to access streams. The giant kokopu (*Galaxias argenteus*) classified as 'At Risk – declining' and Inanga (*Galaxias maculatus*) classified as 'At Risk – declining' are less competent climbers although the inanga can jump and these species would be more likely to occur in the Grey River tributary. The dwarf galaxias (*Galaxias divergens*) classified as 'At Risk – declining' is non-migratory meaning it resides permanently in the stream and is typically less affected by barriers to fish passage. The lamprey (*Geotria australis*) classified as Threatened – Nationally Vulnerable, is an accomplished climber, while the torrent fish (*Cheimarrichthys fosteri*) classified as 'At Risk – declining' is more a fast water specialist with a lesser climbing ability and least likely to reside in Noname Creek.

Bioresarches suggest that a fish barrier may be present where the Noname Creek passes under Taylorville Road but provide no images or assessment. I note that fish passage barrier assessments can be routinely undertaken and uploaded to a national database administrated by NIWA free of charge¹. A further fish passage barrier may occur where the Grey River tributary flows into the Grey River. Passage at this point has not been assessed but is likely to be partial, in that the Grey River tributary may flow within porous river gravels during periods of low flows, but that elevated flows in either the Grey River or Grey River tributary will allow passage.

In summary, the Noname Creek described by Bioresarches has the potential for a diverse macroinvertebrate and fish community with the caveat that there may be barriers to fish migration. Accordingly, I would consider that in the absence of information to the contrary, the Noname Creek would have the potential for at least moderate ecological values.

¹ <https://fishpassage.niwa.co.nz/>. The Fish Passage Assessment Tool has been developed to provide an easy to use, practical tool for recording instream structures and assessing their likely impact on fish movements. Data is collected using the Fish Passage Assessment Survey available in the NIWA Citizen Science app. A User Guide for the Fish Passage Assessment Survey is available [on NIWA's website](#). The information collected using the app is automatically uploaded to a national fish passage database.

2.2 Grey River Tributary

The Grey River tributary appears to arise on the Grey River flats, fed by groundwater at the base of terraces and the Noname Creek. Based on the map provided by EHS Support (Figure 2), it does not have an upstream connection to Noname Creek 2. The area where it arises on the flats, appears to be more wetland/pond habitat than a stream, but the waterway is described as subsequently becoming narrow and swift further downstream. The slow flowing wide reaches of the stream are natural deposition areas for fine sediments and so the predominance of sand and silt is not necessarily an indication of degradation. Sand and silts would derive from upstream erosion or potentially inundation from the Grey River during times of high flow and sediment load.

Bioresearches describe the Grey River tributary as having a variety of flow types (run, riffle, pool) and a diversity of riparian vegetation. As such, the stream provides a spectrum of habitats for both fish and macroinvertebrates. In particular, the slow flowing wetland reaches of the tributary would be favoured habitat for giant kokopu, shortfin eel and inanga. Bioresearches describe fish passage into the stream as being limited by topography but provide no other detail or assessment other than to state that only climbing species could gain access. However, Bioresearches decline to list or consider those climbing species at any point in the assessment. In the absence of a meaningful assessment of fish passage or fish communities it is prudent to assume that a more diverse community may be present than described in the Bioresearches assessment.

Water quality in the Grey River tributary during January 2025 shows potential evidence of some contamination from landfill leachate based on Table 2 of the ENGEO March 2025 memo. Noname Creek 2 (sample SW6), here used as a reference site, had low sulphates² (12.7 mg/l), whereas the Grey River tributary downstream of the roadside drain discharge at TRD5 (SW2) had much higher concentrations (197 mg/l). Further downstream in the Grey River tributary at site SW3 (slightly downstream of the Noname Creek discharge into the Grey River tributary), sulphate levels had begun to decline (58 mg/l). The same pattern was seen for a number of other parameters, potentially indicating that water quality impacts from the landfill can migrate significant distances along the surrounding surface water features. While these January 2025 results suggest a source of contamination from the landfill to the Grey River tributary, the measured concentrations at that time are unlikely to cause significant negative effects on stream ecology. However, based on earlier water quality results in Noname Creek during 2023, it is possible that contaminant concentrations in the Grey River tributary may have been higher at that time.

The Grey River tributary, based on Bioresearches descriptions, is likely to contain a diversity of fish and macroinvertebrates and there is no basis for determining its ecological value to be low. On the basis of the descriptions available, I suspect the headwaters of the Grey River tributary constitute a natural inland wetland as suggested by the ENGEO report, however a formal assessment would be required to confirm that determination.

2.3 Roadside Drain

The Roadside drain is described as artificial with limited flow or habitat heterogeneity. Being artificial does not automatically mean a waterway has low, very low or negligible ecological values but based on the images and description I agree with that conclusion. The fish passage assessment is un-qualified, merely assertion, but there appears to be very limited habitat available on the day images where taken. The key concern about the roadside drain is the collection of potentially contaminated groundwater and conveyance to natural water bodies such as the Grey River Tributary and Noname Creek 2. Water quality

² Sulphates are considered a useful indicator of leachate from the Taylorville Landfill. WCRC Selva Selvarajah pers. Comm. May 15, 2025.

data collected by the applicant confirms that significantly high concentrations of landfill derived contaminants have been measured in the drain.

3.0 Ecological Impact Assessment

It is not clear from the EcIA undertaken by Bioresarches what their understanding of the potential flow paths of leachate and stormwater to the receiving environment might be. The map included in the assessment shows a discharge point at the head of the roadside drain immediately west of the landfill (corresponding with the known spring location), however ENGEO confirm the location of the proposed discharge to land is east of the landfill as shown above in Figure 3 of this memo.

My understanding based on discussion with Carl Steffens is that contaminated water potentially being discharged to land at the location shown in Figure 3 will likely cause significant mounding of groundwater beneath the point of discharge, potentially causing a local increase in groundwater level by up to 8 m (groundwater may reach near the ground surface at the discharge point). This will affect local groundwater flow directions with potential local groundwater flow from the discharge point occurring radially in all directions. This could cause local migration of contaminants in the discharge in a northerly direction toward the upper reaches of Noname Creek, as well as in the natural direction of groundwater flow (generally from west through to south) and also to the south-east. Therefore, PDP's view is that the inferred contaminant flow path indicated by ENGEO and which is shown in Figure 3 may not be accurate.

PDP considers there is the potential for contaminants from the discharge to reach the headwaters of Noname Creek, the existing landfill cells, spring discharges and drains that occur in and along the roadside cutting on the northern side of Taylorville Road (including as far east as TRD4). As a result, contaminants incorporated in the discharge may potentially reach all the drains and streams described in Section 1.3 of this memo. Carl Steffens has indicated that contaminant concentrations will decrease in groundwater with distance from the discharge point, however he considers that information provided to date by ENGEO has not adequately confirmed the scale of reduction in contaminant concentrations prior to discharge of the contaminant plume to surface water. He considers the surface water bodies where contaminant discharges from the proposal may potentially be highest are Noname Creek (near its headwaters) and the roadside drains in the vicinity of TRD2 and TRD4. At TRD2 and TRD4, any discharges of contaminants from the proposal only need to travel a short distance through the Taylorville Road culverts into Noname Creek 2. The Grey River tributary is expected to be less likely to be impacted from contaminant discharges from the proposed discharge due to dilution, however as discussed previously water quality sampling indicates a potential impact in the Grey River tributary from the general landfill activities. While far field and large water bodies such as the Grey River are unlikely to show any demonstrable effect of relatively small volumes of contaminants, the uncertainty around the pathways and eventual fate and concentrations of contaminants at this site is of general concern.

ENGEO provide indicative estimated concentrations derived from the mixing of the discharge with flow in the Noname Creek. Their estimates are presented in Table 3 of the ENGEO Preliminary Hydrogeological Assessment and we have used those estimates to consider potential outcomes for near field receiving environments at the site. However, I note that ENGEO has carried out more recent RISC5 contaminant modelling to estimate contaminant concentrations at the nearest offsite receptor (310 m away as reported by ENGEO). Based on discussions with Carl Steffens, he has informed me he considers the modelling cannot be relied upon based on the level of detail provided by ENGEO to date. In addition, taking into account the mounding discussed previously, he also considers the nearest receptors may be closer (potentially more akin to 100 m distant from the discharge point). Further to the above, I note that the ENGEO text describes the generation of Table 3 using maximum values of stormwater and leachate (Table 3: column 2), but those values don't match the maximum values in the raw data presented in ENGEO Table A6.1. For example, table 3 uses a maximum value of 16.1 mg/l (total ammoniacal nitrogen) collected in

cell B/C, while the actual maximum value from cell B/C is 98 mg/l. The value used by ENGEO for their later RISC5 modelling was 98 mg/l. Therefore, we also have concerns that their estimates presented in Table 3 of their report may not be conservative enough.

Despite the issues described above and significant uncertainty with regard to potential contaminant discharge concentrations, we have used the information provided in Table 3 from ENGEO to make an initial assessment of potential effects. ENGEO Table 3 contaminant concentrations are presented alongside measured maximum values from the downstream site in Noname Creek (Table 1). As discussed above, there is the suggestion that maximum values measured in the Noname Creek may be representative of an inappropriate deposition of organic matter into the landfill and not representative of long-term natural stream conditions. The effects of that organic matter are considered potentially to have passed (although it can't be ruled out that future interactions between rainfall, groundwater and the fill within the landfill cells may result in further periods of significant stream impacts) and therefore water quality results at the downstream Noname Creek site collected in January 2025 have been used in Table 1 to indicate long-term water quality.

It is important to note that none of these concentrations are adjusted for bioavailability (e.g. adjustments based on pH, dissolved organic carbon or hardness) and that the concentrations shown are total metals not the dissolved fraction conventionally used for toxicity assessments. The latter option was chosen because Table 3 ENGEO only provided total concentrations and lacked estimates of any of the toxicant modifying factors.

Measured maximum values in Noname Creek were typically less than the thresholds for protection of 95 % of species and would be considered suitable for a moderately impacted stream. Exceptions were the maximum concentrations of Nitrate-N, total aluminium, total chromium and total zinc. However, it should be noted that the maximum aluminium and zinc concentrations were an order of magnitude greater than the three other samples collected, while the Nitrate -N concentration was not out of character for the creek during 2023-24.

Concentrations recorded in January 2025 were in some cases lower than the maximums by an order of magnitude. A high level of species protection was provided by concentrations of all contaminants other than aluminium, arsenic (V) and total chromium (VI) that provided 95 % protection. These latest results suggest that water quality in the Noname Creek is generally good and would support diverse ecological communities.

Assessment undertaken by ENGEO indicated that many contaminant concentrations could increase relative to the maximum values, including Ammoniacal-N, total manganese, total arsenic, total chromium, total copper, total lead, total nickel and total zinc. In some cases, the extent of the potential increase is that ANZG (2018) would indicate the site is a highly disturbed system and measurably degraded with chronic toxicity effects on more than 80 % of species. The level of impact is greater when the January 2025 Noname Creek water quality data is used which we consider is likely more representative.

Table 1. Comparison of modelled and measured contaminant concentrations in Noname Creek

	Noname Creek maximum (SPFDS)		Noname Creek Jan 2025 (SPFDS)		Table 3	
Contaminant	Max Concentration (mg/l)	ANZG Protection level (and threshold, mg/l)	Concentration (mg/l)	ANZG Protection level (and threshold, mg/l)	Concentration (mg/l)	ANZG Protection level (and upper threshold, mg/l)
Total Ammoniacal N	0.014	>99 %	<0.010	>99 %	5.07	<80 % (2.3)
Nitrate N	1.33	>90 % (1)	0.25	>99 %	0.11	>99 % ³
Total ⁴ aluminium	0.26	<80 % (0.15)	0.044	>95 % (0.027)	0.066	>90 % (0.055)
Total manganese	0.039	>99 %	0.0127	>99 %	0.82	>99 %
Total arsenic (AsV)	0.0012	>95 % (0.0008)	<0.0011	>95 % (0.0088)	0.02	>90 % (0.013)
Total chromium (Cr VI)	0.00102	>90 % (0.001)	0.0006	>95 % (0.00001)	0.0052	>90 % (0.001)
Total copper	0.00123	>95 % (0.001)	0.00088	>99 %	0.056	<80 % (0.0025)
Total lead	0.00074	>99 %	0.00012	>99 %	0.0012	>95 % (0.001)
Total nickel	<0.00053	>99 %	<0.00053	>99 %	0.018	<80 % (0.017)
Total zinc	0.0103	>90% (0.008)	0.0024	99 %	0.066	<80 % (0.031)

³ Note the ANZECC 2000 default guideline for Nitrate was erroneous and ANZG (2018) refer to the grading guidelines in (Hickey, 2013).

⁴ Metal toxicity is typically compared against concentrations of the dissolved fraction rather than total. As such this comparison is inherently conservative.

4.0 Summary

There is considerable uncertainty around the ecological values of receiving environments potentially affected by the proposed TRPL discharge. I do not agree that the ecological values of the receiving environments that have been assessed can be determined as low. In contrast, based on the information available I would consider values to be at least moderate. Values in Noname Creek 2 have not been assessed.

There is further uncertainty around the flow paths of contaminant laden groundwater to receiving environments, in particular in the event that groundwater mounding results in leachate and stormwater being discharged to Noname Creek and the Taylorville Road drain in relatively close proximity to the discharge location.

Finally, the concentrations of contaminants that reach receiving environments are uncertain although previous estimates by ENGEO suggest potential effects would result in a degradation of ecological values. We note ENGEO has carried out more recent modelling of contaminant discharges at surface water receptors using RISC5, however we don't consider the outputs from that modelling can be relied upon.

In light of the available information, it seems possible that there is the potential for demonstrable and long lasting effects on the ecological values of Noname Creek (and possibly Noname Creek 2, and potentially the Grey River tributary). Due to the large dilution available, there are highly unlikely to be effects observed in the Grey River. While it is realistic that contaminant concentrations will reduce with distance from the discharge point, we don't consider that ENGEO has reliably demonstrated to date that concentrations will reduce to a level whereby there are no potential effects in nearby surface water bodies. In addition, based on the ENGEO Preliminary Hydrological Assessment, the potential effects in Noname Creek would be significant.

5.0 References

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6.0 Limitations

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